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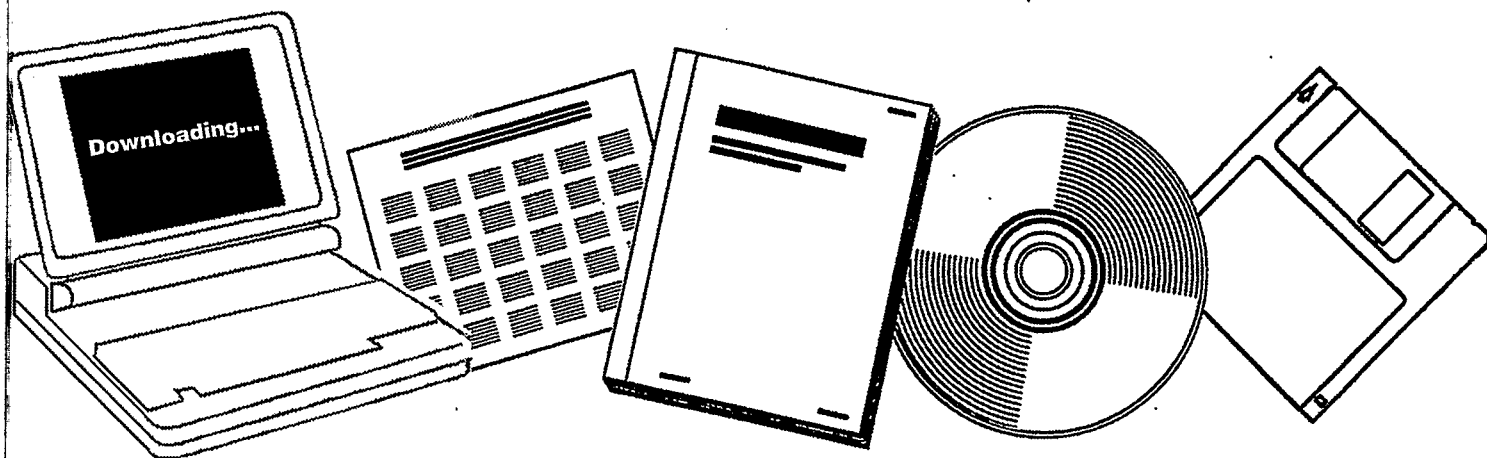
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**ASSESSMENT OF TRACE CONTAMINANTS FROM A
MODEL INDIRECT LIQUEFACTION FACILITY.
VOLUME II. STREAM CHARACTERIZATION OF
LURGI/FISCHER-TROPSCH COAL LIQUEFACTION**

**GENERAL RESEARCH CORP.
MCLEAN, VA**

JAN 1982



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Assessment of Trace Contaminants From a Model Indirect Liquefaction Facility

VOLUME II -- STREAM CHARACTERIZATION OF LURGI/FISCHER-TROPSCH
COAL LIQUEFACTION

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FOREWORD

Development and deployment of a commercial indirect liquefaction industry has been proposed as a means of reducing United States dependence on foreign sources of energy.

Deployment of a commercial industry on an environmentally acceptable basis requires identification and evaluation of potential environmental problems. This assessment is an attempt to anticipate potential environmental hazards that may be posed by commercial-scale facilities to provide an improved basis for planning and implementing environmental research.

The study comprises four major tasks: characterization of hazardous materials released from an indirect liquefaction facility; assessment of ecological hazards; assessment of public health hazards; and assessment of occupational health hazards. The report is organized in the same manner. Volume I is an overview and summary of the results; volume II presents stream characterization data; and volumes III, IV and V present assessments of ecological, public health and occupational health hazards, respectively.

This study was sponsored by the Technology Assessment Division of the Department of Energy. Organizations participating in the assessment were General Research Corporation, Oak Ridge National Laboratory, and Argonne National Laboratory.

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1 INTRODUCTION .

Indirect liquefaction is a promising technology for producing synfuels. Recent studies indicate that if any portion of the national synfuels goals for 1987 and 1992 is to be met with coal liquefaction, the bulk of the production is likely to come from indirect processes. An indirect coal liquefaction industry will benefit the nation by providing a critically needed supplement to our dwindling oil and gas reserves. However, development and deployment of a commercial coal indirect liquefaction industry is not without risks. To ensure development and deployment of indirect processes in an environmentally acceptable manner, potential hazards to the general public, occupational personnel and ecosystems must be assessed and factored into the design, siting and operation of commercial facilities. At present, the potential adverse environmental impacts of indirect liquefaction facilities are not well understood.

Especially lacking is information on the identity and quantity of trace contaminants that may be released from process and waste streams. At the present time chemical characterization data for streams from indirect liquefaction facilities are very limited. In response to these data gaps the Department of Energy (DOE) is attempting to develop preliminary estimates of the types and quantities of potential pollutants that may be released into the environment by commercial scale indirect liquefaction facilities. The purpose of these estimates is to provide a basis for estimating the types and magnitudes of potential environmental impacts of a commercial facility and a commercial industry.

The purpose of this draft report is to document the methodology, assumptions, caveats, results, and references used in developing these preliminary estimates. It is hoped that such documentation will enhance understanding of both the value and limitations of the estimates, and facilitate modification of the estimates as data become available from ongoing research.

2 METHODOLOGY

2.1 Overview

Chemical characterization data were very limited for indirect coal liquefaction; no commercial-scale facilities existed in the U.S., and data from other nations (e.g., South Africa) were generally limited and often not representative of American conditions. Because data were so limited, it was recognized that development of an information base useful for assessing environmental implications would require considerable manipulation of the data which were available. In order to minimize confusion and misinterpretation, data sources, calculations, assumptions and results were made explicit wherever possible.

Development of the data base was accomplished in several steps:

1) Choice of Indirect Liquefaction Process-

A single process was chosen to serve as the basis of the analysis to avoid the confusion which would arise from attempting to deal with several processes or a "generic" process. The Lurgi/Fischer-Tropsch process was chosen because it represented a proven, commercial technology, and more data were available for it than most other processes.

2) Development of a Conceptual Plant Configuration-

A conceptual Lurgi/Fischer-Tropsch plant, processing approximately 28,000 tons per stream day (TPSD) of low sulfur Wyoming subbituminous coal, was developed to provide the basis for estimating the types and quantities of pollutants which might be released from a commercial facility. Process operations and streams were based largely, but not exclusively, upon information in the report titled Research Guidance Studies to Assess Gasoline from Coal by Methanol-to-Gasoline and SASOL-type Fischer-Tropsch Technologies⁸⁷. The process description and basic premises used in conceptualizing the plant are presented in Appendix A. A simplified block flow diagram, developed from the report,

is presented in Figure 1. As indicated in the diagram, the conceptual plant included environmental controls as well as process and auxiliary operations. The major process streams and environmental streams of interest in the analysis are listed in Table 1.

3) Identification of Stream Components of Interest-

Review of the literature indicated that a variety of stream components associated with coal liquefaction would be of environmental concern. Potential stream components included in the analysis were selected by representatives of participating groups from Oak Ridge National Laboratory, Argonne National Laboratory and Battelle Pacific National Laboratory at a workshop, and are listed in Table 2. In selecting components of interest it was assumed that the plant complex would be designed to meet all existing local and federal environmental regulations, as of July 1977, for liquid and gaseous effluents. The analysis was focused largely on trace organics and trace elements.

4) Estimation of Flow Rates of Stream Components by Stream-

Flow rates of each stream component were estimated for each stream to provide initial estimates of types and quantities of compounds to be expected in each stream. Estimates of the flow rates of the major stream components were derived largely from Base Case II of the Mobil report.⁸⁷ Flow rates of the minor components (i.e., those constituting less than one percent of the stream flow rate) were estimated based on sparse information. The results are presented in Appendix B.

5) Estimation of Concentrations of Individual Trace Elements and Organic Compounds in Each Stream-

No data could be located in the literature regarding concentrations of individual trace elements and organic compounds in streams from commercial-scale Lurgi/Fischer-Tropsch plants utilizing Wyoming subbituminous coal. The types and concentrations of individual trace elements and organic compounds were estimated using data from a

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TABLE 1

PROCESS AND ENVIRONMENTAL STREAMS OF INTEREST*

| STREAM NUMBER | STREAM NAME | SOURCE | STREAM NUMBER | STREAM NAME | SOURCE | STREAM NUMBER | STREAM NAME | SOURCE |
|---------------|---------------------------------------|------------------------------|---------------|-----------------------|------------------------|---------------|-------------------------------|-----------------------|
| 1 | Feed Coal | | 35 | Vent Gas | Ash Handling | 64 | OFF Gas | F-T Product Upgrading |
| 2 | Sized Coal | Coal Preparation | 36 | Ash/Scrubber Sludge | Ash Handling | 65 | Waste Stream | F-T Product Upgrading |
| 3 | Undersized Coal | Coal Preparation | 37 | Sluice Water | Ash Handling | 66 | Heater Stack Gas | F-T Product Upgrading |
| 4 | Oxygen | Air Separation | 38 | Waste Water | Ash Handling | 67 | CO ₂ Purge Gas | F-T Product Upgrading |
| 5 | Nitrogen | Air Separation | 39 | Waste Liquor | Waste Steam Generator | 68 | CO ₂ Off Gas | F-T Product Upgrading |
| 6 | Steam | Utilities Generation | 40 | Waste Liquor | Waste Steam Generator | 69 | Leachate from Ash | Stream 36 |
| 7 | Raw Gas | Gasifier | 41 | Waste Liquor | Waste Steam Generator | 70 | Biological Sludge | Biological Treatment |
| 8 | Washed Gas | Wash Cooler | 42 | Waste Liquor | Gas Cooling | 71 | Leachate from Bio/sludge | Stream 70 |
| 9 | Cooled Gas | Waste Steam Generator | 43 | Waste Liquor | Gas Cooling | 72 | Lockhopper vent Gas Emissions | Coal Lockhopper |
| 10 | Gas | Gas/Liquor Separator | 44 | Tars | Streams 40,41,42,57,58 | 73 | Evaporative Losses | Stream 18-23 |
| 11 | Gas | Gas Re-compression/Separator | 45 | Oils | Gas/Liquor Separation | 74 | Evaporative Losses | Stream 60 |
| 12 | Gas | Gas/Liquor Separator | 46 | Oil/Tar Free Liquor | Gas/Liquor Separation | 75 | Evaporative Losses | Stream 45 |
| 13 | Cooled Shifted Gas | Gas Cooling | 47 | Phenols | Phenol Recovery | 76 | Evaporative Losses | Stream 44 |
| 14 | Cooled Gas | Gas Cooling | 48 | Phenol-Free Liquor | Phenol Recovery | 77 | Evaporative Losses | Stream 47 |
| 15 | Combined Gas | Streams 13 & 14 | 49 | Ammonia | Ammonia Recovery | 78 | Evaporative Losses | Stream 49 |
| 16 | Purified Gas | Rectisol Unit | 50 | Ammonia-Free Liquor | Ammonia Recovery | 79 | Spent Shift Catalyst | Raw Gas Shift |
| 17 | Fischer-Tropsch Products | Fischer-Tropsch Synthesis | 51 | Waste Air | Biological Treatment | | | |
| 18 | SNG | F-T Product Upgrading | 52 | Biological Effluent | Biological Treatment | | | |
| 19 | C ₂ HFC | F-T Product Upgrading | 53 | Concentrated Waste | Reverse Osmosis | | | |
| 20 | C ₃ HFC | F-T Product Upgrading | 54 | Treated Effluent | Reverse Osmosis | | | |
| 21 | Gasoline | F-T Product Upgrading | 55 | Sour Gas | Rectisol Unit | | | |
| 22 | Distillate Fuel Oil | F-T Product Upgrading | 56 | Incineration Gas | Stretford Unit | | | |
| 23 | Heavy Fuel Oil | F-T Product Upgrading | 57 | Waste Liquor | Rectisol Unit | | | |
| 24 | Fugitive Emissions | Coal Preparation | 58 | Waste Liquor | Alcohol Recovery | | | |
| 25 | Coal Dust | Bag House | 59 | Aqueous Liquor | F-T Synthesis | | | |
| 26 | Vent Gas | Bag House | 60 | Heavy Alcohol Product | Alcohol Recovery | | | |
| 27 | Coal Pile Leachate | Bag House | 61 | Catalyst Raw Material | | | | |
| 28 | Stack Gas | Coal Preparation | 62 | F-T Catalyst | Catalyst Preparation | | | |
| 29 | Evaporative Losses | Utilities Generation | 63 | Spent F-T Catalyst | F-T Synthesis | | | |
| 30 | Deaerator Losses | Utilities Generation | | | | | | |
| 31 | Make-up Water | Utilities Generation | | | | | | |
| 32 | Blowdown | Utilities Generation | | | | | | |
| 33 | Ash & SO ₂ Scrubber Sludge | Utilities Generation | | | | | | |
| 34 | Ash Slurry | Ash Lockhopper | | | | | | |

*Stream numbers, names and sources are keyed to the flow diagram in Figure 1.

TABLE 2

STREAM COMPONENTS SELECTED FOR INCLUSION
IN THE ANALYSIS

| <u>Gaseous Components</u> | | |
|------------------------------------|-------------------------------|---------------------|
| CO | C ₂ H ₄ | H ₂ S |
| CO ₂ | C ₂ H ₆ | COS |
| H ₂ | N ₂ + Inerts | H ₂ O |
| CH ₄ | O ₂ | Others |
| <u>Liquid and Solid Components</u> | | |
| H ₂ O | Thiophenes | Ni(CO) ₄ |
| Methanol | Ammonia | Minerals |
| Tar | HCN | Coal |
| Oil | Aromatic amines | Sulfur |
| Naphtha | Nitrosamines | Particulates |
| Crude phenols | Polynuclear aromatics (PNA) | Trace elements |
| Mercaptans | Fatty Acids | Others |

variety of sources.* Detailed descriptions of data sources and values, assumptions, and calculations are presented in Sections 2.2, 2.3, and 2.4.

6) Evaluation of Results-

Results of the effort were evaluated and compared with the results and projections from other studies to provide some insight into the reasonableness of the values.

2.2 Characterization of Liquid Streams

Review of the literature indicated that characterization data for liquid effluents for commercial-scale Lurgi/Fischer-Tropsch plants were very limited. No experimental data could be found regarding specific constituents in treated effluents from commercial-scale Lurgi/Fischer-Tropsch processing of Wyoming subbituminous coal.

Liquid streams selected for the analysis are listed in Table 3; their inter-relationships are illustrated in Figure 2.

Wash process liquor (Stream 43) was chosen as the key liquid stream of interest because it represents the largest and most highly contaminated wastewater stream in a Lurgi/Fischer-Tropsch plant.³⁷ It was also chosen for analysis because it is the major influent to the wastewater treatment facility in the plant; the types and quantities of constituents present in Stream 43 largely determine the types and quantities of pollutants present in the liquid and solid effluents from the wastewater treatment facility (i.e., streams 53, 54, 70 and 71).

The analysis focused on two general classes of pollutants which have been identified as primary causes of concern in liquid streams from coal conversion facilities: trace elements and organic compounds. The

*Data from both theoretical and experimental studies were used when they were deemed appropriate and useful.

TABLE 3
LIQUID STREAMS CHARACTERIZED IN THE ANALYSIS

| STREAM NUMBER | STREAM NAME |
|---------------|---|
| 27 | Coal pile leachate |
| 31 | Make-up water |
| 38 | Ash sluice water |
| 43 | Raw process water |
| 46 | Gas liquor separator effluent |
| 48 | Phenosolvan effluent |
| 50 | Ammonia stripper effluent |
| 52 | Biological treatment effluent |
| 53 | Reverse Osmosis concentrated waste solution |
| 54 | Reverse Osmosis permeate |
| 54 & 31 | Feed to cooling towers |

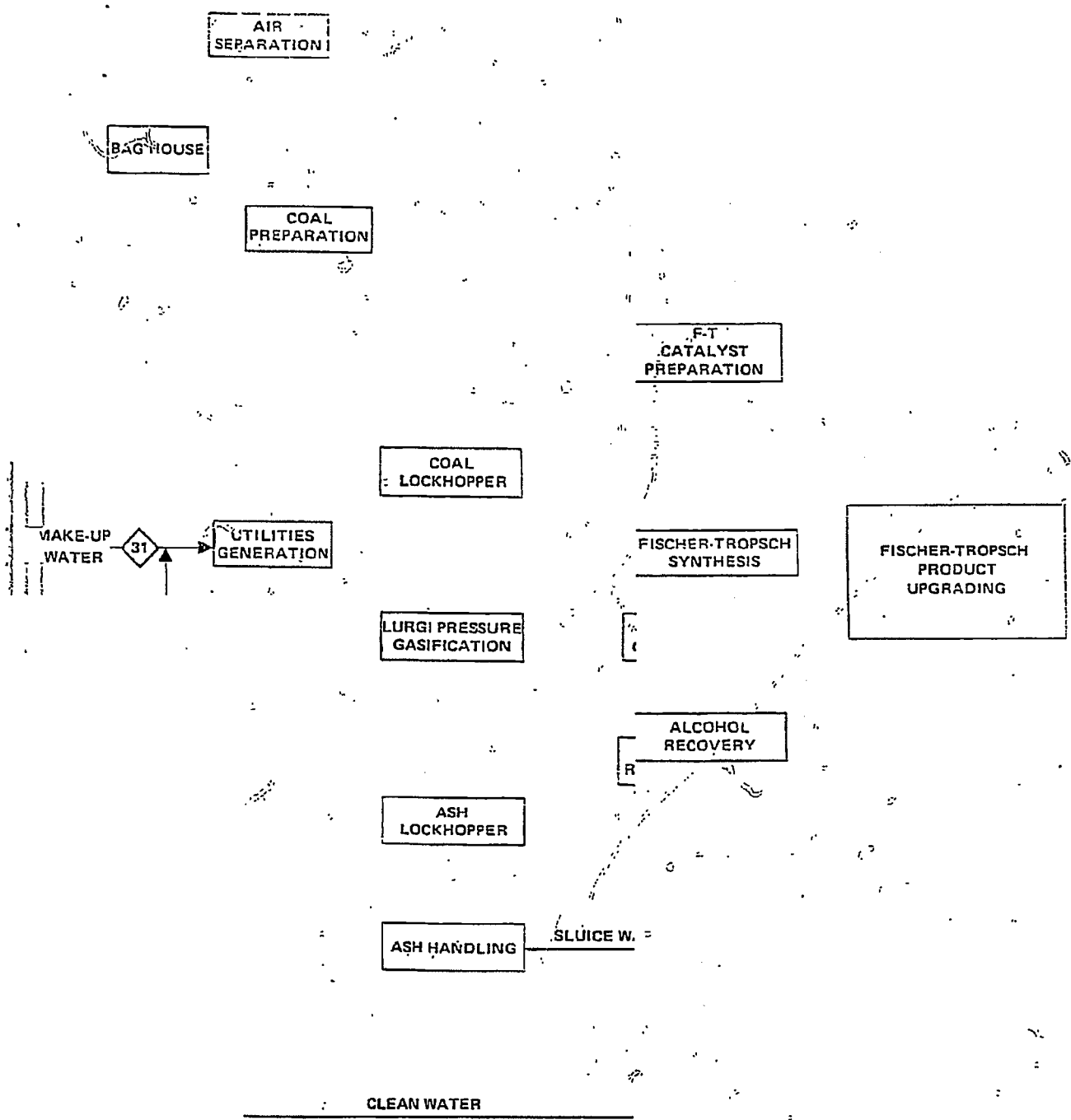


FIGURE 2

**LIQUID STREAMS
FOR
JAL LIQUEFACTION PROCESS
(31-FISCHER-TROPSCH)**

types and quantities of pollutants were estimated based on three sets of data. Estimates of trace elements were based largely on experimental trace element distribution coefficients developed for Lurgi at SASOL.³⁷ Estimates of organic compounds were based on experimental effluent data from SASOL⁸ and on characterization data for coal tar produced by a bench-scale gasifier using Wyoming subbituminous coal.¹⁹

The types and quantities of constituents which may be present in liquid streams were estimated in 3 basic steps:

- 1) Identification of the types and quantities of constituents potentially present in the major raw process waste stream (Stream 43).
- 2) Identification or estimation of efficiency of removal of each constituent (identified in Step 1) by each wastewater treatment process specified in the flow diagram (i.e., gas/liquor separation, phenol recovery, ammonia recovery, biological treatment, and reverse osmosis).
- 3) Calculation of constituent concentrations in liquid effluents from the water treatment facility by sequentially applying the removal efficiencies for each constituent as the stream passed through each water treatment process. Additional pollutant loadings added to the liquid stream (via Streams 27 and 28) were estimated from literature data and incorporated into the calculations at the appropriate point.

The specific calculations, data and assumptions used in estimating the types and concentrations of trace elements and organic compounds which may be present in the liquid streams are discussed in Sections 2.2.1, 2.2.2 and 2.2.3.

2.2.1. Estimation of Trace Elements in Liquid Streams

Estimation of Trace Element Concentrations in Stream 43-

The concentrations of trace elements in Wyoming subbituminous coal on a moisture free, whole coal basis were identified in the literature. (See Table 4).

The flow rate of each trace element into the gasifier was estimated:

$$F_{TE-G} = (C_{TE})(F_{DC}) / 1,000,000$$

TABLE 4

TRACE ELEMENT CONCENTRATIONS FOR WYOMING AND ROSEBUD MONTANA
SUBBITUMINOUS COALS (ppm)

| Trace Elements | Wyoming Subbituminous ⁴⁰ | Montana Rosebud Subbituminous ⁴⁰ | Ratio of Concentrations of Trace Elements - Montana to Wyoming |
|----------------|-------------------------------------|---|--|
| Ag | .06-.43 | .06 | 1-.14 |
| As | .57-1.2 | .08-1.2 | .14-1 |
| B | 32 | 32 | 1 |
| Ba | 87 | 87 | 1 |
| Be | .71-.8 | .7-.8 | .99-1 |
| Br | - | - | - |
| Cd | .31-.8 | .31-.8 | 1-1 |
| Ce | - | - | - |
| Co | .55 | .6-4 | 1.09-7.27 |
| Cr | 4.2-16 | 4-16 | .95-1 |
| Cs | - | - | - |
| Cu | 8.9-10 | 9-10 | 1.01-1 |
| F | 65-67 | 66 | .98-1.01 |
| Ga | - | - | - |
| Ge | - | - | - |
| Hg | .11-.17 | .11-.17 | 1-1 |
| I | - | - | - |
| In | - | - | - |
| La | - | - | - |
| Li | 3.6-15.0 | - | - |
| Mo | 2.2 | 2.2 | 1 |
| Mn | 2.8-3.4 | 2.8-3.4 | 1-1 |
| Ni | 1.7-14 | 2-14 | 1.18-1 |
| P | - | - | - |
| Pb | .51-12 | .51-12 | 1-1 |
| Rb | - | - | - |
| Ru | - | - | - |
| Sb | .08-1.5 | - | - |
| Sc | - | - | - |
| Se | .33 | .33 | 1 |
| Sn | .14 | .14 | 1 |
| Sr | - | - | - |
| Ta | - | - | - |
| Te | - | - | - |
| U | .88 | .88 | 1 |
| V | 10-14 | 10-14 | 1 |
| W | - | - | - |
| Y | - | - | - |
| Zn | .23-8 | 2-8 | 8.70-1 |
| Zr | - | 170 | - |

where-

F_{TE-G} is the flow rate of each trace element into the gasifier (in lb/hr).

C_{TE} is the concentration of each trace element in dry, whole coal (in ppm).

F_{DC} is the flow rate of dry, whole coal to the gasifier (Water-free components of Stream 2 in the flow diagram).*

The distribution of each trace element from Lurgi gasifier into ash, liquor, tar and oil streams at SASOL were identified in the literature. (See Table 5).

The distribution of trace elements in the conceptual plant was assumed to be the same as the distribution at SASOL.

The flow rate of each trace element into the gas liquor was estimated:

$$F_{TE-L} = \frac{(F_{TE-G})(D_{TE-L})}{100\%}$$

where-

F_{TE-L} is the flow rate of each trace element into the gas liquor (in lb/hr).

F_{TE-G} is the flow rate of each trace element into the gasifier (in lb/hr).

D_{TE-L} is the distribution factor of each trace element in the liquor based on SASOL data (in %) (See Table 5).

It was assumed that the total flow of each trace element into the gas liquor dissolved in the water component of the gas liquor stream, then

$$C_{TE-43} = \frac{F_{TE-L} \times 1,000,000}{F_{W-43}}$$

*Flow rates of each stream are presented in Appendix B.

TABLE 5

TRACE ELEMENT DISTRIBUTION FOR LURGI AT SASOL*,³⁷
 (Percent of Element in Coal)

| Element | Ash | Liquor | Tar | Oil |
|---------|--------|--------|------|-----|
| Be | 33.3 | 53.3 | 17.0 | 0.3 |
| B | 90 | 8.8 | 2.0 | 0.0 |
| V | 99.9 | 0.1 | 0.0 | 0.0 |
| Mn | 99.9 | 0.2 | 0.0 | 0.0 |
| Ni | 99.4 | 0.4 | 0.0 | 0.0 |
| As | 26.9 | 67.2 | 1.9 | 3.9 |
| Cd | 51.9 | 45.5 | 0.6 | 1.4 |
| Sb | 50.0 | 45.0 | 3.8 | 0.6 |
| Ce | 99.9 | 0.1 | 0.0 | 0.0 |
| Hg | 51.9 | 41.6 | 6.4 | 0.6 |
| Pb | 94.2 | 1.7 | 4.3 | 0.0 |
| Br | 10.0 | 88.9 | 0.1 | 0.0 |
| F | 56.3** | 43.8** | 0.0 | 0.0 |
| Cl | 52.6** | 47.4** | 0.3 | 0.0 |

* Analysis by spark source mass spectrometer (which can give a semi-quantitative analysis) for El Paso by SASOL.

** % distribution calculated on analyses as done by SASOL previously.

where-

C_{TE-43} is the concentration of each trace element in the water component of Stream 43 (in ppm).

F_{TE-L} is the flow rate of each trace element into the gas liquor (in lb/hr).

F_{W-43} is the flow rate of the water component of Stream 43, i.e., 1,993,000 lb/hr.

Estimation of Trace Element Concentrations in Stream 27-

The concentration of each trace element in coal pile leachate (Stream 27) was identified in the literature (Table 6). The literature data were assumed to be representative of Wyoming subbituminous coal.

Identification/Estimation of Efficiencies of Removal of for Each Constituent by Each Wastewater Treatment Process-

Efficiencies of removal of each constituent in the liquid streams (43 and 27) were identified or estimated from the literature for each wastewater control technology process identified in the flow diagram. Removal efficiencies assumed in the assessment are presented in Table 7.

