7. Diesohol

7.1 Background

Diesohol is a fuel containing alcohol that comprises a blend of diesel fuel (84.5%), hydrated ethanol (15%) and an Australian developed emulsifier (0.5%). Hydrated ethanol is ethyl alcohol that contains approximately 5% water. The emulsifier is an important component in the preparation of the fuel. It has been developed in Australia by APACE Research.

Development and use of alcohol fuels in transport have, for the most part, been driven by the desire in many countries to find substitutes for imported petroleum based fuels. Alcohol fuels have also been used as additives to conventional fuels to improve fuel characteristics. More recently they have been the focus of attention as a possible means of reducing greenhouse gas emissions and noxious urban emissions from transport.

Anhydrous ethanol will readily blend with petrol. Hydrated ethanol containing more than 2% v/v water is not completely miscible with petrol. Hydrated ethanol is not miscible with diesel but can form an emulsion using a suitable emulsifier. Alcohols can be used in diesel engines by either modifying the fuel or by extensive engine adaptations. Work in Australia by APACE Research Ltd has produced an ethanol and diesel emulsion called 'diesohol'. APACE claims that a diesohol emulsion containing up to 30 per cent ethanol will run in a diesel engine, with the engine requiring little or no modification. The ACTION bus fleet in Canberra trialed three new buses running on diesohol (Scott et al., 1995; Joseph, 1996). Sydney Buses also used such buses from 1993 to 1998 (Figure 7.1).



Figure 7.1 Diesohol bus used by Sydney Buses from 1993 to 1998.

7.2 Characteristics of Diesohol

Table 7.1 lists some of the physical properties of diesohol prepared from regular diesel and from low sulfur diesel.

	,	, ,
Fuel parameter	Regular	Low Sulfur Diesohol
	Diesohol	
Sulfur (ppm)	1000	300
Density at 15°C (kg/m ³)	846.5	846.5
Distillation T95 (°C)	336.4	330.7
Calculated cetane index	52	52
Ash & suspended solids (ppm)	100	100
Viscosity (mm ² /s)	3.568	3.256
Water content (mg/L)	8860	10551
Hydrogen content (mass %)	13.7	12.8
Carbon content (mass %)	86.2	87.2

 Table 7.1

 Diesohol fuel quality specifications (APACE Research Ltd, 1999)

The lower calorific value of ethanol is 20.6 MJ/L (25.6MJ/kg), which drops to 19.41 MJ/L (23.96 MJ/kg) when hydrated 5% v/v with water. Thus the lower calorific value of a blend of 15% hydrated ethanol with diesel (which has a lower calorific value of 35.70 MJ/L or 42.75 MJ/kg) is 33.26 MJ/L. According to APACE Research Ltd. the thermodynamic cycle is affected by the extended ignition delay due to the alcohol. This tends to increase the thermal efficiency, especially under full load conditions. The power reduction is thus less than calculated from calorific values alone. For example, use of 15% v/v ethanol emulsion is calculated to result in a 7.3% reduction in power. However, a reduction of only 3-4 % is usually obtained in practice.

7.3 Production and Distribution

Because ethanol comprises only 15% of diesohol, this section briefly reviews the upstream processes associated with ethanol production. Greater detail is given in the previous examination of ethanol as a fuel in its own right.

7.3.1 Ethanol production

At present there are only two sources of ethanol in Australia. It is manufactured from biomass via the fermentation of sugar that is derived either from wheat starch or from molasses. Starch and sugar crops in Australia that have received attention as a potential source of ethanol include cassava in Queensland; sugarcane in Queensland and northern NSW; sweet sorghum in Queensland, NSW and Victoria; Jerusalem artichokes and potatoes in Victoria; sugar beet in Victoria and Tasmania; and cereals in NSW and Victoria. In Sweden, much of the ethanol used as a fuel comes from excess European wine production (Ericson and Odehn, 1999).

7.3.2 Ethanol from sugar

Ethanol has traditionally been produced in Australia from molasses, a by-product of the sugarcane industry. CSR supplies around half of the Australian ethanol market with an annual plant capacity of 55 million litres (www.csr.com.au/about/Facts_Distilling.htm).