Estimation of Trace Element Concentrations in Stream 46-

The concentration of each trace element in Stream 46 was estimated by weighting the concentrations in streams 43 and 27 by their respective flow rates, applying the efficiency of removal of each constituent by the gas liquor separator, and dividing by the flow rate of Stream 46:

$$C_{TE-46} = \frac{(C_{TE-43})(F_{W-43}) + (C_{TE-27})(F_{W-27}) \left[\frac{100 - RE_{TE-ES}}{100} \right]}{F_{W-46}}$$

where-

C_{TE-46} is the concentration of each trace element in the water component of Stream 46 (in ppm).

C_{TE-43} is the concentration of each trace element in the water component of Stream 43 (in ppm).

TABLE 6
LITERATURE VALUES FOR STREAM CONSTITUENTS

| CONCENTRATIONS IN STREAMS CONSTITUENTS | STREAM CONCENTRATIONS (ppm) | | |
|---|-------------------------------|-----------------------------------|---|
| | RAW GAS LIQUOR (STREAM 43) | COAL PILE LEACHATE (STREAM 27) | ASH SLUICE WATER BLOWDOWN (STREAM 38) |
| <u>TRACE ELEMENTS</u> | | | |
| Arsenic | 0.55 (1) | 0.01 (2) | 0.02 (3) |
| Boron | 1.9 (1) | | |
| Beryllium | 0.29 (1) | 0.01 (2) | |
| Cadmium | 0.25 (1) | 0.006 (2) | 0.00064 (3) |
| Fluorine | 2.0 (1) | | |
| Lead | 0.14 (1) | 0.023 (2) | 0.09 (3) |
| Mercury | 0.049 (1) | 0.027 (2) | 0.0003 (3) |
| Manganese | 0.0047 (1) | 110.0 (2) | 0.5 (3) |
| Nickel | 0.0036 (1) | 0.32 (2) | 9.35 (3) |
| Vanadium | 0.0096 (1) | | |
| <u>CARBOXYLIC ACIDS</u> | | | |
| Acetic Acid | 171. (4) | | |
| Propanoic Acid | 26. (4) | | |
| Butanoic Acid | 13. (4) | | |
| 2-Methylpropionic Acid | 2. (4) | | |
| Pentanoic Acid | 12. (4) | | |
| 3-Methylbutanoic Acid | 1. (4) | | |
| Hexanoic Acid | 1. (4) | | |
| <u>BENZENE AND SUBSTITUTED BENZENES</u> | | | |
| Biphenyl | 0.7 (5) | | |
| Ethylbenzene | 15. (5) | | |
| Indan | 8.9 (5) | | |
| Toluene | 50. (5) | | |
| 1,2,4 Trimethylbenzene | 6. (5) | | |
| O-Xylene | 18. (5) | | |
| <u>MONOHYDRIC PHENOLS</u> | | | |
| Phenol | 3100. (6) | | |
| 2-Methylphenol | 340. (6) | | |
| 3-Methylphenol | 420. (6) | | |
| 4-Methylphenol | 300. (6) | | |
| 2,4-Xylenol | 120. (6) | | |
| 3,5-Xylenol | 50. (6) | | |
| <u>DIHYDRIC PHENOLS</u> | | | |
| Catechol | 550. (6) | | |
| 3-Methylcatechol | 400. (6) | | |
| 4-Methylcatechol | 385. (6) | | |
| 3,6-Dimethylcatechol | 45. (6) | | |
| Resorcinol | 275. (6) | | |
| 5-Methylresorcinol | 65. (6) | | |
| 4-Methylresorcinol | 36. (6) | | |

TABLE 6 (CON'T)

| CONCENTRATIONS IN STREAMS CONSTITUENTS | STREAM CONCENTRATIONS (ppm) | | |
|---|-------------------------------|-----------------------------------|---|
| | RAW GAS LIQUOR (STREAM 43) | COAL PILE LEACHATE (STREAM 27) | ASH SLUICE WATER BLOWDOWN (STREAM 3B) |
| <u>POLYNUCLEAR AROMATICS</u> | | | |
| Acenaphthylene. | 0.4 (5) | | |
| Anthracene | 0.1 (5) | | |
| Benz(a) Anthracene | 0.01 (5) | | |
| Benzo(g,h,i)perylene | 0.0003 (5) | | |
| Benzo(a)pyrene | 0.004 (5) | | |
| Benzo(e)pyrene | 0.004 (5) | | |
| Chrysene | 0.002 (5) | | |
| Fluoranthene | 0.2 (5) | | |
| Fluorene | 0.2 (5) | | |
| Naphthalene | 3.2 (5) | | |
| Perylene | 0.0004 (5) | | |
| Phenanthrene | 0.1 (5) | | |
| Pyrene | 0.2 (5) | | |
| <u>SULFUR HETEROCYCLICS</u> | | | |
| Methylthiophene | | | |
| Thiophene | | | |
| <u>NITROGEN HETEROCYCLICS</u> | | | |
| Acridine | 2.2 (7) | | |
| 2,4-Dimethylpyridine | 1. (4) | | |
| 2,5-Dimethylpyridine | 1. (4) | | |
| 2-Methylpyridine | 70. (4) | | |
| 3-Methylpyridine | 26. (4) | | |
| 4-Methylpyridine | 6. (4) | | |
| Pyridine | 117. (4) | | |
| Quinoline | 45. (4) | | |
| <u>OXYGEN HETEROCYCLICS</u> | | | |
| Benzofuran | | | |
| Dibenzofuran | | | |
| <u>MERCAPTANS</u> | | | |
| Methanethiol | 20. (7) | | |
| <u>AROMATIC AMINES</u> | | | |
| Aniline | 12. (4) | | |

- (1) Estimate based on SASOL distribution coefficients (37).
(2) Reference 18.
(3) Reference 91.
(4) Reference 8.
(5) Estimate based on RTI data (19). Limited by compound solubility.
(6) Reference 92.
(7) Estimate based on RTI data (19). Limited by compound production rate.

TABLE 7

REMOVAL EFFICIENCIES FOR LIQUID STREAM
CONSTITUENTS BY CONTROL PROCESS
(Percent Removed)

| COMPOUND | GAS LIQUOR SEPARATOR (RE _{X-LS}) | PHENO- SOLVAN- (RE _{X-PH}) | AMMONIA RECOVERY (RE _{X-AR}) | BIOLOGICAL TREATMENT (RE _{X-BT}) | REVERSE OSMOSIS (RE _{X-RO}) |
|---|--|--|--|--|---|
| ALIPHATICS, ALICYCLICS AND FATTY ACIDS | 0 | 15 (1)* | 0 | 95 (2,3) | |
| Acetic Acid | | | | | 51 (9) |
| Propanoic Acid | | | | | 65 (9) |
| Butanoic Acid | | | | | 63 (10) |
| 2-Methylpropionic Acid | | | | | 63 (10) |
| Pentanoic Acid | | | | | 63 (10) |
| 3-Methylbutanoic Acid | | | | | 63 (10) |
| Hexanoic Acid | | | | | 63 (10) |
| BENZENE AND SUBSTITUTED BENZENES | | 15 (1) | | 40-90 ⁺ (3,4,5) | 63 (10) |
| Ethylbenzene | | | | 90 (4) | |
| Toluene | | | | 90 (4) | |
| 1,2,4-Trimethylbenzene | | | | 90 (4) | |
| O-Xylene | | | | 90 (5) | |
| MONOHYDRIC PHENOLS | | 99.5 (1) | | 92-94 (6) | |
| Phenol | | | | 99 (7) | 69 (9) |
| 2-Methylphenol | | | | 97 (7) | 74 (9) |
| 3-Methylphenol | | | | 97 (7) | 87 (9) |
| 4-Methylphenol | | | | 97 (7) | 75 (9) |
| 2,4-Xylenol | | | | 81 (7) | 90 (9) |
| 3,5-Xylenol | | | | 37 (7) | 90 (9) |
| DIHYDRIC PHENOLS | | 60 (1) | | HIGH (3,7) | |
| Catechol | | | | 97 (7) | 75 (9) |
| 3-Methylcatechol | | | | 100 (5) | (11) |
| 4-Methylcatechol | | | | 97 (8) | (11) |
| 3,6-Dimethylcatechol | | | | 97 (8) | (11) |
| Resorcinol | | | | 98 (7) | (9) |
| 5-Methyl Resorcinol | | | | 97 (8) | (11) |
| 4-Methyl Resorcinol | | | | 97 (8) | (11) |
| TRIHYDRIC PHENOLS | | 60 (1) | | LOW (3) | |
| Phloroglucinol | | | | | |

*Numbers in parenthesis refer to references.

- (1) Reference 78.
- (2) Reference 8.
- (3) Reference 15.
- (4) Reference 80.
- (5) Reference 102.
- (6) Reference 6.
- (7) Reference 93.
- (8) Assumed same removal as catechol.
- (9) Reference 25.
- (10) Assumed value based on rejection of non-phenolic organics in reference 25.

- (11) Assumed same removal as catechol and resorcinol.
- (12) DOE/ECT, unpublished.
- (13) Assumed same removal as nitrogen heterocyclics.
- (14) Assumed values: assuming little degradation, but absorption/sedimentation removes 90% compounds (>2 rings) and 70% compounds (1,2 rings).
- (15) Assumed same value as phenanthrene.
- (16) Assumed same value as for pyridine.
- (17) Assumed value.
- (18) Reference 51.
- (19) Average value, reference 34.
- (20) Reference 17.
- (21) Average value, reference 35.

TABLE 7 (Cont.)

| COMPOUND | GAS LIQUOR SEPARATOR (RE _X -LS) | PHENO- SOLVAN (RE _X -PH) | AMMONIA RECOVERY (RE _X -AR) | BIOLOGICAL TREATMENT (RE _X -BT) | REVERSE OSMOSIS (RE _X -RO) |
|----------------------------------|--|---|--|--|---|
| <u>POLYAROMATIC HYDROCARBONS</u> | 0 | 15 (1) | 0 | 30-90 ⁺ (4,12) | 99 (15) |
| Acenaphthylene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Acridine | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Anthracene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Benz(a)anthracene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Benzo(g,h,i)perylene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Benzo(a)pyrene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Benzo(e)pyrene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Biphenyl | ↓ | ↓ | ↓ | 70 (5) | ↓ |
| Chrysene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Fluoranthene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Fluorene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Indan | ↓ | ↓ | ↓ | 70 (14) | ↓ |
| Indene | ↓ | ↓ | ↓ | 75 (5) | ↓ |
| Naphthalene | ↓ | ↓ | ↓ | 70 (14) | ↓ |
| Perylene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| Phenanthrene | ↓ | ↓ | ↓ | 90 (14) | ↓ (9) |
| Pyrene | ↓ | ↓ | ↓ | 90 (14) | ↓ |
| <u>SULFUR HETEROCYCLICS</u> | | 15 (2) | | | 63 (10) |
| Methylthiophene | | ↓ | | 70 (14) | ↓ |
| Thiophene | | ↓ | | 70 (14) | ↓ |
| <u>NITROGEN HETEROCYCLICS</u> | | | | | 74 (16) |
| 2,4-Dimethylpyridine | | 99 (2) | | 90+ (5) | ↓ |
| 2,5-Dimethylpyridine | | 99 (2) | | 90+ (5) | ↓ |
| 2-Methylpyridine | | 99 (2) | | 90+ (5) | ↓ |
| 3-Methylpyridine | | 99 (2) | | 90+ (5) | ↓ |
| 4-Methylpyridine | | 99 (2) | | 90+ (5) | ↓ |
| Pyridine | | 99 (2) | | 99+ (5) | ↓ (9) |
| Quinoline | | 99 (13) | | 90 (13) | ↓ |
| <u>OXYGEN HETEROCYCLICS</u> | | | | | 74 (13) |
| Benzo-furan | | 15 (1) | | 90 (13) | ↓ |
| Dibenzo-furan | | 15 (1) | | 90 (13) | ↓ |
| <u>MERCAPTANS</u> | | | | | 90 (17) |
| Methanethiol | | 15 (1) | | 90 (17) | ↓ |
| <u>AROMATIC AMINES</u> | | | | | 63 (10) |
| Aniline | | 15 (1) | | 95 (5) | ↓ |
| <u>TRACE ELEMENTS</u> | | 0 | | 30-90 (12,18) | |
| Arsenic | ↓ | ↓ | ↓ | 50 (17) | 89 (21) |
| Boron | ↓ | ↓ | ↓ | 50 (17) | 90 (21) |
| Beryllium | ↓ | ↓ | ↓ | 50 (17) | 90 (17) |
| Cadmium | ↓ | ↓ | ↓ | 32.5 (19) | 90 (17) |
| Fluorine | ↓ | ↓ | ↓ | 50 (17) | 93 (6) |
| Lead | ↓ | ↓ | ↓ | 70 (19) | 60 (21) |
| Mercury | ↓ | ↓ | ↓ | 47.5 (19) | 90 (17) |
| Manganese | ↓ | ↓ | ↓ | 50 (17) | 100 (6) |
| Nickel | ↓ | ↓ | ↓ | 27.5 (19) | 98 (6) |
| Vanadium | ↓ | ↓ | ↓ | 45 (20) | 95 (21) |

F_{W-43} is the flow rate of the water component of Stream 43, i.e., 1,993,000 lb/hr.

C_{TE-27} is the concentration of each trace element in the water component of Stream 27 (in ppm).

F_{W-27} is the flow rate of the water component of Stream 27, i.e., 1,030 lb/hr.

RE_{TE-LS} is the efficiency of removal of each trace element in the gas liquor separator (in %).

F_{W-46} is the flow rate of the water component of Stream 46, i.e., 1,991,000 lb/hr.

Estimation of Trace Element Concentrations in Stream 48

The concentration of each trace element in Stream 48 was estimated using the following equation:

$$C_{TE-48} = \frac{(C_{TE-46})(F_{W-46}) \frac{[100 - RE_{TE-PH}]}{100}}{F_{W-48}}$$

where-

C_{TE-48} is the concentration of each trace element in Stream 48 (in ppm).

C_{TE-46} is the concentration of each trace element in Stream 46 (in ppm).

F_{W-46} is the flow rate of the water component of Stream 46, i.e., 1,991,000 lb/hr.

RE_{TE-PH} is the efficiency of removal of each trace element by the Phenosolvan control unit (in %).

F_{W-48} is the flow rate of the water component of Stream 48, i.e., 1,990,000 lb/hr.

Estimation of Trace Element Concentrations in Stream 50

The concentration of each trace element in Stream 50 was estimated using the following equation:

$$C_{TE-50} = \frac{(C_{TE-48})(F_{W-48}) [100 - RE_{TE-AR}]}{F_{W-50} \cdot 100}$$

where-

C_{TE-50} is the concentration of each trace element in Stream 50, (in ppm).

C_{TE-48} is the concentration of each trace element in Stream 48, (in ppm).

F_{W-48} is the flow rate of the water component of Stream 48, i.e., 1,990,000 lb/hr.

RE_{TE-AR} is the efficiency of removal of each trace element by the Ammonia Recovery unit (in %).

F_{W-50} is the flow rate of the water component of Stream 50, i.e., 1,991,000 lb/hr.

Estimation of Trace Element Concentrations in Stream 52

The concentration of each trace element in Stream 52 was estimated using the following equation:

$$C_{TE-52} = \frac{(C_{TE-50})(F_{W-50}) [100 - RE_{TE-BT}]}{F_{W-52} \cdot 100}$$

where-

C_{TE-52} is the concentration of each trace element in Stream 52, (in ppm).

C_{TE-50} is the concentration of each trace element in Stream 50, (in ppm).

F_{W-50} is the flow rate of the water component of Stream 50, i.e., 1,991,000 lb/hr.

RE_{TE-BT} is the efficiency of removal of each trace element by the Biological Treatment unit, (in %).

F_{W-52} is the flow rate of the water component of Stream 52, i.e., 1,962,000 lb/hr.

Estimation of Trace Element Concentrations in Stream 38-

The concentration of each trace element in Stream 38 was assumed to be equal to the concentration of elements in column volume leachate fractions of ash from Lurgi gasification of Montana Rosebud coal.⁹¹ The data were assumed to give a good approximation of Wyoming subbituminous coal as the concentration of almost all trace elements were essentially equal in both coals, as shown in Table 4. The concentration of each trace element in Stream 38 is presented in Table 6.

Estimation of Trace Element Concentrations in Stream 53-

The concentrations of each trace element in Stream 53 was estimated using the following equation:

$$C_{TE-53} = \frac{[(C_{TE-52})(F_{W-52}) + (C_{TE-38})(F_{W-38})] [RE_{TE-RO}]}{F_{W-53} \cdot 100}$$

where-

C_{TE-53} is the concentration of each trace element in Stream 53, (in ppm).

C_{TE-52} is the concentration of each trace element in Stream 52, (in ppm).

F_{W-52} is the flow rate of the water component of Stream 52, i.e., 1,962,000 lb/hr.

C_{TE-38} is the concentration of each trace element in Stream 38, (in ppm).

F_{W-38} is the flow rate of the water component of Stream 38, i.e., 879,000 lb/hr.

RE_{TE-RO} is the efficiency of removal of each trace element by the Reverse Osmosis unit (in %).

F_{W-53} is the flow rate of the water component of Stream 53 which is assumed to be 20% of the sum of the water components of Streams 52 and 38, i.e., 568,200 lb/hr.

Estimation of Trace Element Concentrations in Stream 54-

The concentration of each trace element in Stream 54 was estimated using the following equation:

$$C_{TE-54} = \frac{(C_{TE-52})(F_{W-52}) + (C_{TE-38})(F_{W-38}) \frac{[100-RE_{TE-RO}]}{100}}{F_{W-54}}$$

where-

C_{TE-54} is the concentration of each trace element in Stream 54 (in ppm).

C_{TE-52} is the concentration of each trace element in Stream 52 (in ppm).

F_{W-52} is the flow rate of the water component of Stream 52, i.e., 1,962,000 lb/hr.

C_{TE-38} is the concentration of each trace element in Stream 38 (in ppm)

F_{W-38} is the flow rate of the water component of Stream 38, i.e., 879,000 lb/hr.

RE_{TE-RO} is the efficiency of removal of each trace element by the Reverse Osmosis unit (in %).

F_{W-54} is the flow rate of the water component of Stream 54, assumed to be equal to 80% of the sum of the water components of Stream 52 and 38, i.e., 2,273,000 lb/hr.

Estimation of Trace Element Concentrations in Stream 31-

The concentration of each trace element in Stream 31 (make-up water to utilities generation) was assumed to be zero.

Estimation of Trace Element Concentrations in Streams 31+54-

The concentration of each trace element in Streams 31+54 was estimated using the following equation:

$$C_{TE-(31+54)} = \frac{(C_{TE-31})(F_{W-31}) + (C_{TE-54})(F_{W-54})}{(F_{W-31}) + (F_{W-54})}$$

where-

- $C_{TE-(31+54)}$ is the concentration of each trace element in Stream 31 + 54, (in ppm).
- C_{TE-31} is the concentration of each trace element in Stream 31, (in ppm).
- F_{W-31} is the flow rate of the water component of Stream 31, i.e., 1,180,000 lb/hr.
- C_{TE-54} is the concentration of each trace element in Stream 54, (in ppm).
- F_{W-54} is the flow rate of the water component of Stream 54, assumed to be equal to 80% of the sum of the water components of Streams 52 and 38, i.e., 2,273,000 lb/hr.

2.2.2 Estimation of Organic Compounds in Liquid Streams

Data on the types and concentrations of organic compounds in liquid streams were very limited. No data could be found regarding organic constituents in process waters from Lurgi/Fischer-Tropsch processing of Wyoming subbituminous coal.

In order to provide estimates of the types and concentrations of organic compounds which may be present in liquid streams from wastewater treatment units, data from two sources were used.

- A search of the literature indicated that limited data were available from the SASOL operation in South Africa and Lurgi gasifiers in Westfield, Scotland.⁹² Although the coal feed type, liquid stream flow rates, and operating conditions may not be the same as those specified for the conceptual plant used in this analysis, the available data were assumed to be representative. The data are presented in Table 6.
- Ongoing studies sponsored by the Environmental Protection Agency recently have characterized the organic components of tars produced by gasification of Wyoming subbituminous coal. Although there are differences between the conceptual plant and the EPA study regarding reactor configuration and operating conditions, the EPA data were assumed to be useful in providing preliminary estimates of the types and concentrations of organic compounds which may be produced by Lurgi gasification of Wyoming subbituminous coal. Parameter values for the EPA

study, Lurgi gasifiers and the conceptual plant configuration are compared in Table 8. Data from the EPA Study which were used in estimating the types and concentrations of organic compounds in liquid streams are summarized in Table 9 (column 4).

These two sources of data were used to estimate the concentrations of organics in Stream 43. Estimates of concentrations of organics in subsequent liquid streams in the wastewater treatment units were estimated by sequentially applying removal efficiencies of control units for each organic constituent to calculate the amount of each constituent which would remain in the liquid stream, and then dividing the quantity of constituent by the flow rate of the water component of each stream.*

Estimation of Organics Concentrations in Stream 43-

As in the case of trace elements, Stream 43 was considered to be the crucial liquid stream for the analysis; it was the largest, most highly contaminated liquid waste stream in the plant, and its composition largely determined the composition and flow rates of subsequent streams from the wastewater treatment units.

Characterization of organics in Stream 43 was accomplished using data from the SASOL and Westfield, Scotland plants⁹² and gasification screening tests¹⁹ described in the previous section. The SASOL and Westfield data, presented in Table 6, were assumed to be representative of Stream 43 and were used directly. The tar characterization data from the gasification screening tests were used in the following way to estimate quantities of organics in Stream 43.

- 1) It was assumed that the compounds detected in the tar produced by fixed-bed gasification of Wyoming subbituminous coal and the rates of production would be representative of Lurgi gasification of Wyoming subbituminous coal.

*The water component of each stream was used rather than the whole stream because it was assumed that the total quantity of each constituent was dissolved in the water component.

TABLE 8

COMPARISON OF PARAMETER VALUES FOR GASIFICATION REACTORS

| | RTI Test #33 ¹⁹ | Lurgi ¹⁹ | Conceptual Plant ⁸⁷ |
|-----------------------------------|-------------------------------|----------------------|--------------------------------|
| Air/Coal, g/g | 1.5 | 3.0 | 1.2 |
| Steam/Coal, g/g | .36 | 1.5 | .89 |
| Carbon Conversion, % | 98.9 | 95 | 99.5 |
| Coal Residence Time (min) | 110 | 60 | No Data |
| Tar Produced, g/g | .012 | No Data | .019 |
| Gas Produced, SCF/lb | 35 | 52 | 38 |
| HHV of Raw Gas, Btu/SCF | 201 | 195 | No Data |
| Throughput, lb/hr/ft ² | 45 | 248 | No Data |
| Coal Type | Wyoming Subbit | New Mexico Subbit | Wyoming Subbit |
| Pressure, Psia | 200 | 300 | 450 |
| Mesh Size | 8 x 16 | 1.75" x .08" | 1/4" x 2" |
| Maximum Temperature °C | 1040 | No Data | No Data |
| Heatup Time to 800°C, Min. | 8 | NA* | NA* |

* NA-Not applicable

TABLE 9
ESTIMATION OF COMPOUND QUANTITIES BASED ON TAR
CONSTITUENTS AND COMPOUND SOLUBILITY*

| 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 |
|------------------------|--|---|---|---------------------------------|---------------------------------|------------------------------------|-----------------------------|--|--|-------------------------------------|
| Compound | Identified in Tar, Oil or Liqueur (10 ⁶ lb) | Identified in Tar From Spent Subbituminous Coal | Compound Production Rate (1) (gal/yr of coal) | Compound Production (2) (lb/yr) | Maximum Concentration (3) (ppm) | Compound Solubility in Water (ppm) | Number of Rings in Compound | Assumed Percent of Saturation Level (11) | Assumed Concentration Solubility (ppm) | Assumed Concentration in Stream (3) |
| Methanol | | / | 2.3x10 ⁻⁵ | 43.7 | 70.6 | 23,100 (9) | - | - | - | 30. |
| Ethyl Benzene | 0 (10) | / | 1.5x10 ⁻⁶ | 285.2 | 136.6 | 152 (7) | 1 | 10 | 15. | 15. |
| Methylthiophene | | / | 1.6x10 ⁻⁶ | 100.1 | 89.7 | 1 (5) (6) | 1 | 10 | (9.5)** | (9.5)** |
| Phenol | 1 (12) | / | 1.6x10 ⁻³ | 3062.1 | 1475.7 | 82,000 (6) | 1 | 10 | 8,200 | 1635. |
| Thiophene | 0 (10) | / | 3.0x10 ⁻⁵ | 57.0 | 26.9 | 1 (5) (6) | 1 | 10 | (3.) | (3.) |
| Toluene | 0 (10) | / | 2.2x10 ⁻³ | 4182.9 | 1976.1 | 500 (6) | 1 | 10 | 50. | 50. |
| 1,2,4-trimethylbenzene | 0 (10) | / | - | - | - | 57 (7) | 1 | 10 | 5.7 | 6. |
| Xylene (8) | 0 (10) | / | 8.0x10 ⁻⁶ | 1521.1 | 717.9 | 175 (7) | 1 | 10 | 17.5 | 18. |
| Benzofuran | 0 (10) | / | 1.0x10 ⁻⁶ | 190.1 | 89.7 | 1 (5) (6) | 2 | 10 | ** | ** |
| Biphenyl | 0 (10) | / | 6.3x10 ⁻⁶ | 12.0 | 5.7 | 7 (4) | 2 | 10 | 0.7 | 0.7 |
| Indan | 0 (10) | / | 7.6x10 ⁻⁵ | 144.5 | 68.1 | 88.9 (13) | 2 | 10 | 0.7 | 9.0 |
| Indole | 0 (10) | / | 5.5x10 ⁻⁶ | 1065.7 | 497.5 | 1 (5) (6) | 2 | 10 | ** | ** |
| Naphthalene | 0 (10) | / | 1.5x10 ⁻⁶ | 665.5 | 316.1 | 32 (4) | 2 | 10 | 3.2 | 3. |
| Quinoline | 0 (10) | / | 5.3x10 ⁻⁵ | 96.9 | 45.7 | 6,100 (4) | 2 | 10 | 610. | 65. |
| Acenaphthylene | | / | 4.9x10 ⁻⁵ (8) | 93.2 | 43.9 | 4 (4) | 3 | 10 | 0.4 | 0.4 |
| Acridine | | / | 2.6x10 ⁻⁶ | 4.6 | 2.2 | vss (5) (6) | 3 | 10 | - | 2.2 |
| Anthracene | | / | 8.5x10 ⁻⁵ | 161.6 | 76.3 | 1 (4) | 3 | 10 | 0.1 | 0.1 |
| Dibenzofuran | 0 (10) | / | 5.9x10 ⁻⁵ | 112.2 | 52.9 | 1 (4) | 3 | 10 | - | 0.2 |
| Fluorene | | / | 5.7x10 ⁻⁵ | 108.4 | 51.1 | 2 (6) | 1 | 10 | 0.2 | 0.2 |
| Phenanthrene | | / | 4.6x10 ⁻⁵ | 83.7 | 39.5 | 1 (6) | 1 | 10 | 0.1 | 0.1 |
| Benzo(a)Anthracene | | / | 5.2x10 ⁻⁶ (8) | 9.9 | 4.7 | 0.01 (4) | 4 | 100 | 0.01 | 0.01 |
| Chrysene | | / | 2.0x10 ⁻⁵ | 38.3 | 17.9 | 0.002 (4) | 4 | 100 | 0.002 | 0.002 |
| Fluoranthene | | / | 3.2x10 ⁻⁵ | 60.8 | 28.7 | 0.2 (4) | 4 | 100 | 0. | 0. |
| Pyrene | | / | 1.5x10 ⁻⁵ | 66.5 | 31.6 | 0.2 (4) | 4 | 100 | 0.2 | 0.2 |
| Benzo(a)pyrene | | / | 1.6x10 ⁻⁶ (8) | 2.7 | 1.3 | 0.004 (4) | 5 | 100 | 0.004 | 0.004 |
| Benzo(b)pyrene | | / | 8.7x10 ⁻⁷ (8) | 1.6 | 0.8 | 0.004 (4) | 5 | 100 | 0.004 | 0.004 |
| Perylene | | / | 1.3x10 ⁻⁵ | 26.7 | 11.7 | 0.004 (4) | 5 | 100 | 0.004 | 0.004 |
| Benzo(g,h,i)perylene | | / | 2.9x10 ⁻⁷ (11) | 0.6 | 0.3 | 0.0003 (4) | 6 | 100 | 0.0003 | 0.0003 |

(1) Reference 19.
 (2) Total Production = (Pollutant Production Rate)(1,901,324 lb coal/yr).
 (3) Maximum Concentration = (Total Production) / (Flow Rate of Stream 43).
 (4) Reference 9.
 (5) vss = very slightly soluble; as = slightly soluble; l = insoluble.
 (6) Reference 75.
 (7) Reference 4.
 (8) Reference 79.
 (9) Reference 113.
 (10) Reference 40.
 (11) Reference 45.
 (12) Reference 92.
 *Numbers in parentheses refer to footnotes.
 **Estimates highly uncertain due to lack of data on compound solubility.