Production of ethanol from molasses constitutes part of the sugar refining process. The overall process consists of the following main steps:

- 1. **Crushing** : Sugar cane "as farmed" is chopped at the sugar mill to facilitate handling and processing.
- 2. **Sugar extraction** : This is effected in a countercurrent flow of warm water. The solids after extraction (bagasse) containing less than 0.5% sugar are squeeze-dried to remove maximum of sugar solution (liquor). Dry bagasse is used as fuel to power sugar mill operation.
- 3. **Raw sugar production** : Sugar-containing liquor is concentrated in evaporators. Crystalline sugar is separated in centrifuges. This process is repeated several times yielding raw sugar. It may be further refined if necessary.
- 4. **Fermentation of molasses** : Liquid residue from sugar production (molasses) containing approximately 50% sugar and 50% mineral matter is mixed with yeast and fermented yielding 6 to 7% ethanol. Solid residue after fermentation (dunder) contains mostly yeast and minerals and is used as fertiliser. Yeast is sometimes separated and used by the food industry.
- 5. **Distillation**: The fermented mash, now called "beer," contains about 10% alcohol, as well as all the non-fermentable solids from the sugar and the yeast cells. The mash is pumped to the continuous flow, multi-column distillation system where the alcohol is removed from the solids and the water. The alcohol leaves the top of the final column at about 96% strength, and the residue mash, called stillage, is transferred from the base of the column to the co-product processing area.
- 6. **Dehydration**: The alcohol from the top of the column then passes through a dehydration system where the remaining water is removed. Most ethanol plants use a molecular sieve to capture the last drop of water in the ethanol. The alcohol product at this stage is called anhydrous (pure, without water) ethanol and is approximately 200 proof.
- 7. **Denaturing**: Ethanol that will be used for fuel is then denatured with a small amount (0-5%) of some product, such as gasoline, to make it unfit for human consumption.

APACE Research (R. Reeves, pers, comm.) notes that molasses is the residue from the production of crystal sugar for food. As residue it has a lower (though non-zero) economic value than the primary output. In the case of CSR's azeotropic ethanol-from-molasses plant at Sarina in Queensland, the processing energy input is supplied from combustion of the sugar cane bagasse. Surplus bagasse is also used by CSR for electrical power cogeneration.

7.3.3 Ethanol from starch

Ethanol is also produced from wheat at Manildra's gluten and starch plant at Nowra, Figure 7.2. The major products of the mill are gluten and starch. The ethanol produced from the waste starch stream with further supplementations of starch is essentially a by-product of the gluten manufacturing process.



Figure 7.2 The ethanol plant at Manildra's Nowra plant. (http://www.manildra.com.au/prospectus/prospectus6.html)

There are basically eight steps in the ethanol production process from wheat starch:

- 1. **Milling**: The wheat (or corn, barley, etc.) first passes through hammer mills, which grind it into a fine powder called meal.
- 2. **Liquefaction**: The meal is then mixed with water and alpha-amylase, and passes through cookers where the starch is liquefied. Heat is applied at this stage to enable liquefaction. Cookers with a high temperature stage (120-150°C) and a lower temperature holding-period (90°C) are used. These high temperatures reduce bacteria levels in the mash.
- 3. **Saccharification**: The mash from the cookers is then cooled and the secondary enzyme (gluco-amylase) added to convert the liquefied starch to fermentable sugars (dextrose), a process called saccharification.
- 4. **Fermentation**: Yeast is then added to the mash to ferment the sugars to ethanol and carbon dioxide. This carbon dioxide, being completely renewable in origin, is not included in the calculations. Using a continuous process, the fermenting mash flows, or cascades, through several fermenters until the mash is fully fermented and then leaves the final tank. In a batch fermentation process, the mash stays in one fermenter for about 48 hours before the distillation process is started.
- 5. **Distillation**: The fermented mash, now called "beer," contains about 10% alcohol, as well as all the non-fermentable solids from the wheat and the yeast cells. The mash is then pumped to the continuous flow, multi-column distillation system where the alcohol is removed from the solids and the water. The alcohol leaves the top of the final column at about 96% strength, and the residue mash, called stillage, is transferred from the base of the column to the co-product processing area.
- 6. **Dehydration**: The alcohol from the top of the column then passes through a dehydration system where the remaining water is removed. Most ethanol plants use a molecular sieve to capture the last drop of water in the ethanol. The alcohol product at this stage is called anhydrous (pure, without water) ethanol and is approximately 200 proof.
- 7. **Denaturing**: Ethanol for fuel is then denatured with a small amount (0-5%) of some product, such as gasoline, to make it unfit for human consumption.
- 8. **Co-Products**: There are two main co-products created in the production of ethanol: carbon dioxide and distillers grain. Carbon dioxide is given off in great quantities during

fermentation and many ethanol plants collect that carbon dioxide, clean it of any residual alcohol, compress it and sell it for use to carbonate beverages or in the flash freezing of meat. This carbon dioxide, also being completely renewable in origin, is not included in the calculations. Distillers grains, wet and dried, are high in protein and other nutrients and are a highly valued livestock feed ingredient. Some ethanol plants also create a "syrup" containing some of the solids that can be a separate production sold in addition to the distiller's grain, or combined with it. Manildra uses this process to produce fructose.

Energy and emission data for ethanol production are available from a number of sources including a NREL study (Kadam et al., 1999) and from Swedish data published on the BioAlcohol Fuels Website (Bioalcohol Fuel Foundation, 2000). These data sources look at different processes (from acid to enzyme) and different feedstocks including woodwaste and straw. However, given the nature of diesohol as a proprietary fuel blend being produced at a specific plant from specific feedstocks, data on ethanol for diesohol production has been taken from documents and personal communications with APACE Research (R. Reeves, pers. comm.). They point out that modern, integrated ethanol-from starch plants, such as that of Manildra, have a processing energy input of approximately 4.5 MJ/L of azeotropic ethanol, and 5.9 MJ/L of anhydrous ethanol. Based on a lower calorific value of 19.43 MJ/L for azeotropic ethanol and 21.15 MJ/L for anhydrous ethanol, and assuming natural gas to steam conversion efficiency of 70%, Reeves estimates the processing energy input to be 0.33 of the lower calorific value for ethanol for azeotropic ethanol, and 0.40 for anhydrous ethanol (as described in Appendix 6).

No individual process data is available for the Manildra process so it has been modelled as a black box with waste product and coal based heat into the plant, with ethanol as the main output. The ethanol was assumed to be azeotropic so the energy use of ethanol production was 9 MJ/L (as in Table 6.10 in the chapter on hydrated ethanol).

There are no solid residues available for combustion from Manildra's ethanol-from-starch plant. All liquid effluent streams, principally the underflow from the stripping distillation column, are irrigated onto surrounding land for intensive pasture production. Thus the liquid effluent has displaced use of conventional fertilisers and significantly increased the soil carbon content. Given that the source of carbon is from renewable sources, no credit for fixing fossil carbon is given from a greenhouse perspective. For the same reason carbon dioxide emissions from fermentation are not included as greenhouse impacts as they are from short-term carbon cycles.

Without clear estimates of the nutrient replacement achieved through land application of effluents, and evidence of this lowering fertiliser use, it is not possible to provided credits for avoided fertiliser use. The effect of these credits is thought to be small in any case.

It is assumed that the starch feedstock used by Manildra for ethanol production is waste starch from Manildra's gluten production, or is derived from reject grain. Because of the low value of these feedstocks, they are treated as waste products and not as by-products of the starch process, and thus have no environmental burdens associated with them. If the value of these feedstocks increase, or higher grade grain is used in the Manildra plant, then an alternative allocation will be needed to include environmental burdens of the feedstock. Modelling of ethanol, as for fuels other than diesohol, included in the next stage of the report, will include allocation procedures for production from dedicated feedstocks and valuable by-products.

Emulsifier for diesohol

According to APACE research the emulsifier that allows the ethanol and the diesel to blend consists of a styrene-butadiene copolymer which is dissolved in the diesel fuel, and a polyethyleneoxide-polystyrene (PEOPS) copolymer which is dissolved in the hydrated alcohol. No values are known as to the proportions of these substances so a total emulsifier is assumed to consist of 50% of each co-polymer. The co-polymers are then also assumed to

Component	Amount Assumed	Inventory data source
Sytrene	250 g	Steinhage (1990) modified with Australian feedstocks
Butadiene	250 g	Steinhage (1990) & Reinders (1983) modified with Australian
		feedstocks
Polyethylene Oxide	250 g	Grant (1999) as polyethylene
Polystyrene	250 g	Same as for Styrene with polymerisation data from Steinhage (1990)

consist of 50% of the two polymer constituents. The resultant mixture for the emulsifier is show in Table 7.2

Table 7.2

Data summary for 1kg of emulsifier used in diesohol

7.4 **Diesohol Emissions**

7.4.1 Upstream

Hydrated (or azeotropic) ethanol derived from sugar, or ethanol derived from wheat starch, may be used for production of diesohol. Hydrated ethanol production is a one-stage refining process, unlike the two-stage anhydrous ethanol. However, from the viewpoint of the LCA, the upstream emissions for ethanol production will be different for both processes.

There are two reasons for this. First, there are differences in energy demand for both processes. Second, as in both cases ethanol is a co-produced with other value added products, there will be differences in emissions allocation as per ISO 14040.