- 2) It was further assumed that the concentration of each compound in the gas liquor would be limited by either the production rate of the compound or the solubility of the compound in water.
- 3) The quantity of each compound produced as a function of the quantity of coal fed to the gasifier was identified in the study. The data are presented in Table 9 (column 4).
- 4) It was assumed that the total quantity of each organic compound in the tar dissolved in the water component of Stream 43. The resultant concentration of each organic compound in the water component of Stream 43 was calculated using the following equation:

$$C_{O-43} = \frac{(R_o)(F_{C-2})}{F_{W-43}}$$

where-

C_{O-43} is the concentration of each organic compound in the water component of Stream 43, (in ppm).

R_o is the rate of production of each organic compound per unit of coal charged to the gasifier (lb compound produced/lb of coal gasified).

F_{C-2} is the flow rate of whole coal to the gasifier, i.e., Stream 2; (1,901,324 lb/hr).

F_{W-43} is the flow rate of the water component of Stream 43, i.e., 1,993,000 lb/hr.

The results are presented in Table 9, Column 6.

- 5) The concentration of each compound, which would result if the limiting factor were the solubility of the compound in water, was estimated:

$$C_{O-43} = \frac{(S_{O-W})(L_{O-W})}{100\%}$$

where-

C_{O-43} is the concentration of each organic compound in Stream 43 if the solubility of the compound in water were the limiting factor (in ppm).

S_{O-W} is the solubility of the compound in water (in ppm).

L_{O-W} is the assumed level of solubility (in %) which would be attained by each compound as a function of molecular size. Based upon findings at Oak Ridge National Laboratory⁴⁵ it was assumed that for compounds with 4 or more rings $L_{O-W} = 100\%$, and for compounds with 3 or fewer rings $L_{O-W} = 10\%$.

The results are presented in Table 9, Column 10.

- 6) The concentration of each organic compound in the water component of Stream 43 was assumed to be the lesser of the two concentrations estimated in steps 4 and 5. Concentrations in Stream 43 used in the analysis are presented in Table 9, Column 11.

Estimation of Organics Concentrations in Stream 27-

Concentrations of organic compounds in Stream 27 were assumed to be zero due to lack of data on specific compounds.

Estimation of Organics Concentrations in Stream 46-

The concentration of each organic compound in the water component of stream 46 was estimated:

$$C_{O-46} = \frac{(C_{O-43})(F_{W-43}) + (C_{O-27})(F_{W-27}) [100 - RE_{O-LS}]}{F_{W-46} \cdot 100}$$

where-

C_{O-46} is the concentration of each compound in the water component of Stream 46, (in ppm).

C_{O-43} is the concentration of each compound in the water component of Stream 43, (in ppm).

F_{W-43} is the flow rate of the water component of Stream 43, i.e., 1,993,000 lb/hr.

C_{O-27} is the concentration of each organic compound in Stream 27, i.e., assumed = 0.

F_{W-27} is the flow rate of the water component of Stream 27.

RE_{O-LS} is the efficiency of removal of each compound in the tar/oil separation unit (in %).

F_{W-46} is the flow rate of the water component of Stream 46, i.e., 1,991,000 lb/hr.

Estimation of Organics Concentrations in Stream 48-

The concentration of each compound in Stream 48 was estimated using the following:

$$C_{O-48} = \frac{(C_{O-46})(F_{W-46}) \left[\frac{100-RE_{O-PH}}{100} \right]}{F_{W-48}}$$

where-

- C_{O-48} is the concentration of each compound in the water component of Stream 48 (in ppm).
- C_{O-46} is the concentration of each compound in the water component of Stream 46.
- F_{W-46} is the flow rate of the water component of Stream 46, i.e., 1,991,000 lb/hr.
- RE_{O-PH} is the efficiency of removal of each compound in the Phenosolvan unit (in %).
- F_{W-48} is the flow rate of the water component of Stream 48, i.e., 1,990,000 lb/hr.

Estimation of Organics Concentrations in Stream 50-

The concentration of each organic compound in Stream 50 was estimated using the following:

$$C_{O-50} = \frac{(C_{O-48})(F_{O-48}) \left[\frac{100-RE_{O-AR}}{100} \right]}{F_{W-50}}$$

where-

- C_{O-50} is the concentration of each compound in the water component of Stream 50 (in ppm).
- C_{O-48} is the concentration of each compound in the water component of Stream 48 (ppm).
- F_{W-48} is the flow rate of the water component of Stream 48, i.e., 1,990,000 lb/hr.
- RE_{O-AR} is the efficiency of removal of each compound in the ammonia recovery unit (in %).
- F_{W-50} is the flow rate of the water component of Stream 50, i.e., 1,991,000 lb/hr.

Estimation of Organic Concentrations in Stream 52-

The concentration of each compound in Stream 52 was estimated using the following:

$$C_{O-52} = \frac{(C_{O-50})(F_{W-50}) \left[\frac{100-RE_{O-BT}}{100} \right]}{F_{W-52}}$$

where-

C_{O-52} is the concentration of each compound in the water component of Stream 52 (in ppm).

C_{O-50} is the concentration of each compound in the water component of Stream 50 (ppm).

F_{W-50} is the flow rate of the water component of Stream 50, i.e., 1,991,000 lb/hr.

RE_{O-BT} is the efficiency of removal of each compound in the Biological Treatment unit (in %).

F_{W-52} is the flow rate of the water component of Stream 52, i.e., 1,962,000 lb/hr.

Estimation of Organic Concentrations in Stream 38-

Concentrations of organic compounds in Stream 38 were assumed to be zero, due to lack of data.

Estimation of Organics Concentrations in Stream 53-

The concentration of each organic compound in Stream 53 was estimated using the following:

$$C_{O-53} = \frac{[(C_{O-52})(F_{W-52}) + (C_{O-38})(F_{W-38})] \left[\frac{RE_{O-RO}}{100} \right]}{F_{W-53}}$$

where-

C_{O-53} is the concentration of each compound in the water component of Stream 53 (in ppm).

C_{O-52} is the concentration of each compound in the water component of Stream 52 (in ppm).

F_{W-52} is the flow rate of the water component of Stream 52, i.e., 1,962,000 lb/hr.

C_{O-38} is the concentration of each compound in Stream 38, assumed = 0.

F_{W-38} is the flow rate of the water component of Stream 38, i.e., 879,000 lb/hr.

RE_{O-RO} is the efficiency of removal of each compound in the Reverse Osmosis unit (in %).

F_{W-53} is the flow rate of the water component of Stream 53, assumed to be equal to 20% of the combined water flow rate of Streams 38 and 52, i.e., 568,200 lb/hr.

Estimation of Organic Concentrations in Stream 54-

Concentrations of each compound in Stream 54 were estimated using the following:

$$C_{O-54} = \frac{[(C_{O-52})(F_{W-52}) + (C_{O-38})(F_{O-38})] [100 - RE_{O-RO}]}{F_{W-54} \cdot 100}$$

where-

C_{O-54} is the concentration of each compound in the water component of Stream 54 (in ppm).

C_{O-52} is the concentration of each compound in the water component of Stream 52 (in ppm).

F_{W-52} is the flow rate of the water component of Stream 52, i.e., 1,962,000 lb/hr.

C_{O-38} is the concentrations of each organic compound in Stream 38, assumed = 0.

RE_{O-RO} is the efficiency of removal of each compound in the Reverse Osmosis unit (in %).

F_{W-54} is the flow rate of the water component of Stream 54, assumed to be equal to 80% of the combined water flow rate of Stream 52 and 38, i.e., 2,273,000 lb/hr.

Estimation of Organic Concentrations in Stream 31-

Concentration of organics in treated makeup water (Stream 31) was assumed to be equal to zero.

Estimation of Organic Concentrations in Streams 31+54-

The concentration of each organic compound in Streams 31+54 was estimated using the following:

$$C_{O-(31+54)} = \frac{(C_{O-31})(F_{W-31}) + (C_{O-54})(F_{W-54})}{F_{W-31} + F_{W-54}}$$

where-

$C_{O-(31+54)}$ is the concentration of each compound in the water component of Streams (31+54).

C_{O-31} is the concentration of each compound in Stream 31, assumed = 0.

F_{W-31} is the flow rate of the water component of Stream 31, i.e., 1,179,700 lb/hr.

C_{O-54} is the concentration of each compound in the water component of Stream 54.

F_{W-54} is the flow rate of the water component of Stream 54, i.e., 2,273,000 lb/hr.

$F_{W-31} + F_{W-54}$ is the sum of the flow rate of the water component of Streams 31+54, i.e., 3,452,000 lb/hr.

2.2.3 Results

The estimated concentrations and flow rates of each stream constituent, including both trace elements and organic compounds, are presented by stream in Table 10 and 11.

TABLE 10

CONCENTRATION IN STREAM (PPM)

| COMPOUNDS/ TRACE ELEMENTS | 27 | 31 | 38 | 43 | 46 | 48 | 50 | 52 | 53 | 54 | 54151 |
|------------------------------|---------|-----|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Arsenic | 1.0E-02 | 0.0 | 2.0E-02 | 5.5E-01 | 5.5E-01 | 5.5E-01 | 5.5E-01 | 2.0E-01 | 8.9E-01 | 2.0E-02 | 1.0E-02 |
| Benzilium | 1.0E-02 | 0.0 | 0.0 | 2.9E-01 | 2.9E-01 | 2.9E-01 | 2.9E-01 | 1.5E-01 | 4.6E-01 | 1.3E-02 | 8.5E-03 |
| Boron | 0.0 | 0.0 | 0.0 | 1.9E100 | 3.9E100 | 1.9E100 | 1.9E100 | 9.0E-01 | 3.1E100 | 8.5E-02 | 5.4E-02 |
| Cadmium | 6.0E-03 | 0.0 | 5.4E-04 | 2.5E-01 | 2.5E-01 | 2.5E-01 | 2.5E-01 | 1.7E-01 | 5.3E-01 | 1.5E-02 | 9.0E-03 |
| Fluorine | 0.0 | 0.0 | 0.0 | 2.0E101 | 2.0E101 | 2.0E101 | 2.0E101 | 1.0E101 | 3.3E101 | 6.2E-01 | 4.1E-01 |
| Lead | 2.3E-02 | 0.0 | 9.0E-02 | 1.4E-01 | 1.4E-01 | 1.4E-01 | 1.4E-01 | 4.3E-02 | 1.7E-01 | 2.9E-02 | 1.9E-02 |
| Manganese | 1.1E102 | 0.0 | 5.0E-01 | 4.7E-03 | 4.2E-02 | 4.2E-02 | 4.2E-02 | 3.1E-02 | 8.0E-01 | 0.0 | 0.0 |
| Mercury | 2.7E-02 | 0.0 | 3.0E-04 | 4.9E-02 | 4.9E-02 | 4.9E-02 | 4.9E-02 | 2.6E-02 | 8.1E-02 | 2.2E-03 | 1.5E-03 |
| Nickel | 3.2E-01 | 0.0 | 3.5E-02 | 3.0E-02 | 3.9E-02 | 3.9E-02 | 3.9E-02 | 2.0E-02 | 1.8E-01 | 7.4E-04 | 5.0E-04 |
| Vanadium | 0.0 | 0.0 | 0.0 | 9.6E-03 | 9.6E-03 | 9.6E-03 | 9.6E-03 | 5.9E-03 | 1.0E-02 | 2.3E-04 | 1.5E-04 |
| Acetic acid | 0.0 | 0.0 | 0.0 | 1.7E102 | 1.7E102 | 1.5E102 | 1.5E102 | 7.4E100 | 1.3E101 | 3.1E100 | 2.1E100 |
| Aniline | 0.0 | 0.0 | 0.0 | 1.2E101 | 1.2E101 | 1.2E101 | 1.2E101 | 6.1E-03 | 1.3E-02 | 1.9E-03 | 1.3E-03 |
| Benzoic acid | 0.0 | 0.0 | 0.0 | 1.3E101 | 1.3E101 | 1.1E101 | 1.1E101 | 5.6E-01 | 1.2E100 | 1.0E-01 | 1.2E-01 |
| Catechol | 0.0 | 0.0 | 0.0 | 5.5E102 | 5.5E102 | 2.2E102 | 2.2E102 | 6.7E100 | 1.7E101 | 1.4E100 | 9.5E-01 |
| Benzoic Acid | 0.0 | 0.0 | 0.0 | 1.0E100 | 1.0E100 | 0.5E-01 | 0.5E-01 | 4.3E-02 | 9.4E-02 | 1.4E-02 | 9.1E-03 |
| Pentanoic Acid | 0.0 | 0.0 | 0.0 | 1.2E101 | 1.2E101 | 1.0E101 | 1.0E101 | 5.2E-01 | 1.1E100 | 1.7E-01 | 1.1E-01 |
| Phenol | 0.0 | 0.0 | 0.0 | 3.1E103 | 3.1E103 | 1.6E101 | 1.6E101 | 1.6E-01 | 1.0E-01 | 4.2E-02 | 2.0E-02 |
| Propionic Acid | 0.0 | 0.0 | 0.0 | 2.6E101 | 2.6E101 | 2.2E101 | 2.2E101 | 1.1E100 | 2.5E100 | 3.9E-01 | 2.2E-01 |
| Pyridine | 0.0 | 0.0 | 0.0 | 1.2E102 | 1.2E102 | 1.2E100 | 1.2E100 | 1.2E-02 | 3.0E-02 | 2.7E-03 | 1.0E-03 |
| Resorcinol | 0.0 | 0.0 | 0.0 | 2.0E102 | 2.0E102 | 1.1E102 | 1.1E102 | 2.2E100 | 5.0E100 | 4.0E-01 | 3.2E-01 |
| 2-Naphthol | 0.0 | 0.0 | 0.0 | 3.4E102 | 3.4E102 | 1.7E100 | 1.7E100 | 5.2E-02 | 1.2E-01 | 1.2E-02 | 7.7E-03 |
| 2-Methylpropanoic Ac | 0.0 | 0.0 | 0.0 | 2.0E100 | 2.0E100 | 1.7E100 | 1.7E100 | 8.6E-02 | 1.9E-01 | 2.0E-02 | 1.0E-02 |
| 2-Methylpyridine | 0.0 | 0.0 | 0.0 | 7.0E101 | 7.0E101 | 7.0E-01 | 7.0E-01 | 7.1E-02 | 1.0E-01 | 1.6E-02 | 1.1E-02 |
| 2,4-Dimethylpyridine | 0.0 | 0.0 | 0.0 | 1.0E100 | 1.0E100 | 1.0E-02 | 1.0E-02 | 1.0E-03 | 2.6E-03 | 2.4E-04 | 1.5E-04 |
| 2,4-Xylenol | 0.0 | 0.0 | 0.0 | 1.2E102 | 1.2E102 | 6.0E-01 | 6.0E-01 | 1.2E-01 | 3.6E-01 | 1.0E-02 | 6.2E-03 |
| 2,5-Dimethylpyridine | 0.0 | 0.0 | 0.0 | 1.0E100 | 1.0E100 | 1.0E-02 | 1.0E-02 | 1.0E-03 | 2.5E-03 | 2.3E-04 | 1.5E-04 |
| 3-Naphthol | 0.0 | 0.0 | 0.0 | 1.0E100 | 2.0E100 | 0.5E-01 | 0.5E-01 | 4.3E-02 | 9.4E-02 | 1.4E-02 | 9.1E-03 |
| 3-Methylcatechol | 0.0 | 0.0 | 0.0 | 4.0E102 | 4.0E102 | 1.5E102 | 1.5E102 | 0.0 | 0.0 | 0.0 | 0.0 |
| 3-Methylphenol | 0.0 | 0.0 | 0.0 | 4.2E102 | 4.2E102 | 2.1E100 | 2.1E100 | 6.4E-02 | 1.9E-01 | 2.2E-03 | 1.7E-03 |

TABLE 10 (CONT.)
CONCENTRATION IN STREAMS (PPM)

| COMPOUNDS/ TRACER ELEMENT | 27 | 31 | 33 | 43 | 44 | 40 | 50 | 52 | 53 | 54 | 54131 |
|------------------------------|-----|-----|-----|---------|---------|---------|---------|---------|---------|---------|---------|
| 3-Methylbutadiene | 0.0 | 0.0 | 0.0 | 2.6E101 | 2.6E101 | 2.6E101 | 2.4E-01 | 3.4E-02 | 6.7E-02 | 5.9E-03 | 3.9E-03 |
| 3,3-Xylool | 0.0 | 0.0 | 0.0 | 5.0E101 | 5.0E101 | 2.5E-01 | 2.5E-01 | 1.4E-01 | 5.0E-01 | 1.4E-02 | 9.1E-03 |
| 3,4-Dimethylbutadiene | 0.0 | 0.0 | 0.0 | 4.5E101 | 4.5E101 | 1.0E101 | 1.0E101 | 5.5E-01 | 1.4E100 | 1.2E-01 | 7.0E-02 |
| 4-Methylbutadiene | 0.0 | 0.0 | 0.0 | 3.6E101 | 3.6E101 | 1.4E101 | 1.4E101 | 4.4E-01 | 1.1E100 | 9.5E-02 | 4.2E-02 |
| 4-Methylbutadiene | 0.0 | 0.0 | 0.0 | 3.9E102 | 3.9E102 | 1.5E102 | 1.5E102 | 4.7E100 | 1.2E101 | 1.0E100 | 6.7E-01 |
| 4-Methylbutadiene | 0.0 | 0.0 | 0.0 | 3.0E102 | 3.0E102 | 1.5E100 | 1.5E100 | 4.6E-02 | 1.2E-01 | 9.9E-03 | 5.5E-03 |
| 4-Methylbutadiene | 0.0 | 0.0 | 0.0 | 5.0E100 | 6.0E100 | 6.0E-02 | 6.0E-02 | 6.1E-03 | 1.6E-02 | 1.4E-03 | 9.0E-04 |
| 5-Methylbutadiene | 0.0 | 0.0 | 0.0 | 4.5E101 | 4.5E101 | 2.6E101 | 2.6E101 | 7.9E-01 | 2.1E100 | 1.7E-01 | 1.1E-01 |
| Acetophenone | 0.0 | 0.0 | 0.0 | 4.0E-01 | 4.0E-01 | 3.4E-01 | 3.4E-01 | 3.5E-02 | 1.2E-01 | 3.0E-04 | 2.0E-04 |
| Anthracene | 0.0 | 0.0 | 0.0 | 1.0E-01 | 1.0E-01 | 0.5E-02 | 0.5E-02 | 8.6E-03 | 3.0E-02 | 7.5E-05 | 4.9E-05 |
| Benzo(a)anthracene | 0.0 | 0.0 | 0.0 | 1.0E-02 | 1.0E-02 | 0.5E-03 | 0.5E-03 | 0.6E-04 | 3.0E-03 | 2.5E-05 | 4.9E-06 |
| Benzo(a)pyrene | 0.0 | 0.0 | 0.0 | 4.0E-03 | 4.0E-03 | 3.4E-03 | 3.4E-03 | 3.5E-04 | 1.2E-03 | 3.0E-06 | 3.0E-06 |
| Benzo(b)pyrene | 0.0 | 0.0 | 0.0 | 4.0E-03 | 4.0E-03 | 3.4E-03 | 3.4E-03 | 3.5E-04 | 1.2E-03 | 3.0E-06 | 3.0E-06 |
| Benzo(k)fluoranthene | 0.0 | 0.0 | 0.0 | 3.0E-04 | 3.0E-04 | 2.6E-04 | 2.6E-04 | 2.6E-05 | 0.9E-05 | 2.2E-07 | 1.5E-07 |
| Benzo(l)fluoranthene | 0.0 | 0.0 | 0.0 | 7.0E-01 | 7.0E-01 | 6.0E-01 | 6.0E-01 | 1.0E-01 | 6.2E-01 | 1.6E-03 | 1.0E-03 |
| Chrysene | 0.0 | 0.0 | 0.0 | 2.0E-03 | 2.0E-03 | 1.7E-03 | 1.7E-03 | 1.7E-04 | 5.9E-04 | 1.5E-06 | 9.0E-07 |
| Fluoranthene | 0.0 | 0.0 | 0.0 | 1.0E-01 | 1.0E-01 | 0.5E-02 | 0.5E-02 | 0.6E-03 | 2.2E-02 | 1.9E-03 | 1.3E-03 |
| Ethylbenzene | 0.0 | 0.0 | 0.0 | 1.5E101 | 1.5E101 | 1.3E101 | 1.3E101 | 1.3E100 | 2.9E100 | 4.2E-01 | 2.4E-01 |
| Fluorene | 0.0 | 0.0 | 0.0 | 2.0E-01 | 2.0E-01 | 1.7E-01 | 1.7E-01 | 1.7E-02 | 5.9E-02 | 1.5E-04 | 9.0E-05 |
| Fluorene | 0.0 | 0.0 | 0.0 | 2.0E-01 | 2.0E-01 | 1.7E-01 | 1.7E-01 | 1.7E-02 | 5.9E-02 | 1.5E-04 | 9.0E-05 |
| Indole | 0.0 | 0.0 | 0.0 | 0.9E100 | 0.9E100 | 7.6E100 | 7.6E100 | 2.3E100 | 7.9E100 | 2.0E-02 | 1.3E-02 |
| Naphthalene | 0.0 | 0.0 | 0.0 | 2.2E101 | 2.2E101 | 1.9E101 | 1.9E101 | 1.9E100 | 5.9E100 | 1.6E-01 | 1.1E-01 |
| Naphthalene | 0.0 | 0.0 | 0.0 | 3.2E100 | 3.2E100 | 2.7E100 | 2.7E100 | 0.5E-01 | 2.0E100 | 7.2E-03 | 4.7E-03 |
| Perylene | 0.0 | 0.0 | 0.0 | 4.0E-04 | 4.0E-04 | 3.4E-04 | 3.4E-04 | 3.5E-05 | 1.2E-04 | 3.0E-07 | 2.0E-07 |
| Phenanthrene | 0.0 | 0.0 | 0.0 | 1.0E-01 | 1.0E-01 | 0.5E-02 | 0.5E-02 | 0.6E-03 | 3.0E-02 | 7.8E-05 | 4.9E-05 |
| Pyrene | 0.0 | 0.0 | 0.0 | 2.0E-01 | 2.0E-01 | 1.7E-01 | 1.7E-01 | 1.7E-02 | 5.9E-02 | 1.5E-04 | 9.0E-05 |
| Quinoline | 0.0 | 0.0 | 0.0 | 9.5E100 | 9.5E100 | 9.5E-02 | 9.5E-02 | 9.7E-03 | 2.5E-02 | 2.2E-03 | 1.4E-03 |
| Indene | 0.0 | 0.0 | 0.0 | 5.0E101 | 5.0E101 | 4.3E101 | 4.3E101 | 4.3E100 | 9.4E100 | 1.4E100 | 9.1E-01 |
| 0-Xylene | 0.0 | 0.0 | 0.0 | 1.0E101 | 1.0E101 | 1.5E101 | 1.5E101 | 1.5E100 | 5.3E100 | 4.0E-01 | 3.2E-01 |