In the past, the ethanol used for diesohol came from the Manildra refinery. The calculations in this report are based on the present source of ethanol for diesohol, namely the CSR refinery at Sarina.

7.4.2 Tailpipe

APACE Research (Ernie Lom, pers. comm.) provided results from Swedish tests of diesohol conducted in 1997¹. These results (Table 7.3) are for fuels that blend diesohol with Swedish Diesel fuel and with European diesel (EDsl) meeting EN590 specifications. European diesel is a low-sulfur fuel. Swedish diesel is an ultra-low sulfur fuel.

tis of 5 wearsh tests of a		in low s	unui (Ei	b) and t	1010	ow sulful (SwL	
Fuel	CO ₂	СО	NOx	HC	PM	Fuel Use	
SwD	205.6	0.4	1.9	0.2	0.047	98.81	
SwD+OXC	211.1	0.4	1.9	0.2	0.033	101.97	
SwDhol	200.0	0.5	2.0	0.2	0.030	99.72	
SwDhol+OXC	200.0	0.4	2.0	0.2	0.017	98.39	

0.4

0.5

1.9

2.0

0.2

0.2

0.061

0.039

97.42

100.11

205.6

205.6

Table 7.3 Results of Swedish tests of diesohol (g/MJ) with low sulfur (ED) and ultra-low sulfur (SwD) fuels

OXC = Oxidation catalyst

EDsl

EDhol

¹ Westerholm, R., Christensen, A., Tornqvist, M., Ehrenberg, L. & Haupt, D. (1997) Chemical and biological characterisation of exhaust emissions from ethanol and ethanol blended diesel fuels in comparison with neat diesel fuels, KFB Report 1997:17, Kommunikations Forsknings Beredningen (Swedish Transport and Communications Research Board) Stockholm.

The ACTION bus fleet in Canberra trialed three new buses running on diesohol (Scott et al., 1995; Joseph, 1996). Sydney buses also used such buses, until 1998, from their Burwood depot (Figure 7.1) and the results of emission testing of these buses is given in APACE Research Ltd (1999).

The tests on diesohol that were conducted by the NSW EPA (Scott et al., 1995) compared the performance of three ACTION ethanol-fuelled buses with three buses fuelled by diesel. The results are given in g/kWh. They have been converted to g/MJ and to g/km on the basis of the observed fuel consumption, which ranged from 217 to 341 g/kWh, and on the fuel economy, which ranged from 36.79 L/100 km to 46.96 L/100 km. The density for all fuels was assumed to be 840 g/L. The results are summarised in Table 7.4, Table 7.5 and Table 7.6. We have analysed the results presented on both the Canberra and Sydney buses

 Table 7.4

 Results of testing of Canberra buses (Scott et al., 1995; Joseph, 1996)

CO ₂	СО	NOx	HC	Fuel Use
296	0.47	4.25	0.25	101
296	0.39	4.81	0.25	95
981	1.57	14.09	0.83	
963	1.27	15.66	0.81	
	CO ₂ 296 296 981 963	CO2 CO 296 0.47 296 0.39 981 1.57 963 1.27	CO2 CO NOx 296 0.47 4.25 296 0.39 4.81 981 1.57 14.09 963 1.27 15.66	CO2CONOxHC2960.474.250.252960.394.810.259811.5714.090.839631.2715.660.81

Results of resultg of byuncy buses (AT ACE Research T ty Erd, 1777)						
Fuel	CO ₂	СО	NO _x	HC	PM	Fuel Use
Diesel (g/MJ)	212	0.22	1.98	0.14	0.05	60.8
E15 (g/MJ)	212	0.25	1.88	0.15	0.04	65.6
E17 (g/MJ)	210	0.22	1.97	0.14	0.04	65.0
E20 (g/MJ)	213	0.25	1.87	0.15	0.03	66.4
LSD (g/MJ)	207	0.22	2.08	0.17	0.04	60.3
LSDiesohol(E15) (g/MJ)	206	0.26	1.97	0.16	0.03	65.6
Diesel (g/km)	1310	1.37	12.20	0.89	0.28	
E15 (g/km)	1274	1.49	11.33	0.89	0.24	
E17 (g/km)	1274	1.31	11.95	0.84	0.21	
E20 (g/km)	1263	1.49	11.09	0.91	0.20	
LSD (g/km)	1291	1.39	12.97	1.04	0.25	
LSDiesohol(E15) (g/km)	1242	1.57	11.85	0.99	0.19	

 Table 7.5

 Results of testing of Sydney buses (APACE Research Pty Ltd, 1999)