FLOW RATES IN STREAMS (LBS/HR)

| COMPOUND/ ISOMER ELEMENTS | 27 | 31 | 30 | 43 | 32 | 40 | 50 | 52 | 53 | 54 | 54131 |
|------------------------------|---------|-----|---------|---------|---------|---------|---------|---------|---------|---------|---------|
| Acetone | 1.0E-02 | 0.0 | 1.0E-02 | 1.1E100 | 1.1E100 | 1.1E100 | 1.1E100 | 5.5E-01 | 5.1E-01 | 6.3E-02 | 6.3E-02 |
| Benzylamine | 1.6E-03 | 0.0 | 0.0 | 5.0E-01 | 5.0E-01 | 5.0E-01 | 5.0E-01 | 2.9E-01 | 2.5E-01 | 2.9E-02 | 2.9E-02 |
| Bornol | 0.0 | 0.0 | 0.0 | 3.9E100 | 3.9E100 | 3.9E100 | 3.9E100 | 1.9E100 | 1.7E100 | 1.9E-01 | 1.9E-01 |
| Carbonium | 6.2E-06 | 0.0 | 5.2E-04 | 5.0E-01 | 5.0E-01 | 5.0E-01 | 5.0E-01 | 3.4E-01 | 3.0E-01 | 3.4E-02 | 3.4E-02 |
| Fluorine | 0.0 | 0.0 | 0.0 | 4.0E101 | 4.0E101 | 4.0E101 | 4.0E101 | 2.0E101 | 1.9E101 | 1.4E100 | 1.4E100 |
| Linal | 2.9E-03 | 0.0 | 7.9E-02 | 2.0E-01 | 2.0E-01 | 2.0E-01 | 2.0E-01 | 0.4E-02 | 9.0E-02 | 6.5E-02 | 6.5E-02 |
| Nitrobenzene | 1.1E-01 | 0.0 | 4.9E-01 | 9.3E-03 | 1.2E-01 | 1.2E-01 | 1.2E-01 | 6.1E-02 | 5.0E-01 | 0.0 | 0.0 |
| Hexane | 2.0E-03 | 0.0 | 2.4E-04 | 9.7E-02 | 9.7E-02 | 9.7E-02 | 9.7E-02 | 5.1E-02 | 4.2E-02 | 5.1E-03 | 5.1E-03 |
| Nitroethane | 3.3E-04 | 0.0 | 3.1E-02 | 7.7E-02 | 7.7E-02 | 7.7E-02 | 7.7E-02 | 5.6E-02 | 0.5E-02 | 1.7E-03 | 1.7E-03 |
| Vanadium | 0.0 | 0.0 | 0.0 | 1.9E-02 | 1.9E-02 | 1.9E-02 | 1.9E-02 | 1.1E-02 | 1.0E-02 | 5.3E-04 | 5.3E-04 |
| Acetic acid | 0.0 | 0.0 | 0.0 | 3.4E102 | 3.4E102 | 3.4E102 | 3.4E102 | 1.4E101 | 7.4E100 | 7.1E100 | 7.1E100 |
| Aniline | 0.0 | 0.0 | 0.0 | 2.4E101 | 2.4E101 | 2.4E101 | 2.4E101 | 1.2E-02 | 7.5E-03 | 4.4E-03 | 4.4E-03 |
| Butanoic Acid | 0.0 | 0.0 | 0.0 | 2.6E101 | 2.6E101 | 2.6E101 | 2.6E101 | 1.1E100 | 6.9E-01 | 4.1E-01 | 4.1E-01 |
| Catechol | 0.0 | 0.0 | 0.0 | 1.1E103 | 1.1E103 | 1.1E103 | 1.1E103 | 1.3E101 | 9.9E100 | 3.3E100 | 3.3E100 |
| Hexanoic Acid | 0.0 | 0.0 | 0.0 | 2.0E100 | 2.0E100 | 2.0E100 | 2.0E100 | 0.5E-02 | 5.3E-02 | 3.1E-02 | 3.1E-02 |
| Pentanoic Acid | 0.0 | 0.0 | 0.0 | 2.4E101 | 2.4E101 | 2.4E101 | 2.4E101 | 1.0E100 | 6.4E-01 | 3.0E-01 | 3.0E-01 |
| Phenol | 0.0 | 0.0 | 0.0 | 6.2E103 | 6.2E103 | 6.2E103 | 6.2E103 | 3.1E-01 | 2.1E-01 | 9.6E-02 | 9.6E-02 |
| Propionic Acid | 0.0 | 0.0 | 0.0 | 5.2E101 | 5.2E101 | 5.2E101 | 5.2E101 | 4.4E101 | 1.4E100 | 7.7E-01 | 7.7E-01 |
| Pyridine | 0.0 | 0.0 | 0.0 | 2.3E102 | 2.3E102 | 2.3E102 | 2.3E102 | 2.3E-02 | 1.7E-02 | 6.1E-03 | 6.1E-03 |
| Resorcinol | 0.0 | 0.0 | 0.0 | 5.5E102 | 5.5E102 | 5.5E102 | 5.5E102 | 4.4E100 | 3.3E100 | 1.1E100 | 1.1E100 |
| 2-hydroxybenzal | 0.0 | 0.0 | 0.0 | 4.0E102 | 4.0E102 | 4.0E102 | 4.0E102 | 1.0E-01 | 7.5E-02 | 2.6E-02 | 2.6E-02 |
| 2-hydroxypropionic Ac | 0.0 | 0.0 | 0.0 | 4.0E100 | 4.0E100 | 4.0E100 | 4.0E100 | 1.7E-01 | 1.1E-01 | 6.3E-02 | 6.3E-02 |
| 2-hydroxypropidine | 0.0 | 0.0 | 0.0 | 1.4E102 | 1.4E102 | 1.4E102 | 1.4E102 | 1.4E-01 | 1.0E-01 | 3.6E-02 | 3.6E-02 |
| 2,4-Dimethylpyridine | 0.0 | 0.0 | 0.0 | 2.6E100 | 2.6E100 | 2.6E100 | 2.6E100 | 2.0E-03 | 1.5E-03 | 5.2E-04 | 5.2E-04 |
| 2,4-Xylenol | 0.0 | 0.0 | 0.0 | 2.4E102 | 2.4E102 | 2.4E102 | 2.4E102 | 2.4E-01 | 2.0E-01 | 2.3E-02 | 2.3E-02 |
| 2,5-Dimethylpyridine | 0.0 | 0.0 | 0.0 | 2.0E100 | 2.0E100 | 2.0E100 | 2.0E100 | 2.0E-02 | 1.5E-03 | 5.2E-04 | 5.2E-04 |
| 3-Hydroxybutanoic Ac | 0.0 | 0.0 | 0.0 | 2.0E100 | 2.0E100 | 2.0E100 | 2.0E100 | 0.5E-02 | 5.3E-02 | 3.1E-02 | 3.1E-02 |
| 3-hydroxyethanol | 0.0 | 0.0 | 0.0 | 0.0E102 | 0.0E102 | 0.0E102 | 0.0E102 | 0.0 | 0.0 | 0.0 | 0.0 |
| 4-hydroxybenzoic | 0.0 | 0.0 | 0.0 | 0.4E101 | 0.4E101 | 0.4E101 | 0.4E101 | 1.3E-01 | 1.1E-01 | 1.4E-01 | 1.4E-01 |

TABLE 11 (CONT.)
FLOW RATES IN STREAMS (LBS/HR)

| COMPOUNDS/ TRACER ELEMENTS | 37 | 38 | 43 | 46 | 48 | 50 | 52 | 53 | 54 | 54H31 |
|-------------------------------|-----|-----|---------|---------|---------|---------|---------|---------|---------|---------|
| 3-methylbutyridine | 0.0 | 0.0 | 5.2E+01 | 5.2E+01 | 5.2E+01 | 5.2E+01 | 5.2E+02 | 3.0E-02 | 1.3E-02 | 1.3E-02 |
| 3,5-Xylenol | 0.0 | 0.0 | 1.0E+02 | 1.0E+02 | 5.0E-01 | 5.0E-01 | 3.1E-01 | 2.0E-01 | 3.1E-03 | 3.1E-02 |
| 3,5-Dimethylcatechol | 0.0 | 0.0 | 9.0E+01 | 9.0E+01 | 3.6E+01 | 3.6E+01 | 1.1E+00 | 0.1E-01 | 2.7E-01 | 2.7E-01 |
| 4-Methyl Resorcinol | 0.0 | 0.0 | 7.2E+01 | 7.2E+01 | 2.9E+01 | 2.9E+01 | 0.6E-01 | 6.5E-01 | 2.2E-01 | 2.2E-01 |
| 4-Methylcatechol | 0.0 | 0.0 | 7.7E+02 | 7.7E+02 | 3.1E+02 | 3.1E+02 | 9.2E+00 | 6.9E+00 | 3.3E+00 | 2.3E+00 |
| 4-ethylphenol | 0.0 | 0.0 | 6.0E+02 | 6.0E+02 | 3.0E+00 | 3.0E+00 | 9.0E-02 | 6.2E-02 | 2.2E-02 | 2.2E-02 |
| 4-Methylpyridine | 0.0 | 0.0 | 1.2E+01 | 1.2E+01 | 1.2E-01 | 1.2E-01 | 1.2E-02 | 0.0E-03 | 3.1E-03 | 3.1E-03 |
| 5-Methyl Resorcinol | 0.0 | 0.0 | 1.3E+02 | 1.3E+02 | 5.2E+01 | 5.2E+01 | 1.6E+00 | 1.2E+00 | 3.9E-01 | 3.9E-01 |
| Acenaphthylene | 0.0 | 0.0 | 0.0E-01 | 0.0E-01 | 6.0E-01 | 6.0E-01 | 6.0E-02 | 6.7E-02 | 6.0E-04 | 6.0E-04 |
| Anthracene | 0.0 | 0.0 | 2.0E-01 | 2.0E-01 | 1.7E-01 | 1.7E-01 | 1.7E-02 | 1.7E-02 | 1.7E-04 | 1.7E-04 |
| Benzo(a)anthracene | 0.0 | 0.0 | 2.0E-02 | 2.0E-02 | 1.7E-02 | 1.7E-02 | 1.7E-03 | 1.7E-03 | 1.7E-05 | 1.7E-05 |
| Benzo(b)fluorene | 0.0 | 0.0 | 0.0E-03 | 0.0E-03 | 6.0E-03 | 6.0E-03 | 6.0E-04 | 6.7E-04 | 6.0E-06 | 6.0E-06 |
| Benzo(e)pyrene | 0.0 | 0.0 | 0.0E-03 | 0.0E-03 | 6.0E-03 | 6.0E-03 | 6.0E-04 | 6.7E-04 | 6.0E-06 | 6.0E-06 |
| Benzo(k)fluorene | 0.0 | 0.0 | 6.0E-04 | 6.0E-04 | 5.1E-04 | 5.1E-04 | 5.1E-05 | 5.0E-05 | 5.1E-07 | 5.1E-07 |
| Biphenyl | 0.0 | 0.0 | 1.4E+00 | 1.4E+00 | 1.2E+00 | 1.2E+00 | 3.4E-01 | 3.5E-01 | 3.6E-03 | 3.6E-03 |
| Chrysene | 0.0 | 0.0 | 4.0E-03 | 4.0E-03 | 3.4E-03 | 3.4E-03 | 3.4E-04 | 3.4E-04 | 3.4E-06 | 3.4E-06 |
| Fluorene | 0.0 | 0.0 | 2.0E-01 | 2.0E-01 | 1.7E-01 | 1.7E-01 | 1.7E-02 | 1.3E-02 | 4.4E-03 | 4.4E-03 |
| Indeno(1,2,3-cd)perylene | 0.0 | 0.0 | 3.0E+01 | 3.0E+01 | 2.6E+01 | 2.6E+01 | 2.6E+00 | 1.6E+00 | 9.5E-01 | 9.5E-01 |
| Indene | 0.0 | 0.0 | 4.0E-01 | 4.0E-01 | 3.4E-01 | 3.4E-01 | 3.4E-02 | 3.4E-02 | 3.4E-04 | 3.4E-04 |
| Fluoranthene | 0.0 | 0.0 | 4.0E-01 | 4.0E-01 | 3.4E-01 | 3.4E-01 | 3.4E-02 | 3.4E-02 | 3.4E-04 | 3.4E-04 |
| Fluorene | 0.0 | 0.0 | 4.0E-01 | 4.0E-01 | 3.4E-01 | 3.4E-01 | 3.4E-02 | 3.4E-02 | 3.4E-04 | 3.4E-04 |
| Indole | 0.0 | 0.0 | 1.0E+01 | 1.0E+01 | 1.5E+01 | 1.5E+01 | 4.5E+00 | 4.5E+00 | 4.5E-02 | 4.5E-02 |
| Anthracene | 0.0 | 0.0 | 4.4E+01 | 4.4E+01 | 3.7E+01 | 3.7E+01 | 3.7E+00 | 3.7E+00 | 3.7E-01 | 3.7E-01 |
| Benzo(a)pyrene | 0.0 | 0.0 | 6.0E+00 | 6.0E+00 | 5.4E+00 | 5.4E+00 | 1.6E+00 | 1.6E+00 | 1.6E-02 | 1.6E-02 |
| Perylene | 0.0 | 0.0 | 0.0E-04 | 0.0E-04 | 6.0E-04 | 6.0E-04 | 6.0E-05 | 6.7E-05 | 6.0E-07 | 6.0E-07 |
| Phenanthrene | 0.0 | 0.0 | 2.0E-01 | 2.0E-01 | 1.7E-01 | 1.7E-01 | 1.7E-02 | 1.7E-02 | 1.7E-04 | 1.7E-04 |
| Pyrene | 0.0 | 0.0 | 4.0E-01 | 4.0E-01 | 3.4E-01 | 3.4E-01 | 3.4E-02 | 3.4E-02 | 3.4E-04 | 3.4E-04 |
| Quinoline | 0.0 | 0.0 | 1.9E+01 | 1.9E+01 | 1.9E-01 | 1.9E-01 | 1.9E-02 | 1.9E-02 | 4.9E-03 | 4.9E-03 |
| Toluene | 0.0 | 0.0 | 1.0E+02 | 1.0E+02 | 0.3E+04 | 0.3E+04 | 0.5E+00 | 0.5E+00 | 3.7E+00 | 3.7E+00 |
| o-Xylene | 0.0 | 0.0 | 3.3E+01 | 3.3E+01 | 3.0E+01 | 3.0E+01 | 3.0E+00 | 1.9E+00 | 1.1E+00 | 1.1E+00 |

2.3 Characterization of Gaseous Streams

Characterization data in the literature were very limited. No experimental data could be found regarding specific gaseous constituents from Lurgi/Fischer-Tropsch processing of Wyoming subbituminous coal.

Gaseous streams selected for the analysis are listed in Table 12; their interrelationships are illustrated in Figure 3. These streams were selected because they may contain constituents of environmental concern and they will be released to the environment.

The analysis focused on criteria pollutants and two general classes of pollutants which have been identified as potential causes of concern in gaseous streams from coal conversion facilities: trace elements and organic compounds.

TABLE 12

GASEOUS STREAMS CHARACTERIZED FOR RISK ASSESSMENT

| <u>STREAM NUMBER*</u> | <u>STREAM NAME</u> |
|-----------------------|--|
| 28 | Utility stack gases to atmosphere |
| 29 | Evaporative losses (incl. cooling tower drift) |
| 72 | Lockhopper vent gas emissions |
| 73 | Evaporative losses (Streams 18-23) |
| 74 | Evaporative losses (Stream 60) |
| 75 | Evaporative losses (Stream 45) |
| 76 | Evaporative losses (Stream 44) |
| 77 | Evaporative losses (Stream 47) |
| 78 | Evaporative losses (Stream 49) |

*See Figure 3 on following page.

AIR SEPARATION

BAG HOUSE

COAL PREPARATION

UNDERSIZED COAL

COOLING TOWER DRIFT

EVAPORATION

3

STACK GAS ATMOSPHERE

28

UTILITIES GENERATION

29

MAKE-UP WATER*

31

EMISSIONS FROM COAL LOCKHOPPER

72

COAL LOCKHOPPER

PRESSUR GAS

LURGI PRESSURE GASIFICATION

WASH COOLER

FISCHER-TROPSCH PRODUCT UPGRADING

PRODUCT EVAPORATIVE EMISSIONS

73

18

19

C3 LPG

C4 LPG

20

21

GASOLINE

22

23

DIST. FUEL OIL

HEAVY FUEL

DUST REMOVAL

ALCOHOL EVAPORATIVE EMISSIONS

74

ASH LOCKHOPPER

ASH HANDLING

CLEAN WATER

FIGURE 3

PROCESS STREAMS FOR LIQUEFACTION PROCESS (FISCHER-TROPSCH)

FOR CHARACTERIZING EAMS.

In general the types and quantities of pollutants in each stream reported here have been derived from the literature and modified, as necessary, to reflect the size and feed coal of the conceptual plant used in this analysis. The types and quantities of trace elements present in the gaseous streams were estimated using the trace element characterization data for Wyoming subbituminous coal presented in Table 13. The distribution of trace elements from the gasifier was based on SASOL data³⁷, and is presented in Table 14. The distribution of trace elements from the utility boiler was based on data from the WESCO Coal Gasification Project¹⁰⁸ and is presented in Table 15.

The specific calculations, data and assumptions used in estimating the types and concentrations of trace elements and organic compounds which may be present in the gaseous streams were analyzed by stream.

Utility Stack Gases (Stream 28)

Stream Constituents:

The major and minor constituents of Stream 28 were reported by Schreiner⁸⁷, and are presented in Table 16.

Any of the trace elements found in the coal could be present in trace amounts. The flow rates of five trace elements in Stream 28 were estimated by calculating the amount of each element entering the utility boiler in the feed coal and tar from the gasifier, and then applying distribution coefficients based on the WESCO report (See Table 15).

The quantity of each trace element in the tar was estimated:

$$Q_{TE-T} = (C_{TE-DC})(F_{DC-G})(D_{TE-T})$$

where-

Q_{TE-T} is the flow rate of the trace element in the tar feed to the boiler (lb/hr).

C_{TE-DC} is the concentration of the trace element in the dry coal, ppm.

TABLE 13

TRACE ELEMENT CONCENTRATIONS FOR WYOMING AND ROSEBUD MONTANA
SUBBITUMINOUS COALS (ppm)

| Trace Elements | Wyoming Subbituminous ⁴⁰ | Montana Rosebud Subbituminous ⁴⁰ | Ratio of Concentrations of Trace Elements - Montana to Wyoming |
|----------------|-------------------------------------|---|--|
| Ag | .06-.43 | .06 | 1-.14 |
| As | .57-1.2 | .08-1.2 | .14-1 |
| B | 32 | 32 | 1 |
| Ba | 87 | 87 | 1 |
| Be | .71-.8 | .7-.8 | .99-1 |
| Br | - | - | - |
| Cd | .31-.8 | .31-.8 | 1-1 |
| Ce | - | - | - |
| Co | .55 | .6-4 | 1.09-7.27 |
| Cr | 4.2-16 | 4-16 | .95-1 |
| Cs | - | - | - |
| Cu | 8.9-10 | 9-10 | 1.01-1 |
| F | 65-67 | 66 | .98-1.01 |
| Ga | - | - | - |
| Ge | - | - | - |
| Hg | .11-.17 | .11-.17 | 1-1 |
| I | - | - | - |
| In | - | - | - |
| La | - | - | - |
| Li | 3.6-15.0 | - | - |
| Mo | 2.2 | 2.2 | 1 |
| Mn | 2.8-3.4 | 2.8-3.4 | 1-1 |
| Ni | 1.7-14 | 2-14 | 1.18-1 |
| P | - | - | - |
| Pb | .51-12 | .51-12 | 1-1 |
| Rb | - | - | - |
| Ru | - | - | - |
| Sb | .08-1.5 | - | - |
| Sc | - | - | - |
| Se | .33 | .33 | 1 |
| Sn | .14 | .14 | 1 |
| Sr | - | - | - |
| Ta | - | - | - |
| Te | - | - | - |
| U | .88 | .88 | 1 |
| V | 10-14 | 10-14 | 1 |
| W | - | - | - |
| Y | - | - | - |
| Zn | .23-8 | 2-8 | 8.70-1 |
| Zr | - | 170 | - |

TABLE 14
TRACE ELEMENT DISTRIBUTION FOR LURGI AT SASOL*, 37
(Percent of Element in Coal)

| Element | Ash | Liquor | Tar | Oil |
|---------|--------|--------|------|-----|
| Be | 33.3 | 53.3 | 17.0 | 0.3 |
| B | 90 | 8.8 | 2.0 | 0.0 |
| V | 99.9 | 0.1 | 0.0 | 0.0 |
| Mn | 99.9 | 0.2 | 0.0 | 0.0 |
| Ni | 99.4 | 0.4 | 0.0 | 0.0 |
| As | 26.9 | 67.2 | 1.9 | 3.9 |
| Cd | 51.9 | 45.5 | 0.6 | 1.4 |
| Sb | 50.0 | 45.0 | 3.8 | 0.6 |
| Ce | 99.9 | 0.1 | 0.0 | 0.0 |
| Hg | 51.9 | 41.6 | 6.4 | 0.6 |
| Pb | 94.2 | 1.7 | 4.3 | 0.0 |
| Br | 10.0 | 88.9 | 0.1 | 0.0 |
| F | 56.3** | 43.8** | 0.0 | 0.0 |
| Cl | 52.6** | 47.4** | 0.3 | 0.0 |

* Analysis by spark source mass spectrometer (which can give a semi-quantitative analysis) for El Paso by SASOL.

** % distribution calculated on analyses as done by SASOL previously.

TABLE 15

DISTRIBUTION OF TRACE ELEMENTS FROM
UTILITY BOILER ¹⁰⁸

| TRACE ELEMENT | DISTRIBUTION (%) | | | |
|------------------|------------------|----------------|-----------------------|--------------------------|
| | BOTTOM ASH | ESP FLY ASH | STACK GAS SCRUBBER | ATMOSPHERIC EMISSIONS |
| As | 4.4 | 94.6 | 0.8 | 0.2 |
| Be | 16.9 | 82.2 | 0.7 | 0.2 |
| Cd | 16.0 | 82.7 | 1.0 | 0.3 |
| F | 1.2 | 26.8 | 57.6 | 14.4 |
| Hg | 4.4 | 13.0 | 0.1 | 82.5 |
| Pb | 9.7 | 89.3 | 0.8 | 0.2 |

F_{DC-G} is the flow rate of dry coal to the gasifier, i.e., 1,369,000 lb/hr.

D_{TE-T} is the fraction of the total quantity of this trace element fed to the gasifier that ends up in the tar. (Table 14).

and the total quantity of each trace element in Stream 28 due to trace element input from the feed coal and tar was estimated:

$$Q_{TE-28} = [(C_{TE-DC})(Q_{DC-B}) + (Q_{TE-T})] D_{TE-28}$$

where-

Q_{TE-28} is the flow rate of the trace element to the atmosphere, lb/hr

C_{TE-DC} is the concentration of the trace element in the coal feed to the boiler, ppm, dry coal basis.

Q_{DC-B} is the flow rate of dry coal to the boiler, i.e., 298,000 lb/hr.

Q_{TE-T} is the flow rate of the trace element in the tar feed to the boiler, lb/hr. (from Table 17, below)

D_{TE-28} is the fraction of the total quantity of this trace element fed to the boiler that is emitted to the atmosphere. (Table 15)

TABLE 16 - Utility Stack Gases, Stream 28

| Component | lbs/hr* | Wt% |
|---------------------------------------|-----------|---------|
| CO ₂ | 2,739,799 | 39.46 |
| N ₂ + Inerts | 3,632,785 | 52.33 |
| O ₂ | 71,264 | 1.03 |
| H ₂ O | 497,676 | 7.17 |
| SO _x (as SO ₂) | 960 | 138 ppm |
| NO _x (as NO) | 176 | 25 ppm |

*Appendix B, Stream 28

The calculations are presented in Tables 17 and 18.

TABLE 17 - Flow Rates of Trace Elements in the Tar

$$C_{TE-DC}^{40} \times F_{DC-G}^{87} \times D_{TE-T}^{*,37} = Q_{TE-T}$$

| Trace Elements | (ppm) | (MM lb/hr) | | (lb/hr) |
|----------------|----------|------------|------|-------------|
| As | .57-1.2 | 1.369 | .019 | .0148-.0312 |
| Be | .71-.80 | 1.369 | .17 | .1652-.1862 |
| Cd | .31-.80 | 1.369 | .006 | .0025-.0066 |
| F | 65-67 | 1.369 | 0.0 | 0.0 |
| Hg | .11-.17 | 1.369 | .064 | .0096-.0149 |
| Pb | .51-12.0 | 1.369 | .043 | .0300-.7064 |

TABLE 18 - Flow Rates of Trace Elements in Stream 28

$$(C_{TE-DC}^{40} \times F_{DC-B}^{87} + Q_{TE-T}^{**}) \times D_{TE-28}^{108} = Q_{TE-28}$$

| Trace Elements | (ppm) | (MM lb/hr) | (lb/hr) | | (lb/hr) |
|----------------|----------|------------|-------------|------|--------------|
| As | .57-1.2 | .298 | .0148-.0312 | .002 | .0004-.0008 |
| Be | .71-.80 | .298 | .1652-.1862 | .002 | .0008-.00085 |
| Cd | .31-.80 | .298 | .0025-.0066 | .003 | .0003-.0007 |
| F | 65-67 | .298 | .0000-.0000 | .144 | 2.789-2.875 |
| Hg | .11-.17 | .298 | .0096-.0149 | .825 | .0350-.0541 |
| Pb | .51-12.0 | .298 | .0300-.7064 | .002 | .0004-.0086 |

* See Table 14

** From Table 17

Atmospheric Losses from Cooling Towers* (Stream 29)

Stream Constituents:

The flow rate of cooling tower atmospheric losses for the plant size studied was 2,413,000 lbs/hr. The flow rates of the constituents in the cooling tower atmospheric losses were calculated using the following equation:

$$Q_C = (C)(F_A)$$

where-

Q_C is the flow rate of the component, lbs/hr.

C is the concentration of this component in the cooling water, ppm.

F_A is the flow rate of the atmospheric losses, MM lbs/hr.

The calculations are presented in Table 19.

Feed Lockhopper Vent Gas (Stream 72)

Stream Constituents:

The following assumptions were made in calculating the composition of the feed lockhopper vent gas.

- 1) The feed lockhopper pressurizing gas has the same composition as the product gas stream leaving the Gas/Liquor Separator. Gas phase composition of the raw gas was assumed to be those components that remained in the product gas stream after condensation and separation of the Lurgi tar and Lurgi oil. Sound engineering practice supports this interpretation.
- 2) 0.1% of the total flow rate of product gas exiting the Gas/Liquor Separator was vented to the atmosphere during lockhopper depressurization.⁴⁰
- 3) The list of individual organic compounds, except for the fatty acids, contained in the vent gas was derived from the literature.¹⁹

*Atmospheric losses from the cooling tower consist of two components, evaporation and drift. Make-up water to the cooling tower consists of boiler blowdown and treated gas liquor.

TABLE 19 - Atmospheric Losses from Cooling Towers, Stream 29

| Component | C^* (ppm) | x | F_A^{87} (MM lb/hr) | $=$ | Q_C (lbs/hr) |
|-------------------------|----------------|-----|--------------------------|-----|-------------------|
| COD | 501 | | 2.413 | | 1210 |
| BOD | 77 | | 2.413 | | 190 |
| TOC** | 89 | | 2.413 | | 220 |
| Tars/Oils** | 9 | | 2.413 | | 20 |
| Other Organics** | 85 | | 2.413 | | 210 |
| Ammonia | 508 | | 2.413 | | 1230 |
| Na ⁺¹ | 363 | | 2.413 | | 876 |
| Ca ⁺¹ | 209 | | 2.413 | | 504 |
| Mg ⁺² | 0 | | 2.413 | | 0 |
| Alkalinity as C_aCO_3 | 0 | | 2.413 | | 0 |
| SO_3^{-2} / SO_4^{-2} | 2254 | | 2.413 | | 5439 |
| Cl ⁻ | 112 | | 2.413 | | 270 |
| SCN ⁻ | 1 | | 2.413 | | 2 |
| TDS | 3083 | | 2.413 | | 7439 |

4) It was assumed that the fatty acids present in the raw gas liquor would also be present in the product gas. The list of fatty acids contained in the gas liquor was presented in the literature.⁹²

5) It was assumed that the trace elements present in the coal would also be present in the product gas. The list of trace elements contained in Wyoming subbituminous coal was derived from the literature⁴⁰ and is presented in Table 13 of this report.

The estimated composition of Stream 72 is presented in Table 20.

*Reference: Unpublished EPA data

** See composition of liquid stream 54 + 31 or organic compounds and trace elements possibly present in the atmospheric losses

TABLE 20--Feed Lockhopper Vent Gas, Stream 72

| <u>Component</u> | <u>Lbs/Hr</u> * | <u>Wt %</u> ** |
|-------------------------------|-----------------|----------------|
| N ₂ | 8.4 | .002 |
| H ₂ O | 1422 | 36.374 |
| CO ₂ | 1483 | 37.934 |
| CO | 602 | 15.399 |
| H ₂ | 89 | 2.277 |
| CH ₄ | 206 | 5.269 |
| C ₂ H ₄ | 2.5 | 0.064 |
| H ₂ S | 5.4 | 0.138 |
| COS | .12 | 0.003 |
| Ni(CO) ₄ | .006 | 1.5 ppm |
| Tars: | 4.4 | .113 |
| Naphthalene | | |
| Anthracene | | |
| Phenanthrene | | |
| Pyrene | | |
| Cresols | | |
| Phenol | | |
| Xylenols | | |
| Fluorene | | |
| Dibenzofuran | | |
| Chrysene | | |
| Perylene | | |
| Aniline | | |

*0.1% of Stream 10 in the flow diagram.

**Calculated using the following equation:

$$\text{wt \%} = \frac{(\text{lbs/hr of the component})}{3909.38} \times 100$$

TABLE 20 (Cont.)

| <u>Component</u> | <u>Lbs/Hr</u> | <u>Wt%</u> |
|---|---------------|------------|
| Oils: | 38.5 | .985 |
| Benzene | | |
| Thiophene | | |
| Toluene | | |
| Xylenes | | |
| Phenol | | |
| Alkylphenols | | |
| Naphthalene | | |
| Indan | | |
| Cresols | | |
| Xylenols | | |
| Naphtha: | 15.7 | .402 |
| C ₂ -C ₆ Aliphatics | | |
| Benzene | | |
| Xylene | | |
| Ethylbenzene | | |
| Phenols: | 5.2 | .133 |
| Phenol | | |
| Xylenols | | |
| Cresols | | |
| Trimethylphenol | | |
| O-Isopropylphenol | | |
| Mercaptans: | .02 | 5.1 ppm |
| Methanethiol | | |
| Ethanethiol | | |
| Thiophenes: | .016 | 4.1 ppm |
| Thiophene | | |
| Methylthiophene | | |
| Ammonia | 8.1 | .207 |
| HCN | .008 | 2.0 ppm |
| Aromatic Amines: | .004 | 1.0 ppm |
| Aniline | | |
| Methylaniline | | |
| Dimethylaniline | | |

TABLE 20 (Concluded)

| <u>Component</u> | <u>Lbs/Hr</u> | <u>Wt %</u> |
|------------------------|----------------|--------------|
| Nitrosamines: | .002 | .5 ppm |
| N-Nitrosamine | | |
| PAHs | .002 | .5 ppm |
| Chrysene | | |
| Perylene | | |
| Pyrene | | |
| Fluorene | | |
| Anthracene | | |
| Naphthalene | | |
| Biphenyl | | |
| Indene | | |
| Benzofuran | | |
| Dibenzofuran | | |
| Fluoranthene | | |
| Quinoline | | |
| Acridine | | |
| Fatty Acids: | 1 | .026 |
| Acetic Acid | | |
| Propanoic Acid | | |
| n-Butanoic Acid | | |
| 2-Methylpropanoic Acid | | |
| n-Pentanoic Acid | | |
| 3-Methyl Butanoic Acid | | |
| n-Hexanoic Acid | | |
| Trace Elements: | .002 | .5 ppm |
| Ag | | |
| As | | |
| B | | |
| Ba | | |
| Be | | |
| Cd | | |
| Co | | |
| Cr | | |
| Cu | | |
| F | | |
| Hg | | |
| Mn | | |
| Ni | | |
| Pb | | |
| Se | | |
| Sn | | |
| U | | |
| V | | |
| Zn | | |
| Total | <u>3909.38</u> | <u>99.79</u> |

Evaporative Emissions (Stream 73-78)

Stream Constituents:

The contaminants believed to be found in evaporative emissions are presented in Table 21. A more detailed breakdown of gasoline in Stream 73 is presented in Table 22.

TABLE 21 - Evaporative Emissions for
Product and By-Product Storage, Streams 73-78

| <u>Product/By-Product</u> | <u>Lbs/Hour</u> |
|--------------------------------|-----------------|
| Diesel Oil (Stream 73) | 0.513 |
| Fuel Oil (Stream 73)* | 0.014 |
| Lurgi Oil (Stream 75)** | 12.625 |
| Lurgi Tar (Stream 76)† | 6.750 |
| Alcohol (Stream 74)†† | 1.500 |
| Phenols (Stream 77)§ | 3.750 |
| Ammonia (Stream 78) | No Data |
| C ₃ LPG (Stream 73) | No Data |
| C ₄ LPG (Stream 73) | No Data |

*The fuel oil was reported to be all C₇ + hydrocarbons.⁸⁷

**The following compounds were reported to be present in the Lurgi oil: benzene, thiophene, toluene, xylenes, phenol, alkyl phenols, naphthalene, cresols, benzothiophene. (EPA unpublished data.)

†The following compounds were reported to be present in the Lurgi tar: naphthalene, anthracene, fluoranthene, pyrene, cresols, phenol, fluorene, dibenzofuran. (EPA unpublished data.)

††The alcohol was reported to be ethanol and C₃ + alcohols.⁸⁷

§ The following phenolic compounds were expected to be present:
phenol: catechol, resorcinol, methylcatechol, methyl resorcinol,
cresols, xylenols.

TABLE 22-Evaporative Emissions of Gasoline, Stream 73

| <u>Component</u> | <u>lbs/Hour</u> * |
|------------------|-------------------|
| Methane | 0.213 |
| Ethylene | 0.001 |
| Propylene | 0.017 |
| Propane | 0.425 |
| Isobutane | 2.071 |
| Isobutylene | 0.738 |
| n-Butane | 2.413 |
| Cis-2-Butane | 0.575 |
| Isopentane | 5.238 |
| n-Pentane | 2.450 |
| Hexanes | 0.446 |
| Heptanes | 0.188 |
| Octanes | 0.299 |
| Total | 15.055 |

2.4 Characterization of Solid Streams

No experimental data could be found regarding specific constituents in solid wastes from Lurgi/Fischer-Tropsch processing of Wyoming subbituminous coal.

Solid streams selected for the analysis are listed in Table 23; their interrelationships are illustrated in Figure 4. These streams were selected because they may contain constituents of environmental concern, and their disposal may result in environmental releases.

The analysis focused on two classes of pollutants which have been identified as potential causes of environmental concern in solid streams from coal conversion facilities; trace elements and organic compounds.

In general the types and quantities of pollutants in each stream either have been

- estimated from the literature and modified, as necessary, to reflect the size and feed coal of the conceptual plant used by ORNL as the basis of this analysis, or

*Unpublished EPA data

TABLE 23
SOLID STREAMS CHARACTERIZED FOR RISK ASSESSMENT

| STREAM NUMBER | STREAM NAME |
|---------------|---|
| 33 | Utility ash and SO ₂ scrubber sludge |
| 34 | Gasifier ash |
| 36 | Combined ash |
| 62 | Fresh Fischer-Tropsch catalyst |
| 63 | Spent Fischer-Tropsch catalyst |
| 69* | Leachate from ash (Stream 36) |
| 70* | Biological sludge |
| 71* | Leachate from biosludge (Stream 70) |
| 79 | Spent shift catalyst |

*Liquid stream, but emanating from a solid waste after disposal.

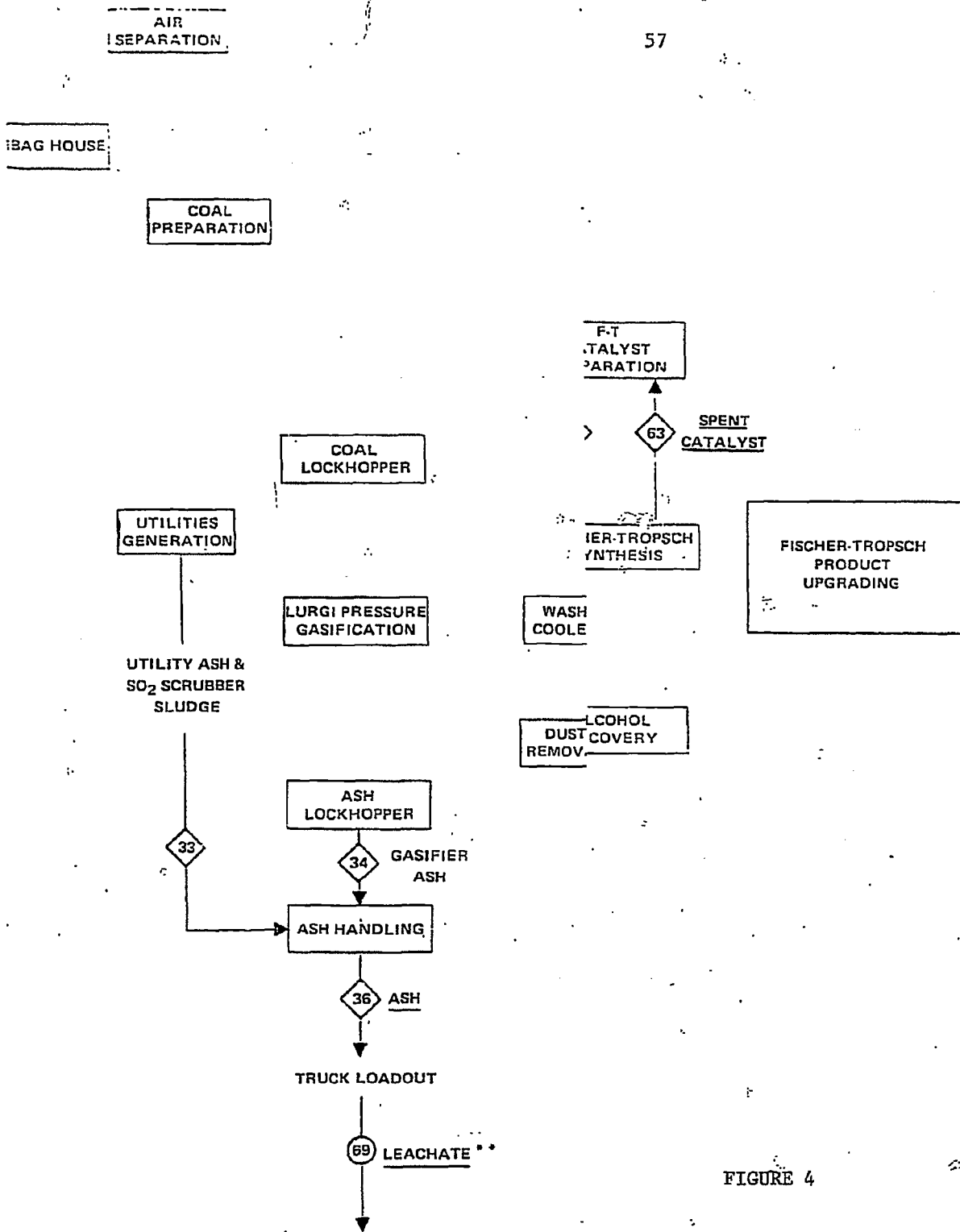


FIGURE 4

LID STREAMS FOR LIQUEFACTION PROCESS (FISCHER-TROPSCH)

FOR CHARACTERIZING

NUMBERS WHICH ARE NOT IN DIAGRAM.

- estimated by assuming that pollutants removed from the gaseous and liquid streams in the solid form via non-destructive mechanisms (e.g., absorption and reverse osmosis) will be present in the resultant solid wastes.

The type and quantities of trace elements present in Stream 33 (bottom ash, fly ash and scrubber sludge) were estimated using the trace element characterization data for Wyoming subbituminous coal presented in Table 24. The distribution of trace elements from the utility boiler and in the gasifier tar were calculated using distribution data from WESCO¹⁰⁸ and SASOL³⁷, respectively, presented in Tables 25 and 26.

The specific calculations, data and assumptions used in estimating the types and concentrations of trace elements and organic compounds which may be present in the solid streams are discussed in the following pages.

Boiler Bottom Ash, Fly Ash and Scrubber Sludge (Stream 33)

Stream Constituents:

The concentrations of six trace elements in Stream 33 were estimated by calculating the total input of each trace element into the utilities generation section (in coal and gasifier tar) and then applying the distribution factor for each trace element (See Table 25).

The flow rates of the trace elements in the tar feed to the boiler were first calculated using the following equation.

$$Q_{TE-T} = (C_{TE-DC})(F_{DC-G})(D_{TE-T})$$

where-

Q_{TE-T} is the flow rate of the trace element in the tar feed to the boiler, lb/hr

C_{TE-DC} is the concentration of the trace element in the dry coal, in ppm

F_{DC-G} is the flow rate of dry coal to the gasifier, MM lbs/hr

D_{TE-T} is the fraction of the total quantity of this trace element fed to the gasifier that ends up in the tar.

TABLE 24

TRACE ELEMENT CONCENTRATIONS FOR WYOMING AND ROSEBUD MONTANA
SUBBITUMINOUS COALS (ppm)

| Trace Elements | Wyoming Subbituminous ⁴⁰ | Montana Rosebud Subbituminous ⁴⁰ | Ratio of Concentrations of Trace Elements - Montana to Wyoming |
|----------------|-------------------------------------|---|--|
| Ag | .06-.43 | .06 | 1-.14 |
| As | .57-1.2 | .08-1.2 | .14-1 |
| B | 32 | 32 | 1 |
| Ba | 87 | 87 | 1 |
| Be | .71-.8 | .7-.8 | .99-1 |
| Br | - | - | - |
| Cd | .31-.8 | .31-.8 | 1-1 |
| Ce | - | - | - |
| Co | .55 | .6-4 | 1.09-7.27 |
| Cr | 4.2-16 | 4-16 | .95-1 |
| Cs | - | - | - |
| Cu | 8.9-10 | 9-10 | 1.01-1 |
| F | 65-67 | 66 | .98-1.01 |
| Ga | - | - | - |
| Ge | - | - | - |
| Hg | .11-.17 | .11-.17 | 1-1 |
| I | - | - | - |
| In | - | - | - |
| La | - | - | - |
| Li | 3.6-15.0 | - | - |
| Mo | 2.2 | 2.2 | 1 |
| Mn | 2.8-3.4 | 2.8-3.4 | 1-1 |
| Ni | 1.7-14 | 2-14 | 1.18-1 |
| P | - | - | - |
| Pb | .51-12 | .51-12 | 1-1 |
| Rb | - | - | - |
| Ru | - | - | - |
| Sb | .08-1.5 | - | - |
| Sc | - | - | - |
| Se | .33 | .33 | 1 |
| Sn | .14 | .14 | 1 |
| Sr | - | - | - |
| Ta | - | - | - |
| Te | - | - | - |
| U | .88 | .88 | 1 |
| V | 10-14 | 10-14 | 1 |
| W | - | - | - |
| Y | - | - | - |
| Zn | .23-8 | 2-8 | 8.70-1 |
| Zr | - | 170 | - |

TABLE 25
DISTRIBUTION OF TRACE ELEMENTS FROM
UTILITY BOILER

| TRACE ELEMENT | DISTRIBUTION (%) | | | |
|------------------|------------------|----------------|-----------------------|--------------------------|
| | BOTTOM ASH | ESP FLY ASH | STACK GAS SCRUBBER | ATMOSPHERIC EMISSIONS |
| As | 4.4 | 94.6 | 0.8 | 0.2 |
| Be | 16.9 | 82.2 | 0.7 | 0.2 |
| Cd | 16.0 | 82.7 | 1.0 | 0.3 |
| F | 1.2 | 26.8 | 57.6 | 14.4 |
| Hg | 4.4 | 13.0 | 0.1 | 82.5 |
| Pb | 9.7 | 89.3 | 0.8 | 0.2 |

TABLE 26
TRACE ELEMENT DISTRIBUTION FOR LURGI AT SASOL^{*,37}
(Percent of Element in Coal)

| Element | Ash | Liquor | Tar | Oil |
|---------|--------|--------|------|-----|
| Be | 33.3 | 53.3 " | 17.0 | 0.3 |
| B | 90 | 8.8 | 2.0 | 0.0 |
| V | 99.9 | 0.1 | 0.0 | 0.0 |
| Mn | 99.9 | 0.2 | 0.0 | 0.0 |
| Ni | 99.4 | 0.4 | 0.0 | 0.0 |
| As | 26.9 | 67.2 | 1.9 | 3.9 |
| Cd | 51.9 | 45.5 | 0.6 | 1.4 |
| Sb | 50.0 | 45.0 | 3.8 | 0.6 |
| Ce | 99.9 | 0.1 | 0.0 | 0.0 |
| Hg | 51.9 | 41.6 | 6.4 | 0.6 |
| Pb | 94.2 | 1.7 | 4.3 | 0.0 |
| Br | 10.0 | 88.9 | 0.1 | 0.0 |
| F | 56.3** | 43.8** | 0.0 | 0.0 |
| Cl | 52.6** | 47.4** | 0.3 | 0.0 |

* Analysis by spark source mass spectrometer (which can give a semi-quantitative analysis) for El Paso by SASOL.

** % distribution calculated on analyses as done by SASOL previously.

The flow rates of the trace elements in the bottom ash plus fly ash stream were then calculated using the equation below.

$$Q_{TE-A} = [(C_{TE-DC})(F_{DC-B}) + (Q_{TE-T})] D_{TE-A}$$

where-

- Q_{TE-A} is the flow rate of the trace element in the bottom ash plus fly ash, lb/hr.
- C_{TE-DC} is the concentration of the trace element in the coal feed to the boiler, ppm dry coal basis.
- F_{DC-B} is the flow rate of dry coal to the boiler, MM lbs/hr.
- Q_{TE-T} is the flow rate of the trace element in the tar feed to the boiler, lb/hr.
- D_{TE-A} is the fraction of the total quantity of this trace element fed to the boiler that is removed with the boiler bottom ash and fly ash.

The calculations are presented in Tables 27, 28, and 29.

Table 27 - Flow Rates of Trace Elements in the Tar

| <u>Trace Element</u> | C_{TE-DC}^{40} (ppm) | x | F_{DC-G}^* (MM lb/hr) | x | D_{TE-T}^{37} | = | Q_{TE-T} (lb/hr) |
|----------------------|---------------------------|---|----------------------------|---|-----------------|---|-----------------------|
| As | .57-1.2 | | 1.369 | | .019 | | .0148-.0312 |
| Be | .71-.80 | | 1.369 | | .17 | | .1652-.1862 |
| Cd | .31-.80 | | 1.369 | | .006 | | .0025-.0066 |
| F | 65-67 | | 1.369 | | 0.0 | | 0 |
| Hg | .11-.17 | | 1.369 | | .064 | | .0096-.0149 |
| Pb | .51-12.0 | | 1.369 | | .043 | | .0300-.7064 |

*Appendix B, Stream 2

TABLE 28 - Flow Rates of Trace Elements in the Bottom Ash and Fly Ash of Stream 33

| Trace Element | $(C_{TE-DC}^{40}) \times$ (ppm) | F_{DC-B}^* (MM lb/hr) | $+ F_{TE-T}^{**}$ (lb/hr) | $\times D_{TE-A}^{37}$ | $= Q_{TE-A}$ (lb/hr) |
|---------------|------------------------------------|----------------------------|------------------------------|------------------------|-------------------------|
| As | .57-1.2 | .298 | .0148-.0312 | .99 | .183-.385 |
| Be | .71-.80 | .298 | .1652-.1862 | .991 | .373-.421 |
| Cd | .31-.80 | .298 | .0025-.0066 | .987 | .094-.242 |
| F | 65-67 | .298 | 0 | .28 | 5.424-5.590 |
| Hg | .11-.17 | .298 | .0096-.0149 | .174 | .0074-.0114 |
| Pb | .51-12.0 | .298 | .0300-.7064 | .99 | .180-4.240 |

The flow rates of six trace elements in the scrubber sludge portion of Stream 33 were estimated using the following equation.

$$Q_{TE-S} = [(C_{TE-DC})(F_{DC-B}) + (F_{TE-T})] D_{TE-S}$$

where-

- Q_{TE-S} is the flow rate of the trace element in the scrubber sludge, lb/hr.
- C_{TE-DC} is the concentration of the trace element in the coal feed to the boiler, ppm, dry coal basis.
- F_{DC-B} is the flow rate of dry coal to the boiler, MM lbs/hr.
- F_{TE-T} is the flow rate of the trace element in the tar feed to the boiler, lb/hr.
- D_{TE-S} is the fraction of the total quantity of this trace element fed to boiler that was removed with the scrubber sludge.

The calculations are presented in Table 29.

*Appendix B, Stream 3

**From Table 27

**TABLE 29 - Flow Rates of Trace Elements in the Scrubber
Sludge Portion of Stream 33**

| Trace Element | (C_{TE-DC}^{40}) (ppm) | $\times F_{DC-B}^*$ (MM lb/hr) | $+ F_{TE-T}^{**}$ (lb/hr) | $\times D_{TE-S}^{108}$ | $= Q_{TE-S}$ (lb/hr) |
|------------------|-----------------------------|-----------------------------------|------------------------------|-------------------------|-------------------------|
| As | .57-.12 | .298 | .0148-.0312 | .008 | .0015-.0031 |
| Be | .71-80 | .298 | .1652-.1862 | .007 | .0026-.0030 |
| Cd | .31-.80 | .298 | .0025-.0066 | .01 | .0009-.0025 |
| F | 65-67 | .298 | 0 | .576 | 11.157-11.500 |
| Hg | .11-.17 | .298 | .0096-.0149 | .001 | .0000-.0001 |
| Pb | .51-12.0 | .298 | .0300-.7064 | .008 | .0015-.0343 |

Gasifier Ash (Stream 34)

Stream Constituents:

Schreiner⁸⁷ reported flow rates for the materials in Stream 34, and these are presented in Table 30.

TABLE 30 - Flow Rates of Materials in Stream 34

| <u>Constituent</u> | <u>Flow Rate</u> |
|--------------------|------------------|
| H ₂ O | 199,486 |
| Minerals | 96,031 |
| Coal | 7,038 |
| Sulfur | 66 |

The flow rates of nine trace elements in the gasifier ash were estimated using the following equation:

$$Q_{TE-34} = (C_{TE-DC})(F_{DC-G})(D_{TE-G})$$

⁸⁷ Appendix B, Stream 3

^{**} From Table 27

where-

Q_{TE-34} is the flow rate of the trace element in the gasifier ash stream, lb/hr.

C_{TE-DC} is the concentration of the trace element in the coal feed to the gasifier, ppm, dry coal basis.

F_{DC-G} is the flow rate of dry coal to the gasifier, MM lbs/hr.

D_{TE-G} is the fraction of the total quantity of this trace element fed to the gasifier that was removed with the gasifier ash.

The calculations are presented in Table 31.

TABLE 31 - Flow Rates of Trace Elements in the Gasifier Ash, Stream 34

| Trace Element | C_{TE-DC}^{40} (ppm) | x | F_{DC-G}^* (MM lb/hr) | x | D_{TE-G}^{37} | = | Q_{TE-34} (lb/hr) |
|---------------|---------------------------|---|----------------------------|---|-----------------|---|------------------------|
| As | .57-1.2 | | 1.369 | | .269 | | .2099-.4419 |
| Be | .71-.80 | | 1.369 | | .330 | | .3210-.3617 |
| Cd | .31-.80 | | 1.369 | | .519 | | .2203-.5684 |
| F | 65-67 | | 1.369 | | .563 | | 50.10-51.64 |
| Hg | .11-.17 | | 1.369 | | .519 | | .0782-.1208 |
| Pb | .51-12.0 | | 1.369 | | .942 | | .6577-15.48 |
| B | 32 | | 1.369 | | .900 | | 39.43 |
| Ni | 1.7-14.0 | | 1.369 | | .994 | | 2.313-19.05 |
| V | 10.0-14.0 | | 1.369 | | .999 | | 13.676-19.147 |

*Appendix B, Stream 2

Dewatered Gasifier Ash, Boiler Ash, and Scrubber Sludge to Final Disposal (Stream 36)

Stream Constituents:

Schreiner⁸⁷ reported flow rates for the materials in Stream 36, and these are presented in Table 32.

TABLE 32 - Flow Rates of Materials in Stream 36

| <u>Constituent</u> | <u>Flow Rate (lbs/hr)</u> |
|--------------------|-------------------------------|
| H ₂ O | 20,000 |
| Minerals | 116,989 |
| Coal | 7,038 |
| Sulfur | -66 |

The flow rates of nine trace elements in Stream 36 were estimated using the following equation.

$$Q_{TE-36} = Q_{TE-A} + Q_{TE-S} + Q_{TE-34} - (C_{TE-38})(F_{38})$$

where-

Q_{TE-36} is the flow rate of the trace element in Stream 36, lb/hr.

Q_{TE-A} is the flow rate of the trace element in the boiler bottom ash and fly ash, lb/hr.

Q_{TE-S} is the flow rate of the trace element in the scrubber sludge, lb/hr.

Q_{TE-34} is the flow rate of the trace element in the gasifier ash, Stream 34, lb/hr.

C_{TE-38} is the concentration of the trace element in the sluice water recycle, Stream 38, ppm.

F_{38} is the flow rate of Stream 38, MM lb/hr.

The calculations are presented in Table 33.

Spent F-T Catalyst (Stream 63)

Stream Constituents:

Commercial Fischer-Tropsch catalysts included cobalt (fixed-bed) and iron (fixed- and fluid-bed). Although iron was the base for both units, catalyst preparation and formulation were extremely different and very specific for each unit type.

Although spent F-T catalyst was removed periodically, not continuously, Schreiner reported the replacement rate of F-T catalyst on a continuous basis to be 30 TPSD. Catalyst life was approximately 50 days.

Due to the proprietary nature of the catalyst, very limited information was available regarding chemical composition, economics of regeneration or metal recovery from spent catalysts. It appeared that the cobalt based catalyst could be economically recovered. On-site regeneration of iron based catalyst may not be practical or economical, and hence it may be either sent to metals/catalyst vendors or disposed of as a solid waste. No data was available regarding leachate characteristics.

Leachate from Solid Wastes in Stream 36 (Stream 69)

Stream Constituents:

Ash and sludge from the boiler and gasifier were slurried and pumped to the ash handling unit where the solids were dewatered. The dewatered solids, therefore, were leached before they went to final disposal. Nevertheless, the data in the following table, which represents first column volume leachate fractions of Montana Rosebud subbituminous ash, were assumed to be representative of the leachate composition from a quenched ash. Stream components are listed in Table 34. Estimation of the leachate flow rate is presented on the following two pages.

TABLE 34 - Leachate from Montana Rosebud* Ash, Stream 69

| <u>Trace Element</u> | <u>Concentration⁹¹ (mg/l)</u> |
|--------------------------|--|
| Al | .20 |
| As | .02 |
| Cd | .00064 |
| Cu | .03 |
| Fe | 1.6 |
| Hg | .0003 |
| Mn | .50 |
| Ni | .035 |
| Pb | .09 |
| Se | .06 |
| Zn | .0029 |
| Other Trace Elements | No data |
| Organics | No data |

*The trace element composition of Montana Rosebud subbituminous coal was very similar to the trace element composition of Wyoming subbituminous coal as shown in Table 24.

Estimation of the Flow Rate of Leachate
From Ash and Scrubber Sludge Disposal

The flow rate of Stream 36* indicated that:

Total Dry Ash + Scrubber Sludge = 127,802 lb/hr.

With an onstream factor of 92% assumed,⁸⁷

Days on line per year = (365 days/hr)(92%) = 336 days/year

Total dry ash and scrubber sludge:

= (127,802 lb/hr)(336 days/year)(24 hr/day) = 1,030,595,328 lb/yr

Four values were averaged for the density of dry ash/sludge³⁶:

Density = [(72 + 75 + 105 + 116)lb/ft³] ÷ 4 = 92 lb/ft³

A density of 92 lb/ft³ was assumed so that the total volume of waste was:

$$\text{Volume} = \frac{1,030,595,328 \text{ lbs/year}}{(92 \text{ lb/ft}^3)(43,560 \text{ ft}^2/\text{acre})} = 257 \text{ acre-feet/year}$$

This waste was assumed to be disposed of by landfill to an average thickness of 30 feet³⁶. The total acreage of solid waste was:

$$\text{Acre/year} = \frac{257.2 \text{ acre-feet/year}}{30 \text{ feet}} = 8.57 \text{ acres/year}$$

Because leachate generation was directly attributable to the average net yearly precipitation/evaporation, it was best to consider rainfall levels in the U.S. as the basis for total leachate production. The range of net precipitation was from 0 to >30 inches per year. This gave a range of 0 to >815,000 gal/acre/year for leachate generation.¹⁶

The landfilling of ash and scrubber sludge required 8.57 acres per year, and the rate of production of leachate ranged from a low of 0 gallons per year to a high of: >6,984,550 gal/year. High leachate production = (8.57 acres)(>815,000) = >6,984,550 gal/year.

*Appendix B, Stream 36

Biological Sludge (Stream 70)

Stream Constituents:

Flow rate of wastewater to the biotreatment unit was 1,990,190 lbs/hr.*

$$1,990,190 \text{ lbs/hr} = 3,977 \text{ gpm}$$

The WESCO EIS reported¹⁰⁸ that 6.7 TPD of dry biological solids were produced from a wastewater flow of 2,810 gpm.

$$6.7 \text{ TPD} \times 2000 \text{ lbs/ton} \times 1 \text{ day/24 hrs} \times 1 \text{ hr/60 min} = 9.3 \text{ lbs/min}$$

$$\frac{9.3 \text{ lbs/min dry biological sludge}}{2810 \text{ gpm wastewater}} = .0033 \text{ lbs/gal}$$

Sludge production was adjusted to flow rate from Schreiner:⁸⁷

$$.0033 \text{ lbs/gal} \times 3977 \text{ gpm} = 13 \text{ lbs/min} = 780 \text{ lbs/hr. dry solids.}$$

Vacuum filtration was assumed to produce a solid concentration of 20%:

$$780 + 0.2 = 3,900 \text{ lbs/hr biological sludge.}$$

Removal in the activated sludge unit was accomplished through absorption and sedimentation for the biorefractory compounds. The biorefractory compounds removed from the water were contained in the biosludge. The flow rate of biorefractory compounds removed with the biosludge was:

$$F_B = (C_{52}) (R_{52}) (F_{52})$$

where-

F_B is the flow rate of each biorefractory compound in the biosludge (lb/hr.)

C_{52} is the concentration of each biorefractory compound in Stream 52 (ppm_w)

R_{52} is the efficiency of removal of each compound in the biotreatment unit

F_{52} is the flow rate of Stream 52 (10⁶ gal/hr.)

*Appendix B, Stream 50

Calculations are presented in Table 35.

TABLE 35 - Biological Sludge, Stream 70

| Compound | C_{52} Removal (ppm)* | R_{52} Flow Rate Efficiency** | F_{52} Biosludge (10^6 lb/hr) | F_B (lb/hr) |
|----------------------|----------------------------|------------------------------------|---------------------------------------|---------------|
| Acenaphthylene | 0.3403 | 0.9 | 1.99 | 0.61 |
| Anthracene | 0.0851 | 0.9 | 1.99 | 0.15 |
| Benz(a)anthracene | 0.00851 | 0.9 | 1.99 | 0.015 |
| Benzo(g,h,i)perylene | 0.00026 | 0.9 | 1.99 | 0.00046 |
| Benzo(a)pyrene | 0.0034 | 0.9 | 1.99 | 0.0061 |
| Benzo(e)pyrene | 0.0034 | 0.9 | 1.99 | 0.0061 |
| Chrysene | 0.0017 | 0.9 | 1.99 | 0.003 |
| Fluoranthene | 0.1702 | 0.9 | 1.99 | 0.31 |
| Fluorene | 0.1702 | 0.9 | 1.99 | 0.31 |
| Naphthalene | 2.723 | 0.70 | 1.99 | 3.80 |
| Phenanthrene | 0.0851 | 0.90 | 1.99 | 0.15 |
| Pyrene | 0.1702 | 0.90 | 1.99 | 0.30 |
| Arsenic | 0.5545 | 0.50 | 1.99 | 0.55 |
| Beryllium | 0.2932 | 0.50 | 1.99 | 0.29 |
| Cadmium | 0.2503 | 0.325 | 1.99 | 0.16 |
| Mercury | 0.0486 | 0.475 | 1.99 | 0.046 |
| Lead | 0.1403 | 0.70 | 1.99 | 0.20 |

*Estimated values. See Table 10, column 8.

**See Table 7.

Leachate from Biosludge (Stream 71)

Stream Constituents:

No data available.

Spent Shift Catalyst (Stream 79)

Stream Constituents:

The average useful life of shift catalyst was 3 years.* An indirect liquefaction plant producing 45,000 BPSD fuel oil equivalent had an inventory of about 400 tons of shift catalyst.** Therefore once every 3 years 400 tons of spent shift catalyst was replaced. The spent catalyst may either be disposed of or reclaimed. Because of the high price of cobalt it is likely that spent shift catalyst will be reclaimed. The trace element content of spent shift catalyst is given in Table 36.

TABLE 36 - Spent Shift Catalyst, Stream 79

| <u>Constituent Elements</u> | <u>Quantity (% wt)*</u> |
|-----------------------------|-------------------------|
| Cobalt | 5-15 |
| Molybdenum | 15-25 |
| Sulfur | 5-20 |
| Carbon | 5-10 |
| Arsenic | 0.7-7.0 |
| Lead | 4-10 |
| Selenium | 0.3-1 |
| Mercury | 0.2-0.8 |

*Unpublished EPA data

**Calculated from unpublished EPA data

3 ESTIMATION OF MAXIMUM POST-DILUTION POLLUTANT CONCENTRATIONS RESULTING FROM INDIRECT LIQUEFACTION STREAMS

3.1 Introduction

Maximum post-dilution concentrations of various pollutants resulting from discharge streams from indirect liquefaction facilities were estimated to provide a basis for subsequent analytical tasks in the risk assessment.

Due to the relatively high degree of uncertainty regarding estimated pollutant concentrations in discharge streams, and the total lack of site specific environmental information, sophisticated modeling of pollutant transport, transformation and fate was not warranted. Instead, maximum post-dilution pollutant concentrations resulting from the various facility discharges were estimated using either dilution factors based on the Source Analysis Model (SAM/I)¹¹² or very simple models.

Short descriptions of the assumptions and calculations used to estimate the pollutant post-dilution concentrations resulting from the release of selected facility streams are presented in Section 3.3 in addition to tables of results. A short description of the Source Analysis Model (SAM/I) is presented in Appendix C.

3.2 Caveats

These estimates of post-dilution concentrations have been prepared to provide a basis for conducting subsequent analytical tasks of the assessment of indirect liquefaction. While a considerable amount of effort has been expended in developing these estimates, users of the data are reminded and cautioned that, at best, the estimates are very rough due to a variety of problems, including but not limited to the following:

- Characterization data (qualitative and quantitative) for process and effluent streams from Lurgi/Fischer-Tropsch plants were very limited. Data used in developing these estimates came from a variety of sources and may not be representative of commercial Lurgi/Fischer-Tropsch plants.

- Dilution factors used to estimate post-dilution concentrations were not representative of any particular plant or site; they were general values derived by modeling nationwide average conditions.
- Environmental interactions and transformations were not considered in the estimates.
- Using SAM/I, dilution factors and resultant post-dilution concentrations were a function of discharge stream flow rates; discharge stream flow rates used in developing the estimates were derived from the conceptual plant flow diagram/matrix, which may or may not be truly representative of a commercial Lurgi/Fischer-Tropsch plant.
- The estimates of post-dilution concentrations presented in Tables 37 through 42 actually represent only incremental increases in pollutants due to the discharge of a single stream, i.e., the estimates did not include environmental background concentrations, nor did they include pollutant loadings from other discharge streams from the indirect liquefaction facility.

3.3 Assumptions, Calculations and Results

Estimation of Pollutant Post-Dilution Concentrations Due to Release of Stream 28 - Utility Stack Gas

Post-dilution concentrations of pollutants due to release of Stream 28 were estimated using a dilution factor from the Source Analysis Model (SAM/I). The assumptions and calculations are presented below. The results are presented in Table 37.

Assumptions:

- Ambient temperature = 25°C, pressure = 1 atm.
- Mean density of Stream 28 - Mean density of air = 1,180 gm/m³

Calculation:

$$PC = \frac{(EC) \left(\frac{\mu g}{gm} / PPM_w \right) (\text{Density of air})}{DF}$$

where-

PC is the post-dilution concentration, in $\mu g/m^3$

EC is the emission stream concentration in PPM_w

Density of air is 1,180 gm/m³

DF is the SAM/I dilution factor, which = 235 for a gaseous stream being discharged into the atmosphere at a flow rate of 8.75×10^5 gm/sec.

$$\frac{\mu\text{g}/\text{PPM}_w}{\text{gm}} = 1.0$$

TABLE 37

ESTIMATED POST-DILUTION CONCENTRATIONS DUE TO
DISCHARGE OF STREAM 28-UTILITY STACK GAS

| Stream Component | Emission Stream Concentration* (PPM _w) | Post-Dilution Concentration (μg/m ³) |
|-------------------------|--|--|
| CO ₂ | 3.9 E+5 | 2.0 E+6 |
| N ₂ + Inerts | 5.2 E+5 | 2.6 E+6 |
| O ₂ | 1.0 E+4 | 5.2 E+4 |
| H ₂ O | 7.2 E+4 | 3.6 E+5 |
| SO _x | 1.4 E+2 | 7.0 E+2 |
| NO _x | 2.5 E+1 | 1.3 E+2 |
| Arsenic | 1.2 E-4** | 5.8 E-4 |
| Beryllium | 1.2 E-4** | 6.1 E-4 |
| Cadmium | 1.0 E-4** | 5.1 E-4 |
| Lead | 1.2 E-3** | 6.0 E-3 |
| Mercury | 7.8 E-3** | 3.9 E-2 |
| Particulates | 4.0 E-1 | 2.0 E 0 |

*See Section 2.3

**Based upon higher value of the estimated range of concentration

Estimation of Pollutant Concentrations Due to
Release of Stream 29 - Atmospheric Losses from Cooling Tower

The risks due to Stream 29 were analyzed differently for public health and ecosystems. Risks to public health were evaluated assuming that essentially all of the atmospheric emissions from the cooling tower were in the gaseous state. This assumption was not strictly accurate. However it appeared to be reasonable in view of the fact that drift would comprise a relatively small part (5%) of the atmospheric losses and it was assumed that the receptor population was situated close to the liquefaction facility.

The uncertainty regarding the transport and fate of drift and the partitioning of pollutants between the drift and evaporative losses appeared to be of greater importance in the ecological evaluation than in the public health evaluation. The ecological evaluation was based on pollutant emission rates rather than post-dilution concentrations.

Post-dilution Concentrations for Public Health:

Atmospheric post-dilution concentrations of pollutants due to discharge of Stream 29 were estimated using a dilution factor from the Source Analysis Model (SAM/I). Assumptions and calculations used in estimating post-dilution concentrations and emission rates are presented below. Results are presented in Table 38.

TABLE 38
ESTIMATED POST-DILUTION CONCENTRATIONS DUE TO
DISCHARGE OF STREAM 29 - ATMOSPHERIC LOSSES FROM COOLING TOWER

| Stream Component | Conc. in Cooling Water (PPM _w) | Conc. in Atmospheric Emissions (PPM _w) | Public Health Post-dilution Concentration (ug/m ³) | Ecological Emmission Rate (mg/sec) |
|------------------------|--|--|--|------------------------------------|
| Ammonia | 2.5E+1 | 4.9E-1 | 1.9E+1 | 4.3E+2 |
| Arsenic | 1.8E-2 | 3.5E-4 | 1.3E-2 | 5.5E 0 |
| Beryllium | 8.5E-3 | 1.7E-4 | 6.5E-3 | 2.6E 0 |
| Boron | 5.6E-2 | 1.1E-3 | 4.3E-2 | 1.7E+1 |
| Cadmium | 9.8E-3 | 1.9E-4 | 7.5E-3 | 3.0E 0 |
| Fluorine | 4.1E-1 | 8.0E-3 | 3.1E-1 | 1.2E+2 |
| Lead | 1.9E-2 | 3.7E-4 | 1.5E-2 | 5.8E 0 |
| Mercury | 1.5E-3 | 2.9E-5 | 1.1E-3 | 4.6E-1 |
| Nickel | 5.0E-4 | 9.8E-6 | 4.0E-4 | 1.5E-1 |
| Vanadium | 1.5E-4 | 2.9E-6 | 1.1E-4 | 4.6E-2 |
| Acetic Acid | 2.1E 0 | 4.1E-2 | 1.6E 0 | 6.4E+2 |
| Aniline | 1.3E-3 | 2.6E-5 | 1.0E-3 | 4.0E-1 |
| Butanoic Acid | 1.2E-1 | 2.3E-3 | 9.1E-2 | 3.6E+1 |
| Catechol | 9.5E-1 | 1.9E-2 | 7.2E-1 | 2.9E+2 |
| Hexanoic Acid | 9.1E-3 | 1.8E-4 | 6.9E-3 | 2.8E 0 |
| Pentanoic Acid | 1.1E-1 | 2.2E-3 | 8.4E-2 | 3.3E+1 |
| Phenol | 2.8E-2 | 5.5E-4 | 2.1E-2 | 8.5E 0 |
| Propanoic Acid | 2.2E-1 | 4.3E-3 | 1.7E-1 | 6.7E+1 |
| Pyridine | 1.8E-3 | 3.5E-5 | 1.4E-3 | 5.5E-1 |
| Resorcinol | 3.2E-1 | 6.3E-3 | 2.4E-1 | 9.7E+1 |
| 2-Methylphenol | 7.7E-3 | 1.5E-4 | 5.9E-3 | 2.3E 0 |
| 2-Methylpropionic Acid | 1.8E-2 | 3.5E-4 | 1.4E-2 | 5.5E 0 |

TABLE 38 (Continued).

| Stream Component | Conc. in Cooling Water (PPM _w) | Conc. in Atmospheric Emissions (PPM _w) | Public Health Post-dilution Concentration (µg/m ³) | Ecological Emission Rate (mg/sec) |
|-----------------------|--|--|--|-----------------------------------|
| 2-Methylpyridine | 1.1E-2 | 2.2E-4 | 8.4E-3 | 3.3E 0 |
| 2,4-Dimethylpyridine | 1.5E-4 | 2.9E-6 | 1.1E-4 | 4.6E-2 |
| 2,4-Xylenol | 6.6E-3 | 1.3E-4 | 5.0E-3 | 2.0E 0 |
| 2,5-Dimethylpyridine | 1.5E-4 | 2.9E-6 | 1.1E-4 | 4.6E-2 |
| 3-Methylbutanoic Acid | 9.1E-3 | 1.8E-4 | 6.9E-3 | 2.8E 0 |
| 3-Methylcatechol | 0.0E 0 | 0.0E 0 | 0.0E 0 | 0.0E 0 |
| 3-Methylphenol | 4.7E-3 | 9.2E-5 | 3.6E-3 | 1.4E 0 |
| 3-Methylpyridine | 3.9E-3 | 7.6E-5 | 3.0E-3 | 1.2E 0 |
| 3,5-Xylenol | 9.1E-3 | 1.8E-4 | 6.9E-3 | 2.9E+1 |
| 3,6-Dimethylcatechol | 7.8E-2 | 1.5E-3 | 5.9E-2 | 2.4E+1 |
| 4-Methyl Resorcinol | 6.2E-2 | 1.2E-3 | 4.7E-2 | 1.9E+1 |
| 4-Methylcatechol | 6.7E-1 | 1.3E-2 | 5.1E-1 | 2.0E+2 |
| 4-Methylphenol | 6.5E-3 | 1.3E-4 | 4.9E-3 | 2.0E 0 |
| 4-Methylpyridine | 9.0E-4 | 1.8E-5 | 6.9E-4 | 2.7E-1 |
| 5-Methyl Resorcinol | 1.1E-1 | 2.2E-3 | 8.4E-2 | 3.3E+1 |
| Acenaphthylene | 2.0E-4 | 3.9E-6 | 1.5E-4 | 6.1E-2 |
| Anthracene | 4.9E-5 | 9.6E-7 | 3.7E-6 | 1.5E-2 |
| Benz(a)anthracene | 4.9E-6 | 9.6E-8 | 3.7E-6 | 1.5E-3 |
| Benzo(a)pyrene | 2.0E-6 | 3.9E-8 | 1.5E-6 | 6.1E-4 |

TABLE 38 (Concluded)

| Stream Component | Conc. in Cooling Water (PPM _w) | Conc. in Atmospheric Emissions (PPM _w) | Public Health Post-dilution Concentration ($\mu\text{g}/\text{m}^3$) | Ecological Emission Rate (mg/sec) |
|----------------------|---|---|---|--------------------------------------|
| Benzo(e)pyrene | 2.0E-6 | 3.9E-8 | 1.5E-6 | 6.1E-4 |
| Benzo(g,h,i)perylene | 1.5E-7 | 2.9E-9 | 1.4E-2 | 4.6E-5 |
| Biphenyl | 1.0E-3 | 2.0E-5 | 7.6E-4 | 3.0E-1 |
| Chrysene | 9.8E-7 | 1.9E-8 | 7.5E-7 | 3.0E-4 |
| Dibenzofuran | 1.3E-3 | 2.5E-5 | 1.0E-3 | 4.0E-1 |
| Ethylbenzene | 2.8E-1 | 5.5E-3 | 2.1E-1 | 8.5E+1 |
| Fluoranthene | 9.8E-5 | 1.9E-6 | 7.4E-5 | 3.0E-2 |
| Fluorene | 9.8E-5 | 1.9E-6 | 7.4E-5 | 3.0E-2 |
| Indan | 1.3E-2 | 2.5E-4 | 9.9E-3 | 4.0E 0 |
| Methanethiol | 1.1E-1 | 2.2E-3 | 8.4E-2 | 3.3E+1 |
| Naphthalene | 4.7E-3 | 9.2E-5 | 3.6E-3 | 1.4E 0 |
| Perylene | 2.0E-7 | 3.9E-9 | 1.5E-7 | 6.1E-5 |
| Phenanthrene | 4.9E-5 | 9.6E-7 | 3.7E-5 | 1.5E-2 |
| Pyrene | 9.8E-5 | 1.9E-6 | 7.4E-5 | 3.0E-2 |
| Quinoline | 1.4E-3 | 2.8E-5 | 1.1E-3 | 4.3E-1 |
| Toluene | 9.1E-1 | 1.8E-2 | 6.9E-1 | 2.8E+2 |
| o-Xylene | 3.2E-1 | 6.2E-3 | 2.4E-1 | 9.7E+1 |

Assumptions:

- Ambient temperature - 25°C, pressure - 1 atm.
- Evaporative losses accounted for essentially all of the atmospheric losses from the cooling tower.*
- Relative humidity of influent air was 0%, and of effluent air was 100%.
- Water content of air was 0.02 lb water/lb of dry air.⁷⁵
- Density of atmospheric emission stream = density of air = 1,180 gm/m³.

Calculation:

The total flow rate of moisture and air from the cooling tower was calculated using the following:

$$TF = WF + AF$$

where-

TF is the total flow rate of air and moisture

WF is the flow rate of moisture

AF is the flow rate of air

At 25°C and 1 ATM, the moisture content of air was 0.02 lb water/lb dry air. Therefore,

$$AF = (WF) (1 \text{ lb}/0.02 \text{ lb}) = 50 \text{ WF}$$

and

$$TF = WF + 50WF = 51 \text{ WF}$$

According to Schreiner, the flow rate of water out of the cooling tower was 2.413×10^6 lb/hr or 3.04×10^5 gm/sec. Therefore,

$$TF = (51)(3.04 \times 10^5 \text{ gm/sec}) = 1.55 \times 10^7 \text{ gm/sec.}$$

*According to Schreiner⁸⁷ drift accounted for 5% and evaporative losses 95%, but to simplify the estimates it was assumed that all of the atmospheric loss was evaporative in nature.

The concentration of each pollutant in the total emission stream (air + water) was calculated by the following:

$$EC = \frac{(CW)(WF)}{TF}$$

where-

EC is the pollutant concentration in the total emission stream, in PPM_w

CW is the pollutant concentration in the cooling tower feed water, in PPM_w

WF is the flow rate of the water in the atmospheric stream, 3.04×10^5 gm/sec

TF is the total flow rate of air and moisture.

But, $TF = 51 \times WF$

therefore-

$$EC = \frac{(CW)(WF)}{51 WF} = \frac{CW}{51}$$

The post-dilution concentration of each pollutant was then calculated:

$$PC = \frac{(EC) \left(\frac{\mu g}{gm} / PPM_w \right) (\text{Density of Air})}{DF}$$

where-

PC is the pollutant post-dilution concentration, in $\mu g/m^3$

EC is the pollutant concentration in the emission stream, in PPM_w .

$\frac{\mu g}{gm} / PPM_w$ is a conversion factor, equal to 1.0

Density of air is $1,180 \text{ gm/m}^3$

DF is the SAM/I dilution factor, which was 30.4 for a gaseous discharge to the atmosphere at a flow rate of 1.55×10^7 gm/sec.

Emission Rates for Ecosystems:

From an ecological perspective, the partitioning of contaminants between drift and evaporation losses was important because drift tends

to settle over a more limited area, relatively close to the plant, than the evaporative losses. Thus, whatever contaminants were present in the drift would tend to become enriched on the surface of vegetation and in soils within one or two kilometers of the plant. Because data were so limited and partitioning was important, the ecological risks were evaluated on the basis of pollutant emission rates rather than post-dilution concentrations in the atmosphere. Two emission rates were calculated as follows:

$$E = (C)(F)$$

where-

E is the emission rate, in mg/sec

C is the pollutant concentration in water fed to the cooling tower, in mg/l

F is the flow rate of Stream 29, i.e., 304 l/sec.

Estimation of Pollutant Post-Dilution Concentrations Due to Disposal of Stream 53 - Reverse Osmosis Concentrated Waste Solution

Two sets of post-dilution concentrations were estimated for disposal of Stream 53. The set used to evaluate risks to public health was developed assuming that groundwater contamination would be the major concern; the set used to evaluate risks to ecosystems was developed assuming contamination of surface streams would be the primary cause of concern. The assumptions and calculations are presented below. The results are presented in Table 39.

TABLE 39
ESTIMATED POST-DILUTION CONCENTRATIONS DUE TO
DISCHARGE OF STREAM 53 - REVERSE OSMOSIS WASTE SOLUTION

| Stream Component | Effluent Stream Concentration (PPM _w) | Public Health Post-dilution Concentration (PPM _w) | Ecological Post-dilution Concentration (PPM _w) |
|------------------|---|---|--|
| Arsenic | 8.9E-1 | 8.9E-2 | 2.3E-1 |
| Beryllium | 4.6E-1 | 4.6E-2 | 1.2E-1 |
| Boron | 3.1E 0 | 3.1E-1 | 7.9E-1 |
| Cadmium | 5.3E-1 | 5.3E-2 | 1.3E-1 |
| Fluorine | 3.3E+1 | 3.3E 0 | 8.3E 0 |
| Lead | 1.7E-1 | 1.7E-2 | 4.3E-2 |
| Manganese | 8.8E-1 | 8.8E-2 | 2.2E-1 |
| Mercury | 8.1E-2 | 8.1E-3 | 2.1E-2 |
| Nickel | 1.5E-1 | 1.5E-2 | 3.8E-2 |
| Vanadium | 1.8E-2 | 1.8E-3 | 4.5E-3 |
| Acetic Acid | 1.3E+1 | 1.3E 0 | 3.3E 0 |
| Aniline | 1.3E-2 | 1.3E-3 | 3.3E-3 |
| Butanoic Acid | 1.2E-0 | 1.2E-1 | 3.1E-1 |
| Catechol | 1.7E+1 | 1.7E 0 | 4.4E 0 |
| Hexanoic Acid | 9.4E-2 | 9.4E-3 | 2.4E-2 |
| Pentanoic Acid | 1.1E 0 | 1.1E+1 | 2.8E-1 |
| Phenol | 3.8E-1 | 3.8E-2 | 9.6E-2 |
| Propanoic Acid | 2.5E 0 | 2.5E+1 | 6.3E-1 |

TABLE 39 (Cont.)

| Stream Component | Effluent Stream Concentration (PPM _w) | Public Health Post-Dilution Concentration (PPM _w) | Ecological Post-Dilution Concentration (PPM _w) |
|------------------------|---|---|--|
| Pyridine | 3.0E-2 | 3.0E-3 | 7.5E-3 |
| Resorcinol | 5.8E 0 | 5.8E-1 | 1.5E 0 |
| 2-Methylphenol | 1.3E-1 | 1.3E-2 | 3.3E-2 |
| 2-Methylpropionic Acid | 1.9E-1 | 1.9E-2 | 4.8E-2 |
| 2-Methylpyridine | 1.8E-1 | 1.8E-2 | 4.6E-2 |
| 2,4-Dimethylpyridine | 2.6E-3 | 2.6E-4 | 6.6E-4 |
| 2,4-Xylenol | 3.6E-1 | 3.6E-2 | 9.1E-2 |
| 2,5-Dimethylpyridine | 2.7E-3 | 2.6E-4 | 6.6E-4 |
| 3-Methylbutanoic Acid | 9.4E-2 | 9.4E-3 | 2.4E-2 |
| 3-Methylcatechol | 0.0E 0 | 0.0E 0 | 0.0E 0 |
| 3-Methylphenol | 1.9E-1 | 1.9E-2 | 4.8E-2 |
| 3-Methylpyridine | 6.7E-2 | 6.7E-2 | 1.7E-2 |
| 3,5-Xylenol | 5.0E-1 | 5.0E-2 | 1.3E-1 |
| 3,6-Dimethylcatechol | 1.4E 0 | 1.4E-1 | 3.6E-1 |
| 4-Methyl Resorcinol | 1.1E 0 | 1.1E-1 | 2.8E-1 |
| 4-Methylcatechol | 1.2E+1 | 1.2E 0 | 3.1E 0 |
| 4-Methylphenol | 1.2E-1 | 1.2E-2 | 3.0E-2 |
| 4-Methylpyridine | 1.6E-2 | 1.6E-3 | 4.1E-3 |
| 5-Methyl Resorcinol | 2.1E 0 | 2.1E-1 | 5.3E-1 |
| Acenaphthylene | 1.2E-1 | 1.2E-2 | 3.0E-2 |

TABLE 39 (Concluded)

| Stream Component | Effluent Stream Concentration (PPM _w) | Public Health Post-dilution Concentration (PPM _w) | Ecological Post-dilution Concentration (PPM _w) |
|----------------------|---|---|--|
| Anthracene | 3.0E-2 | 3.0E-3 | 7.6E-3 |
| Benz(a)anthracene | 3.0E-3 | 3.0E-3 | 7.6E-4 |
| Benzo(a)pyrene | 1.2E-3 | 1.2E-4 | 3.0E-4 |
| Benzo(e)pyrene | 1.2E-3 | 1.2E-4 | 3.0E-4 |
| Benzo(g,h,i)perylene | 8.9E-5 | 8.9E-6 | 2.2E-5 |
| Biphenyl | 6.2E-1 | 6.2E-2 | 1.6E-1 |
| Chrysene | 5.9E-4 | 5.9E-5 | 1.5E-4 |
| Dibenzofuran | 2.2E-2 | 2.2E-3 | 5.6E-3 |
| Ethylbenzene | 2.9E 0 | 2.9E-1 | 7.3E-1 |
| Fluoranthene | 5.9E-2 | 5.9E-3 | 1.5E-2 |
| Fluorene | 5.9E-2 | 5.9E-3 | 1.5E-2 |
| Indan | 7.9E 0 | 7.9E-1 | 2.0E 0 |
| Methanethiol | 5.9E 0 | 5.9E-1 | 1.5E 0 |
| Naphthalene | 2.8E 0 | 2.8E-1 | 7.1E-1 |
| Perylene | 1.2E-4 | 1.2E-5 | 3.0E-5 |
| Phenanthrene | 3.0E-2 | 3.0E-3 | 7.6E-3 |
| Pyrene | 5.9E-2 | 5.9E-3 | 1.5E-2 |
| Quinoline | 2.5E-2 | 2.5E-3 | 6.3E-3 |
| Toluene | 9.4E 0 | 9.4E-1 | 2.4E 0 |
| o-Xylene | 3.3E 0 | 3.3E-1 | 8.3E-1 |

Post-dilution Concentrations for Public Health:

The risk to public health posed by Stream 53 was assumed to originate from improper disposal of the stream and subsequent contamination of the groundwater. The dilution factor was derived from the Source Analysis Model (SAM/I). Assumptions and calculations were as follows.

Assumptions:

- Stream 53 disposed of in a sump or fill site.
- Receiving body (sump or fill site) had a base diameter > 10m.
- SAM/I dilution factors were appropriate for estimating subsequent pollutant concentrations in groundwater.
- Local population used groundwater directly for domestic needs.
- Flow rate of Stream 53 = 569,000 lb/hr (71,700 gm/sec).

Calculations

Post-dilution concentrations of pollutants from Stream 53 for the public health analysis were calculated using the following equation:

$$PC = \frac{EC}{DF}$$

where-

PC is the post-dilution concentration in PPM_w or mg/l.

EC is the concentration in effluent Stream 53, in PPM_w.

DF is the SAM/I dilution factor, which was 10, for a liquid stream discharged into a sump or fill site with a diameter > 10 meters.

Post-dilution Concentrations for Ecosystems:

The risk to ecosystems posed by Stream 53 was assumed to originate from disposal of the stream and subsequent leakage of all of the stream contaminants into a small surface stream. A SAM/I dilution factor was not used.

Assumptions:

- Stream 53 was disposed of in an inadequate evaporation pond.
- All of the stream components except water were subsequently released into a small (10 cfs) surface stream through overflow or by percolation through the soil.
- Loss of stream contaminants through physical, chemical and biological processes prior to entry into the surface stream was negligible.
- Evaporation of the water components of stream 53 in the evaporation pond was significant, resulting in extreme concentration of stream contaminants prior to the entry into the surface stream.

Calculations:

Post-dilution concentrations of pollutants from Stream 53 for the ecological analysis were calculated using the following equation:

$$PC = \frac{(EC)(FR_{53})}{FR_S}$$

where-

PC is the post-dilution concentrations, in PPM_w.

EC is the constituent concentration in Stream 53, in PPM_w.

FR₅₃ is the flow rate of Stream 53, 71.7 gm/sec.

FR_S is the flow rate of the surface stream, 283 l/sec.

Estimation of Pollutant Post-Dilution Concentrations Due to
the Discharge of Stream 69 - Leachate from Gasifier and Utility Ash

Two sets of post-dilution concentrations were estimated for discharge of contaminants in Stream 69. The set used to evaluate risks to public health was developed assuming that groundwater contamination would be the major concern, while the set used to evaluate risks to ecosystems was developed assuming contamination of surface streams would be the primary cause for concern. The assumptions and calculations are presented below. The results are presented in Table 40.

TABLE 40

ESTIMATED POST-DILUTION CONCENTRATIONS DUE TO
DISCHARGE OF STREAM 69 ASH LEACHATE

| Stream Component | Effluent Stream Concentration (PPM _w) | Public Health Post-dilution Concentration (PPM _w) | Ecological Post-dilution Concentration (PPM _w) |
|------------------|---|---|--|
| Aluminum | 2.0E-1 | 2.0E-2 | 6.0E-4 |
| Arsenic | 2.0E-1 | 2.0E-3 | 6.0E-5 |
| Cadmium | 6.4E-4 | 6.4E-5 | 1.9E-6 |
| Copper | 3.0E-2 | 3.0E-3 | 9.0E-5 |
| Iron | 1.6E 0 | 1.6E-1 | 4.8E-3 |
| Lead | 9.0E-2 | 9.0E-3 | 2.7E-4 |
| Manganese | 5.0E-1 | 5.0E-2 | 1.5E-3 |
| Mercury | 3.0E-4 | 3.0E-5 | 9.0E-7 |
| Nickel | 3.5E-2 | 3.5E-3 | 1.1E-4 |
| Selenium | 6.0E-2 | 6.0E-3 | 1.8E-4 |
| Zinc | 2.9E-3 | 2.9E-4 | 1.8E-4 |

Post-dilution Concentrations for Public Health:

The risk to public health was assumed to originate from improper disposal of the stream and subsequent contamination of groundwater. The dilution factor was derived from the Source Analysis Model (SAM/I). Assumptions and calculations were as follows:

Assumptions:

- Gasifier and utility ash (Stream 36) were the source of the leachate.
- The ash was disposed of at a fill site.
- The fill site had a base diameter ≥ 10 meters.
- SAM/I dilution factors were appropriate for estimating subsequent pollutant concentrations in groundwater.
- The local population used the groundwater directly for domestic uses.
- Flow rate of the leachate = 7×10^6 gallons/year.*

Calculations:

Post-dilution concentrations of pollutants from Stream 69 for the public health analysis were calculated using the following equation:

$$PC = \frac{EC}{DF}$$

where-

PC is the pollutant post-dilution concentrations, in PPM_w

EC is the pollutant concentration in effluent Stream 69, in PPM_w*

DF is the SAM/I dilution factor, which was 10, for a liquid stream discharged in a fill site with a base diameter ≥ 10 meters.

*See Section 2.4

Post-dilution Concentrations for Ecosystems:

The risk to ecosystems posed by Stream 69 was assumed to originate from disposal of the ash and subsequent contamination of a small surface stream. A SAM/I dilution factor was not used.

Assumptions:

- Flow rate of the leachate = 7.0×10^6 gal/year* or 0.85 liters/sec.
- Pollutant concentrations in the leachate were not reduced by physical, chemical or biological processes prior to entry into a small surface stream.
- The surface stream flow rate was 10 cubic feet/sec (283 l/sec).

Calculations:

The post-dilution concentration of each pollutant due to release of Stream 69 was calculated using the following:

$$PC = \frac{(EC)(LF)}{SF} = \frac{(EC)(.85 \text{ l/sec})}{283 \text{ l/sec}} = \frac{EC}{333}$$

where-

PC is the post-dilution concentrations, in PPM_w.

EC is the concentration in the effluent stream, in PPM_w.

LF is the flow rate of the leachate, equal to 0.85 l/sec.

SF is the flow rate of the small surface stream, 283 l/sec.

Estimation of Pollutant Post-Dilution Concentrations Due to Disposal of Stream 70 - Biosludge

The post-dilution concentrations of pollutants, which might result from disposal of Stream 70, biosludge, were estimated assuming that all of the trace elements and toxic organic compounds in the biosludge were made available to a small surface stream through

*See Section 2.4

leaching. Assumptions and calculations are presented below. Results are presented in Table 41.

Assumptions:

- Biosludge was disposed of in a landfill.
- All of the trace metals and toxic organics were leached out of the biosludge and enter a surface stream.
- The concentrations of leachate contaminants were not reduced by physical, chemical or biological processes prior to entry into the surface stream.
- The flow rate of the surface stream was 10 cfs.
- The water component of the leachate was negligible.

Calculations

The concentration of each pollutant in the surface stream was calculated as follows:

$$PC = \frac{(PR)(1000\text{mg/gm})}{SF}$$

where-

PC is the post-dilution concentration, in PPM_w

PR is the production rate of the trace elements and organics, in Stream 70*, in gm/sec.

SF is the flow rate of the surface stream, 283 l/sec.

* See Table 35 for pollutant production rate.

TABLE 41
 ESTIMATED POST-DILUTION CONCENTRATION DUE TO
 DISPOSAL OF STREAM 70 - BIOSLUDGE

| Stream Component | Biosludge Removal Rate (gm/sec) | Post-dilution Concentration (PPM _w) |
|---------------------|---------------------------------|---|
| Arsenic | 6.9E-2 | 2.4E-1 |
| Beryllium | 3.7E-2 | 1.3E-1 |
| Cadmium | 2.0E-2 | 7.1E-2 |
| Mercury | 5.8E-3 | 2.0E-2 |
| Lead | 2.5E-2 | 8.9E-2 |
| Acenaphthylene | 7.7E-2 | 2.7E-1 |
| Anthracene | 1.9E-2 | 6.7E-2 |
| Benz(a)anthracene | 1.9E-3 | 6.7E-3 |
| Benz(g,h,i)perylene | 5.8E-5 | 2.0E-4 |
| Benzo(a)pyrene | 7.7E-4 | 2.7E-3 |
| Benzo(e)pyrene | 7.7E-4 | 2.7E-3 |
| Chrysene | 3.8E-4 | 1.3E-3 |
| Fluoranthene | 3.9E-2 | 1.4E-1 |
| Fluorene | 3.9E-2 | 1.4E-1 |
| Naphthalene | 4.8E-1 | 1.7E 0 |
| Phenanthrene | 1.9E-2 | 6.7E-2 |
| Pyrene | 3.8E-2 | 1.3E-1 |

**Estimation of Pollutant Post-Dilution Concentrations Due to
Release of Stream 72 - Lockhopper Vent Gas**

The post-dilution concentrations of pollutants which may result from the discharge of Stream 72 were estimated using a dilution factor from the Source Analysis Model (SAM/I). The assumptions and calculations are presented below. The results are presented in Table 42.

Assumptions:

- Ambient temperature = 25°C, pressure = 1 atm.
- Mean density of Stream 72 = 1,180 gm/m³ (25°, 1 atm)

Calculation:

$$PC = \frac{(EC) \left(\frac{\mu g}{gm/PPM_w} \right) (\text{Density of Air})}{DF}$$

where-

PC is the post-dilution concentration, in $\mu g/m^3$

EC is the concentration in Stream 72 in PPM_w*

$\left(\frac{\mu g}{gm} / PPM_w \right)$ is a conversion factor = 1.0

Density of air is 1,180 gm/m³

DF is the SAM/I dilution factor, which was 24,000 for a gaseous stream being discharged into the atmosphere at a rate of 493 gm/sec.

3.4 Applications of Results

The information presented in Volume II represents estimates of maximum post-dilution concentrations for trace elements and organic compounds released by a model, commercial-scale indirect liquefaction facility. The ecological, public health and occupational health hazards posed by this model facility are based upon these estimates and are described in Volumes III, IV and V, respectively. Volume I presents a summary of Volumes II through V and overall conclusions made in the assessment.

*See Section 2.2

TABLE 42

ESTIMATED POST-DILUTION CONCENTRATIONS DUE TO
DISCHARGE OF STREAM 72 - LOCKHOPPER VENT GAS

| Stream Component | Emission Stream Concentration (PPM _w) | Post-dilution Concentration ($\mu\text{g}/\text{m}^3$) |
|-------------------------------|--|--|
| CO | 1.5E+5 | 7.4E+3 |
| CH ₄ | 5.2E+4 | 2.5E+3 |
| C ₂ H ₄ | 6.4E+2 | 3.1E+1 |
| H ₂ S | 1.4E+3 | 6.6E+1 |
| COS | 3.0E+1 | 1.4E 0 |
| Ni(CO) ₄ | 1.5E 0 | 7.4E-2 |
| Tars | 1.1E+3 | 5.4E+1 |
| Oils | 9.9E+3 | 4.9E+2 |
| Naphtha | 4.0E+3 | 1.9E+2 |
| Phenols | 1.3E+3 | 6.4E+1 |
| Mercaptans | 5.1E 0 | 2.5E-1 |
| Thiophenes | 4.1E 0 | 2.0E-1 |
| Ammonia | 2.0E+3 | 9.8E+1 |
| HCN | 2.0E 0 | 9.8E-2 |
| Aromatic amines | 1.0E 0 | 4.9E-2 |
| Nitrosamines | 5.0E-1 | 2.5E-2 |
| PAHs | 5.0E-1 | 2.5E-2 |
| Fatty acids | 2.6E+2 | 1.3E+1 |
| Trace elements | 5.0E-1 | 2.5E-2 |

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A-1

APPENDIX A

LIQUEFACTION PROCESS DESCRIPTION AND BASIC PREMISES
USED IN DEVELOPING THE MATERIAL BALANCE

Process Description

Sized coal (+ 1/4" to 2") was gasified in the Lurgi gasifiers at 450 psig using steam and oxygen to yield a raw synthesis gas (syngas) product. The raw syngas exited the gasifiers at 900° F. This gas was cooled and scrubbed to remove the attendant particulates, tars, oils, phenols, and other impurities. The impurities were separated from the gas as a gas liquor in the gas/liquor separator. The partially cleaned syngas was then split into two streams; one stream was further cooled while the other stream underwent shift conversion to adjust the H₂/CO ratio in the gas such that H₂/CO ratio of the combined gas stream was in the desired range for the Fischer-Tropsch reaction. The two streams were then comingled and further purified using the Rectisol process to remove the sour gases (H₂S and CO₂) and naphtha from the syngas.

The cleaned gas from the Rectisol process was fed to the Fischer-Tropsch reactors where it was catalytically converted (over a proprietary iron catalyst) to mainly aliphatic hydrocarbon products. The raw products from the Fischer-Tropsch reactors were further processed and upgraded to yield the plant product slate consisting of SNG, C₃ and C₄ LPG, gasoline and fuel oils by using conventional petroleum refinery processes. Details of the product upgrading section have been omitted from the flowsheet because: 1) it consisted of conventional petroleum refinery-type processes, and 2) the major unknowns regarding environmental and occupational health concerns were associated with the front-end or syngas production section of the process.

The acid gases recovered in the various processing steps were taken to the sulfur recovery step where they were converted to sulfur using the Stretford process. The sulfur produced was proposed for sale as plant product. Tail gas from the Stretford process was incinerated in the plant boiler.

APPENDIX A:

(cont.)

Boiler stack gas cleanup facilities for SO₂ removal were provided to reduce the SO₂ emissions to 0.2 pounds of SO₂ per million Btu of fired heat. In addition, an electrostatic precipitator was included to reduce the particulate matter to 0.1 pound per million Btu of fired heat duty.

The gas liquors recovered in the syngas cleanup steps were collected and processed further to separate the oils, dust, tars, crude phenols, and ammonia from the water. The recovered water underwent additional treatment prior to reuse in the process. Waste products recovered from the gas liquor were disposed of appropriately as indicated below.

Oils and tars were recovered from the waste water by settling and decantation. The oils were processed further in the product upgrading section to yield additional plant products; the tars were incinerated in the plant boiler. Crude phenols were recovered from the waste water by using the Phenosolvan process. The crude phenols were incinerated in the plant boiler. Anhydrous NH₃ was recovered for sale by using the Chemie-Linz/Lurgi process.

The denuded waste water from the ammonia recovery step was biologically treated to reduce its BOD and COD levels by removing the residual organics present in the water. The water was then subjected to a reverse osmosis treatment to reduce its metal salts content prior to reuse in the process. The biosludge (from the biological treatment step) was disposed of in an environmentally acceptable manner such as a landfill operation.

APPENDIX A
(cont.)

TABLE A-1- Coal analysis:

Proximate analysis, as-received basis (wt.%)

| | |
|-----------------|--------------|
| Volatile matter | 33.1 |
| Fixed carbon | 33.8 |
| Ash | 5.1 |
| Moisture | 28.0 |
| Total | <u>100.0</u> |

Ultimate analysis, moisture and ash-free (MAF) basis (wt.%)

| | |
|----------|---------------|
| Carbon | 74.45 |
| Hydrogen | 5.10 |
| Oxygen | 19.25 |
| Nitrogen | 0.75 |
| Sulfur | 0.45 |
| Total | <u>100.00</u> |

Heating value, MAF basis, Btu/lb

| | |
|--------------------|--------|
| High heating value | 12,720 |
| Low heating value | 12,236 |

TABLE A-2- Product slate

| | |
|--------------------------------------|--------|
| SNG ⁸⁶ , MMscf/sd* | 173.3 |
| C ₃ LPG, BPSD** | 1,1707 |
| C ₄ LPG, BPSD | .146 |
| Gasoline, BPSD | 13,580 |
| Diesel fuel, BPSD | 2,307 |
| Heavy fuel oil, BPSD | 622 |
| Mixed alcohols, BPSD | 1,825 |
| Sulfur, TPSD | 61 |
| Anhydrous, NH ₃ , TPSD | 103 |
| Total product, BPSD FOE [†] | 44,950 |

* MMscf/sd = Million standard cubic feet per stream day

**BPSD = Barrels per stream day

†FOE = Fuel oil equivalent at 6×10^6 Btu/barrel fuel oil

B-1

APPENDIX B:

ESTIMATION OF STREAM FLOW RATES AND
CONCENTRATIONS BY INDIVIDUAL
STREAM ARE PRESENTED IN
TABLES B-1 AND B-2

TABLE B-1

Appendix B: Estimated process stream concentration for the risk assessment of Indirect (Lurgi/Fischer-Tropsch) coal liquefaction process

| Component | Stream Number | | | | | | | | | | | |
|-------------------------------------|---------------|---------|--------|------------------|------------------|------------------|------------------|---------|---------|---------|--------|---------|
| | 1 | 2 | 3 | 4 | 5 | 6 | 7 | 8 | 9 | 10 | 11 | 12 |
| FLOWRATE, lb mol/h | | | | | | | | | | | | |
| CO | 28.01 | | | | | | 21459 | 21459 | 21459 | 21459 | 4042 | 17416 |
| CO ₂ | 46.01 | | | | | | 33724 | 33724 | 33724 | 33724 | 6353 | 27370 |
| H ₂ | 2.02 | | | | | | 46332 | 46332 | 46332 | 46332 | 8159 | 35973 |
| CH ₄ | 16.04 | | | | | | 12866 | 12866 | 12866 | 12866 | 2424 | 10442 |
| C ₂ H ₆ | 28.05 | | | | | | 86 | 86 | 86 | 86 | 16 | 69 |
| C ₃ H ₈ | 30.07 | | | | | | 614 | 614 | 614 | 614 | 136 | 498 |
| H ₂ + Inerts | 28.01 | | | | | | 310 | 310 | 310 | 310 | 59 | 252 |
| O ₂ | 32.09 | | | | | | | | | | | |
| H ₂ S | 36.08 | | | | | | 158 | 158 | 158 | 158 | 30 | 128 |
| CS ₂ | 60.07 | | | | | | 2 | 2 | 2 | 2 | 4 | 2 |
| H ₂ O | 18.02 | | | | | | 88401 | 93936 | | | | |
| Others | | | | | | | 94379 | | | | | |
| Total gas flowrate, lb mol/h | 14338 | 56335 | 94379 | 202032 | 207487 | 213551 | 117548 | 21400 | 92150 | | | |
| Temperature, °F | 290 | Amb. | 748 | 900 ^a | 700 ^b | 600 ^c | 400 ^d | 367 | 367 | | | |
| Pressure, psia | 600 | Atm. | 660 | 450 ^e | 445 ^f | 425 ^g | 414 | 414 | 414 | | | |
| Flowrate, lb/h | | | | | | | | | | | | |
| H ₂ O | 648687 | 532390 | 115917 | | | | | | | | | |
| Methanol | | | | | 11567 | | | | 1692727 | 1425476 | 268531 | 3156945 |
| Tar | | | | | | | | | 39504 | 4438 | 815 | 3601 |
| Oil | | | | | | | | | 53235 | 38667 | 7246 | 11221 |
| Naphtha | | | | | | | | | 15675 | 15675 | 2951 | 12722 |
| Crude phenols | | | | | | | | | 16189 | 5234 | 986 | 4248 |
| Hexcaptane | | | | | | | | | 42 | 20 | 6 | 14 |
| Thiophene | | | | | | | | | 36 | 16 | 5 | 11 |
| Amines | | | | | | | | | 9985 | 8029 | 1522 | 6556 |
| HCN | | | | | | | | | 8 | 8 | 2 | 6 |
| Aromatic amines | | | | | | | | | 4 | 4 | 1 | 3 |
| Heteroamines | | | | | | | | | 2 | 2 | 4 | 7 |
| Pha | | | | | | | | | 82 | 82 | 2 | 1 |
| Fatty acids | | | | | | | | | 2799 | 1029 | 194 | 835 |
| H ₂ (CO) _n | | | | | | | | | 5 | 6 | 2 | 4 |
| Minerals | 116959 | 96020 | 20904 | | | | | | | | | |
| Coal | 1569420 | 1271963 | 277067 | | | | | | | | | |
| Sulfur | | | | | | | | | 4106 | 4516 | 2 | 1 |
| Particulates | | | | | | | | | 7 | 8 | | |
| Trace elements | | | | | | | | | | | | |
| Others | | | | | | | | | | | | |
| Total Liquid/mol/h | 2316024 | 1901324 | 416089 | | 11567 | | 135983 | 142288 | 1835015 | 1698657 | 202286 | 1216172 |
| TOTAL STREAM FLOWRATE, lb/h | 2316024 | 1901324 | 416089 | 457707 | 1592827 | 1700710 | 4166482 | 4351081 | 4251081 | 3914406 | 737404 | 3177001 |
| Radioactivity | 2.6 | 2.6 | | | | | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 | 0.1 |

Table B-1
Appendix B (Continued)

| Component | Stream Number | | | | | | | | | | | | | |
|-----------------------------------|---------------|---------|---------|--------|---------------------|-------------------|------|------|--------|--------------------|-------------------|------|-----|-----|
| | 13 | 14 | 15 | 16 | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 24 | 25 | 26 |
| FLOWRATE, lb mol/h | | | | | | | | | | | | | | |
| CO | 3218 | 17416 | 18636 | 18500 | 481 | 13 | | | | | | | | |
| CO ₂ | 9177 | 27370 | 36566 | 2900 | 3102 | 05 | | | | | | | | |
| H ₂ | 11184 | 35973 | 67156 | 46985 | 9337 | 719 | | | | | | | | |
| CH ₄ | 2424 | 10662 | 12866 | 32460 | 14258 | 17055 | | | | | | | | |
| C ₂ H ₆ | 16 | 69 | 86 | 21 | 301 | 108 | | | | | | | | |
| C ₃ H ₈ | 116 | 608 | 614 | 277 | 818 | 665 | | | | | | | | |
| N ₂ Inertia | 59 | 252 | 310 | 309 | 309 | 308 | | | | | | | | |
| H ₂ O | 30 | 128 | 157 | | | | | | | | | | | |
| H ₂ S | <1 | 2 | 2 | | | | | | | | | | | |
| CO ₂ | | | | | | | | | | | | | | |
| H ₂ O | | | | | | | | | | | | | | |
| Others | | | | | | | | | | | | | | |
| Total gas flowrate, lb mol/h | 24225 | 92150 | 116371 | 81453 | 28688 | 18813 | | | | | | | | |
| Temperature, °F | 86 | 86 | 86 | 68 | 100 | 130 | | | | | | | | |
| Pressure, psig | 400 | 400 | 400 | 348 | 288 | 1000 | | | | | | | | |
| FLOWRATE, lb/h | | | | | | | | | | | | | | |
| H ₂ O | 710 | 2474 | 3184 | 200 | 117 | 36 | | | | | | | | |
| Methanol | | | | | | | | | | | | | | |
| Tar | | | | | | | | | | | | | | |
| Oil | | | | | | | | | | | | | | |
| Naphtha | | | | | | | | | | | | | | |
| Crude Phlegma | | | | | | | | | | | | | | |
| Hexoptylene ^d | 2953 | 12722 | 15675 | 15675 | PP | 16056 | | | | | | | | |
| Thiophene ^d | 2 | 6 | 6 | <1 | PP | | | | | | | | | |
| Amalgam | 1 | 3 | 4 | <1 | PP | | | | | | | | | |
| Urea ^e | 15 | 66 | 81 | 17 | | | | | | | | | | |
| Aromatic amines ^o | 1 | 3 | 4 | | | | | | | | | | | |
| Hilroamine ^o | <1 | 2 | 3 | | | | | | | | | | | |
| VIA | <1 | 1 | 2 | | | | | | | | | | | |
| Fatty acids | PP | PP | PP | <1 | PP | <1 | | | | | | | | |
| H ₂ (CO) ₂ | <1 | <1 | 2 | <1 | PP | <1 | | | | | | | | |
| Nitrate | | | | | | | | | | | | | | |
| Coal | | | | | | | | | | | | | | |
| Sulfur | | | | | | | | | | | | | | |
| Particulates | | | | | | | | | | | | | | |
| Trace elements ^o | 1 | 1 | 2 | 1 | 185977 ^m | 8903 ^m | | | | | | | | |
| Others | | | | | | | | | | | | | | |
| Total liquid/solid flowrate, lb/h | 3686 | 15277 | 18963 | 15696 | 186095 | 8940 | 8190 | 1245 | 140918 | 25390 ^m | 7436 ^m | 7436 | 600 | |
| ROYAL STREAM FLOWRATE, lb/h | 509687 | 1976108 | 2485795 | 973482 | 627641 | 315340 | 8190 | 1245 | 140918 | 25626 | 7454 | 762 | | |
| Radioactivity | 0.1 | 0.1 | 0.1 | 0.05 | | | | | | | | | | 2.6 |

Table B-1
Appendix B (Continued)

| Component | Stream Number | | | | | | | | | | | |
|--------------------------------------|-----------------|-----------------|-----------------|-----------------|---------|-------|---------|--------|-------------------|------------------|-------|-------------------|
| | 25 ^k | 26 ^k | 27 ^k | 28 | 29 | 30 | 31 | 32 | 33 | 34 | 35 | 36 |
| FLUORIDE, lb mol/h | | | | | | | | | | | | |
| CO | 28.01 | | | | | | | | | | | |
| CO ₂ | 44.01 | | | 62254 | | | | | | | | |
| H ₂ | 2.02 | | | | | | | | | | | |
| Cl ₂ | 16.06 | | | | | | | | | | | |
| C ₂ H ₄ | 28.05 | | | | | | | | | | | |
| C ₂ H ₆ | 30.07 | | | | | | | | | | | |
| H ₂ + Inerts | 28.01 | 4 | | 129696 | | | | | | | | |
| H ₂ | 32.00 | 1 | | 2227 | | | | | | | | |
| H ₂ S | 36.08 | | | | | | | | | | | |
| COS | 60.07 | | | | | | | | | | | |
| H ₂ O | 18.02 | 1 | | 19 ^h | | | | | | | | |
| Others | | | | | | | | | | | | |
| Total gas flowrate, lb mol/h | | 6 | | 194192 | | | | | | | | |
| Temperature, °F | | Amb. | 60 | 100 | Amb. | 60 | 70 | 70 | 70 | 180 ^c | Amb. | Amb. |
| Pressure, psig | | Amb. | 60 | Atm. | Amb. | 5 | 5 | 5 | 5 | 10 ^c | Amb. | Amb. |
| FLUORIDE, lb/h | | 168 | 1030 | 497676 | 2413000 | 10000 | 1179700 | 878977 | | 199686 | 25500 | 20000 |
| H ₂ O | | | | | | | | | | | | |
| Methanol | | | | | | | | | | | | |
| Tar | | | | | | | | | | | | |
| Oil | | | | | | | | | | | | |
| Naphtha | | | | | | | | | | | | |
| Crude phenyle | | | | | | | | | | | | |
| Benzophane | | | | | | | | | | | | |
| Thiophenes ^d | | | | | | | | | | | | |
| Ammonia | | | | | | | | | | | | |
| H ₂ N ^e | | | | | | | | | | | | |
| Aromatic amines ^g | | | | | | | | | | | | |
| Heteroamines ^h | | | | | | | | | | | | |
| PHA | | | | | | | | | | | | |
| Fatty acids | | | | | | | | | | | | |
| Ni(CO) ₄ | 25 | | 2 | 2 | | | | 4420 | 20950 | 96031 | | 116989 |
| Minerals | 310 | | | | | | | | | 7038 | | 7038 |
| Conl | | | | | | | | | | 66 | | 66 |
| Sulfur | 90 | 2 | | 3 | | | | 50 | 200 | 940 | PP | 1172 |
| Particulates | 4 | <1 | <1 | 2 | | | | | 2577 ^p | | | 2577 ^p |
| Trace elements ^l | | | | | | | | | | | | |
| Others | | | | | | | | | | | | |
| Total liquid/solid flowrate, lb/h | 597 | 3 | 1033 | 497683 | 2413000 | 10000 | 1179700 | 883667 | 23727 | 303561 | 25500 | 147802 |
| TOTAL STREAM FLOWRATE, lb/h | 597 | 165 | 1033 | 6942675 | 2413000 | 10000 | 1179700 | 883667 | 23727 | 303561 | 25500 | 147802 |
| Reactivity ^m | 2.6 | | | 0.05 | | | | | 2.5 | 2.5 | | 2.5 |

B-4

Table B-1

Appendix B (Continued)

| Component | Stream Number | | | | | | | | | |
|------------------------------------|-----------------|-----------------|--------|---------|------------------|------------------|------------------|------|---------|---------|
| | 37 | 38 | 39 | 40 | 41 | 42 | 43 | 44 | 45 | 46 |
| Flowrate, lb mol/h | | | | | | | | | | |
| CO | 28.01 | | | | | | | | | |
| CO ₂ | 44.01 | | | | | | | | | |
| H ₂ | 2.02 | | < 1 | < 1 | | | | | | |
| CH ₄ | 16.04 | | | | | | | | | |
| C ₂ H ₆ | 28.05 | | | | | | | | | |
| C ₃ H ₈ | 30.07 | | | | | | | | | |
| H ₂ + Inerts | 28.01 | | < 1 | < 1 | | | | | | |
| O ₂ | 32.00 | | | | | | | | | |
| H ₂ S | 34.00 | | | | | | | | | |
| CO ₂ | 60.07 | | | | | | | | | |
| H ₂ O | 18.02 | | | | | | | | | |
| Others | | | | | | | | | | |
| Total gas flowrate, lb mol/h | 139579 | 878977 | 98623 | 982625 | 340433 | 216938 | 1992989 | 2347 | 1990642 | |
| Temperature, °F | 70 ^d | 70 ^d | 393 | 393 | 170 ^d | 197 ^d | 160 ^d | 157 | 110 | 95 |
| Pressure, psia | 10 ^d | 10 ^d | 251 | 251 | 251 | 250 | 250 ^d | 100 | 80 | 50 |
| Flowrate, lb/h | | | | | | | | | | |
| Ethanol | | | | | | | | | | |
| Tar | | | | | | | | | | |
| Oil | | | | | | | | | | |
| Naphtha | | | | | | | | | | |
| Crude phenyls | | | | | | | | | | |
| Hexoptylene | | | | | | | | | | |
| Styrene | | | | | | | | | | |
| Acetone | | | | | | | | | | |
| RCB ^e | | | | | | | | | | |
| Aromatic amines | | | | | | | | | | |
| Benzonitrile | | | | | | | | | | |
| PhA | | | | | | | | | | |
| Fatty acids | | | | | | | | | | |
| RI(CO) ₂ | | | | | | | | | | |
| Minerals | | | | | | | | | | |
| Coal | 413 | 2600 | | | | | | | | |
| Solids | | | | | | | | | | |
| Particulates | | | | | | | | | | |
| Trace elements | | | | | | | | | | |
| Others | | | | | | | | | | |
| Total liquid/solids flowrate, lb/h | 140000 | 881627 | 104562 | 1045635 | 306817 | 227719 | 2118759 | 4225 | 51716 | 2024115 |
| TOTAL STREAM FLOWRATE, lb/h | 140000 | 881627 | 104634 | 1045707 | 306817 | 227719 | 2118808 | 4225 | 51716 | 2024237 |
| Radioactivity | | | | | | | | | | |

B-5

Table B-1
Appendix B (Continued)

| Component | Stream Number | | | | | | | | | | |
|--------------------------------------|---------------|---------|------|---------|----------------|-----------------|------------------|-----------------|---------|---------|-----------------|
| | 47 | 48 | 49 | 50 | 51 | 52 | 53 ^f | 54 ^f | 55 | 56 | 57 |
| FLOWRATE, lb mol/h | | | | | | | | | | | |
| CO | 28.01 | | | | | | | | 134 | 134 | |
| CO ₂ | 44.01 | | | | | | | | 33646 | 33646 | |
| H ₂ | 2.02 | | | 2 | 314 | | | | 171 | 174 | |
| CH ₄ | 16.06 | | | | | | | | | 411 | |
| C ₂ H ₆ | 28.05 | | | | | | | | | 65 | |
| C ₃ H ₈ | 30.07 | | | | | | | | 337 | 338 | |
| H ₂ + Inerts | 28.01 | | | | 11673 | | | | 1 | 1 | |
| H ₂ | 32.00 | | | | 2704 | | | | 16 | 2 | |
| H ₂ S | 34.08 | | | | pp | | | | 157 | 2 | |
| CO ₂ | 60.07 | | | < 1 | | | | | 2 | 14 | |
| H ₂ O | 18.02 | | | | 2 ^p | | | | | | |
| Others | | | | | | | | | | | |
| Total gas flowrate, lb mol/h | | 3 | | 3 | 14653 | | | | 34527 | 34787 | |
| Temperature, °F | 266 | 95 | | 130 | Amh. | 80 ^d | 70 ^d | 80 ^d | 95 | 95 | 58 |
| Pressure, psia | 130 | 157 | 300 | 50 | Atm. | 10 ^d | 100 ^d | 10 ^d | 12 | 6 | 390 |
| FLOWRATE, lb/h | | | | | | | | | | | |
| H ₂ O | 452 | 1990190 | 40 | 1990190 | 13981 | 1961947 | 27855 | 2813069 | 764 | 41992 | 2420 |
| Methanol | | | | | | | | | 463 | 463 | |
| Tar | | | | | | | | | | | |
| Oil | 200 | | | | | | | | | | |
| naphtha | pp | pp | | pp | pp | pp | pp | pp | | | pp |
| Crude phenols | 14606 | 587 | | 587 | | | | | | | pp |
| Isocaproic acid | pp | pp | | pp | | | | | | | 5 |
| Thiophenes | pp | pp | | pp | | | | | | | 3 |
| Anthracene | pp | pp | | pp | | | | | | | pp |
| HCN | 2 | 9808 | 8521 | 1287 | | | | | 4 | 4 | pp |
| Aromatic amines ^d | 1 | pp | | pp | | | | | | | 3 |
| Nitrosamines ^d | 1 | pp | | pp | | | | | | | 2 |
| PIA | pp | pp | | pp | | | | | | | pp |
| Fatty acids | pp | pp | | pp | | | | | | | pp |
| N(CO) ₂ | | 8268 | | pp | | | | | | | pp |
| Alkanes | | | | pp | | | | | | | pp |
| Coni | | | | | | | 2100 | 500 | | | 3 |
| Sulfur | | | | | | | | | | | |
| Particulates | | | | | | | | | | | |
| Trace elements ^d | | | | | | | | | | | |
| Others | | | | | | | 45 | 5 | | | 61 ^m |
| Total liquid/solid flowrate, lb/h | 15262 | 2008953 | 8561 | 2000252 | 13981 | 1961947 | 59000 | 2813574 | 1237 | 42523 | 2490 |
| TOTAL STREAM FLOWRATE, lb/h | 15262 | 2008975 | 8561 | 2000214 | 440296 | 1961947 | 100000 | 2813574 | 1301999 | 1546631 | 2494 |
| Radioactivity ^h | | | | | | | | | 0.05 | 0.05 | |

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Table B-2

Appendix B: Detailed breakdown of liquid hydrocarbons

| Component | Stream Numbers ^a | | | | | | | | | | | | |
|--------------------------------------|-----------------------------|--------|--------|-------|---------|--------|--------|------|-------|--------|------|------|------|
| | 17 | 18 | 19 | 20 | 21 | 22 | 23 | 54 | 58 | 60 | 64 | 66 | 68 |
| FLUORATE, lb mol/h | | | | | | | | | | | | | |
| C ₂ H ₆ | 30.07 | | 1.06 | | | | | | | | | | |
| C ₃ H ₈ | 42.08 | 189.49 | 17.71 | | | | | | | | | 0.50 | 0.50 |
| C ₃ H ₆ | 44.09 | 18.25 | 168.12 | | | | | | | | 0.18 | 0.07 | 0.05 |
| i C ₄ H ₁₀ | 56.10 | 0.24 | | 1.55 | 4.68 | | | 0.18 | | | | | |
| i C ₄ H ₈ | 56.10 | | | | | | | | | | | | |
| i C ₄ H ₁₀ | 58.12 | 0.06 | 0.06 | | 1.32 | | | | | | 0.83 | | |
| n C ₄ H ₁₀ | 58.12 | 0.03 | 0.03 | 19.93 | 89.05 | | | | | | | | |
| i C ₅ H ₁₂ | 70.13 | 0.04 | 0.04 | | 7.55 | | | | | | | | |
| i C ₅ H ₁₀ | 70.13 | | | | | | | | | | | | |
| i C ₅ H ₁₂ | 72.15 | 9.90 | | | 305.63 | | | | | | | | |
| n C ₅ H ₁₂ | 72.15 | 0.01 | 0.01 | | 55.53 | 0.02 | | 0.07 | | | 0.07 | 0.14 | |
| i C ₆ H ₁₄ | 86.16 | 0.01 | 0.01 | | 173.67 | | | | | | | | |
| i C ₆ H ₁₂ | 86.16 | | | | | | | | | | | | |
| i C ₆ H ₁₄ | 86.17 | 11.17 | 0.01 | | 30.72 | 0.02 | | | | | | | |
| n C ₆ H ₁₄ | 86.17 | 0.01 | 0.01 | | | 0.05 | | | | | | | |
| C ₇ H ₁₆ | 88.17 | 26.08 | | | | | | | | | | | |
| C ₇ H ₁₄ | 88.17 | | | | | | | | | | | | |
| Alkylate | 102.51 | 587.88 | 0.65 | | 332.07 | 129.70 | 22.81 | | | | | | |
| Gas. poly | 118.15 | | | | | | | | | | | | |
| Gasoline | 63.24 | | | | 350.88 | | | | | | | | |
| Acids | 66.07 | | | | | | | | | | | | |
| Ethanol | 64.69 | | | | | | | | | | | | |
| Heavy alcohols | 64.69 | | | | | | | | | | | | |
| Total, lb mol/h | 2370.45 | 208.80 | 186.89 | 21.48 | 1351.10 | 129.79 | 22.81 | 1.08 | 86.48 | 401.53 | 1.08 | 0.71 | 0.55 |
| Total, lb/h | 185977 | 8903 | 8190 | 1245 | 124862 | 25390 | 7436 | 61 | 5469 | 21234 | 61 | 34 | 23 |
| C ₇ ⁺ mol. wt. | 150.6 | 153.24 | | | | 195.70 | 326.00 | | | | | | |

^a Stream numbers indicated correspond to stream numbers in Figure 1.

^b See below for component molecular weights.

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APPENDIX C

EPA SOURCE ANALYSIS MODEL I (SAM/I)*

The SAM/I model has been designed for intermediate screening purposes to evaluate chemical analysis data. To address these objectives, the model includes elementary treatment of pollutant dispersion or dilution to post-dilution levels, but does not incorporate post-dilution chemical reaction or transformation.

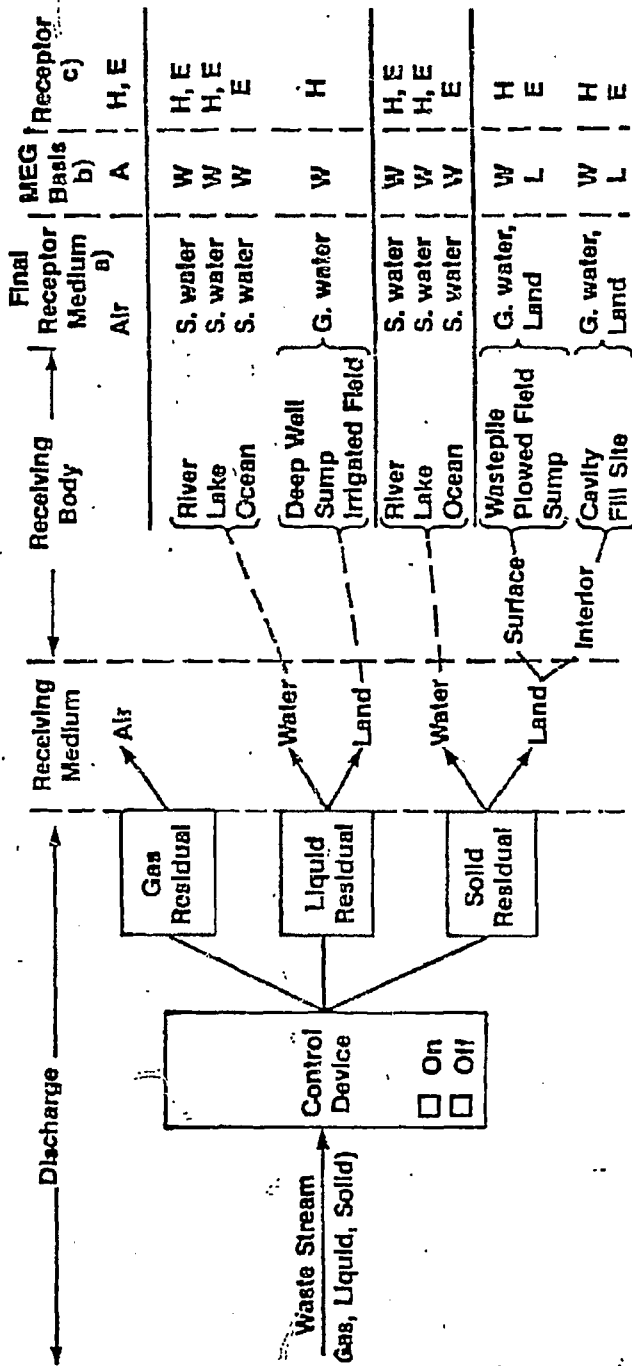
To obtain the estimated maximum post-dilution concentration of a pollutant because of the discharge stream, SAM/I employs approximate dispersion models to account for the dilution of a discharge concentration to a post-dilution concentration. Models have been developed for gaseous, liquid, and solid discharges into appropriate receiving bodies within air, water, and land media. Figure C-1 illustrates the discharge stream/receiving body combinations treated. The figure shows that any given gaseous, liquid, or solid waste stream from a source can be discharged in a number of ways to air, water, or land-receiving media.

Similarly, liquid and solid streams can be discharged to deep well, sump (or waste pond), irrigated field, wastepile, plowed field, cavity, or fill site-receiving bodies in the land medium. The underlying physical picture for all the SAM/I dispersion models is that of a discharge stream entering an entraining post-dilution flow. After mixing takes place, the pollutant stream dispersion, or dilution factor can be approximated by the ratio of the entraining stream volumetric flow rate** to the discharge stream flow rate. SAM/I defines a discharge stream dilution factor, K, in just such a manner:

$$K = \frac{\text{Entraining stream volumetric flow rate}}{\text{Discharge stream volumetric flow rate}}$$

*Information in this Appendix is excerpted from reference 112.

**Entraining stream volumetric flow rate includes the discharge stream volumetric flow rate, i.e. it is the total volumetric flow rate of the two streams after they are mixed.



a) S. water: Surface water, G. water: Ground water
 b) A: Air, W: Water, L: Land
 c) H: Health, E: Ecological

Figure C-1 SAM/I pollutant discharge overview.

Therefore, the estimated maximum* post-dilution concentration for a pollutant species is the ratio of the discharge concentration to the dilution factor.

Dilution factors have been defined for all the receiving bodies shown in Figure C-1. In the dispersion models used to calculate these dilution factors, entraining flow characteristics and certain discharge stream characteristics have been internally parameterized based on estimates of nationwide averages of these characteristics. Thus, only discharge stream flow rate remains a model variable. Further, several model discharge stream flow rates have been defined, spanning discharge flow rate range of interest. Typical dilution factors have been assigned to each of these model streams. Therefore, the SAM/I user need only know the discharge rate of the stream under evaluation, and receiving body discharges into, to perform SAM/I calculations.

As an example, for gaseous effluent streams discharges into the atmosphere, a Gaussian plume dispersion model was used to predict maximum ground level pollutant concentrations. Here, the entraining flow is the atmosphere. The entraining flow characteristics, atmospheric stability, and wind speed are given values within the model typical of nationwide average conditions. Further, discharge stream stack height is internally parameterized by relating average stack height to average source flow rate (e.g., large utility power plants, sources with flue gas discharge rates in the Mg/s range have stack heights around 200m, whereas small commercial or industrial boilers, with flue gas flow rates in the kg/s range have about 10 m stacks). Thus, for SAM/I evaluations a user need only know discharge flow rate to be able to assign an approximate dilution factor.

The defined SAM/I dilution factors, as a function of effluent stream discharge rate, for the various effluent stream/receiving bodies is summarized in Table C-1. Details of the models used to assign these dilution factors are reported elsewhere.

*Maximum under worst probable conditions; it does not consider extreme worst conditions.

Table 33 - Flowrates of Trace Elements in Stream 36

| Trace Element | Q_{TE-A}^* | $+ Q_{TE-S}^{**}$ | $+ Q_{TE-34}^\dagger$ | $- [C_{TE-38}^{\dagger\dagger} (F_{38})^{\S}] =$ | | Q_{TE-36} |
|---------------|------------------|-------------------|-----------------------|--|------------|--------------|
| | (lb/hr) | (lb/hr) | (lb/hr) | (ppm) | (MM lb/hr) | (lb/hr) |
| As | .183-.385 | .0015-.0031 | .2099-.4419 | .02 | .882 | .377-.812 |
| Be | .373-.421 | .0026-.0030 | .3210-.3617 | ND | .882 | <.697-.786 |
| Cd | .094-.242 | .0009-.0025 | .2203-.5684 | .00064 | .882 | .315-.812 |
| F | 5.424-5.590 | 11.157-11.500 | 50.10-51.64 | ND | .882 | <66.68-68.73 |
| Hg | .0074-.0114 | .0000-.0001 | .0782-.1208 | .0003 | .882 | .085-.132 |
| Pb | .180-4.240 | .0015-.0343 | .6577-15.48 | .09 | .882 | .760-19.67 |
| B | ND ^{§§} | ND | 39.43 | ND | .882 | >39.4 |
| Ni | ND | ND | 2.313-19.05 | .035 | .882 | >19.02 |
| V | ND | ND | 13.676-19.147 | ND | .882 | >19.147 |

*From Table 28

**From Table 29

†From Table 31

†† These data are first column volume leachate fractions of Montana Rosebud subbituminous ash from the Lurgi gasifier in Westfield, Scotland. Montana Rosebud coal has a trace element composition very similar to the trace element composition of Wyoming subbituminous. These data were published in Shriner et al., July 1979. It was assumed that the concentrations of trace elements in the sluice water would be the same as the concentrations of trace elements in the first column volume leachate.

§ Appendix B, Stream 38.

§§ No Data

TABLE C-1
SUMMARY OF MODEL STREAM DISCHARGE RATES AND DILUTION FACTORS

| Discharge Stream Type | Gas | | Liquid/Soluble Solid | | | |
|---|-----------------------|-----------------------|----------------------|-----------------------|-----------------------|--------------------------------|
| | Air | | River/Lake | | Ocean | |
| Discharge Rate Q (g/s) and Dilution Factor K | Q | K | Q | K | Q | K |
| | | 2.5 x 10 ⁶ | 1 x 10 ² | 1 x 10 ⁵ | 1.6 x 10 ² | 3 x 10 ⁴ (Barge) |
| | 6.5 x 10 ⁵ | 3 x 10 ² | 1 x 10 ⁴ | 1.6 x 10 ³ | | |
| | 1.3 x 10 ⁵ | 1 x 10 ³ | 1 x 10 ³ | 1.6 x 10 ⁴ | | |
| | 6.8 x 10 ³ | 5 x 10 ³ | 1 x 10 ² | 1.6 x 10 ⁵ | | |
| | 5.4 x 10 ² | 2 x 10 ⁴ | 1 x 10 ¹ | 1.6 x 10 ⁶ | | |

| Discharge Stream Type | Liquid | | Liquid/Soluble Solid | | | | Leached Solid | |
|---|-----------|-----|----------------------|---|-----|---------------|--------------------------------------|-----|
| | Deep Well | | Irrigated Field | Sump, Waste Pile, Plowed Field, Cavity, Fill Site | | Any Land Body | | |
| Discharge Rate Q (g/s) and Dilution Factor K | Q | K | Q | K | Q | K | Q | K |
| | | Any | 1 | Any | 100 | Any | 10 [*] 100 ^{**} | Any |

* Large receiving body with base diameter $d \geq 10m$.

** Large receiving body with base diameter $d < 10m$.