Fuel	Formaldehyde	Acetaldehyde.
Diesel (g/MJ)	0.0014	0.0033
E15 (g/MJ)	0.0011	0.0061
E17 (g/MJ)	0.0017	0.0061
E20 (g/MJ)	0.0014	0.0067
LSD (g/MJ)	0.0022	0.0061
LSDiesohol(E15) (g/MJ)	0.0019	0.0058
Diesel (g/km)	0.0086	0.0206
E15 (g/km)	0.0064	0.0350
E17 (g/km)	0.0096	0.0353
E20 (g/km)	0.0079	0.0377
LSD (g/km)	0.0139	0.0381
LSDiesohol(E15) (g/km)	0.0111	0.0334

 Table 7.6

 Results of testing of Sydney buses for air toxics (APACE Research Pty Ltd.)

In addition, Scott et al. (1995: Table 25) present a summary of the aldehyde emissions from Canberra buses using diesohol, in concentration units. At a speed of 50 km/h under 25% load, the formaldehyde emissions are as given in Table 7.7.

 Table 7.7

 Concentrations of aldehydes emitted from Canberra buses at 50 km/hr under 25% load.

Fuel	Formaldehyde	Acetaldehyde	Acrolein	Total Aldehydes
Diesohol (ppmv)	0.658	1.667	0.483	2.792
Diesel (ppmv)	0.783	1.342	0.85	2.958

7.5 Full Fuel-Cycle Analysis of Emissions

APACE Research (R. Reeves, personal communication) provided estimates of the life-cycle carbon dioxide emissions of diesohol using the energy balance method of Lynd (1996). These calculations are reproduced in Appendix 6. They claim the following emissions:

- 80 gCO₂/MJ for diesel fuel
- 28 gCO₂/MJ for Manildra azeotropic ethanol
- 31 gCO₂/MJ for Manildra anhydrous ethanol
- 4 gCO₂/MJ for CSR azeotropic ethanol from molasses
- 16 gCO₂/MJ for ethanol from dedicated lignocellulosic crops
- 6 gCO₂/MJ for ethanol from lignocellulosic residue material.

These values may be compared with those calculated by Beer et al. (2000) who estimated lifecycle CO_2 emissions to be 80 gCO₂/MJ for diesel fuel and 36 gCO₂/MJ for ethanol from lignocellulose. We believe that discrepancy between this latter value, obtained using a bottom-up approach and the 16gCO₂/MJ estimated by APACE, using a top-down approach are indicative of the range of uncertainty associated with estimates of full fuel cycle greenhouse gas emissions.

Despite the energy savings associated with lignocellulosic ethanol, there is no commercial production of such ethanol in Australia. Even though the buses that were tested in the diesohol tests used diesohol with the ethanol made from wheat starch waste, our calculations are based on an expected supply of ethanol from molasses from Sarina.

7.5.1 Emissions on a mass per unit energy basis

The results obtained by using the SimaPro life-cycle model along with the upstream and tailpipe emissions data specified in the previous chapters of this report are given in Table 7.8 for the full life cycle for greenhouse gases and criteria pollutants. The upstream emissions and the tailpipe emissions that comprise these totals are given in Table 7.9 and Table 7.10 respectively. The greenhouse gas emissions are graphed in Figure 7.3.

Full Lifecycle	Units	LS Diesel	Diesohol
Greenhouse	kg CO ₂	0.0858	0.0800
HC total	g HC	0.140	0.133
HC urban	g HC	0.111	0.106
NOx total	g NOx	1.044	0.966
NOx urban	g NOx	0.987	0.912
CO total	g CO	0.253	0.335
CO urban	g CO	0.242	0.325
PM10 total	mg PM10	40.7	31.8
PM10 urban	mg PM10	39.3	30.5
Energy Embodied	MJ LHV	1.18	1.11

 Table 7.8

 Urban and total life-cycle emissions (per MJ) calculated for diesel and diesohol

The results separate urban and total emissions. Emissions were assumed to occur in urban areas unless they were produced by a known rural or maritime activity.

The apparent discrepancies in certain values, when compared with tabulations earlier in this report, arise because many of the values that are reported in the main text are in terms of g/MJ measured as usable energy from the engine driveshaft (normally represented as g/kWh), whereas the life-cycle calculations are consistent in setting all the calculations in terms of g/MJ based on the inherent chemical energy of the fuel. On average, this reduces quoted engine dynamometer values by a factor of 3.



Figure 7.3 Life-cycle emissions of greenhouse gas emissions for low sulfur diesel, diesohol and ethanol.

Precombustion	Units	LS Diesel (Aus)	Diesohol
Greenhouse	kg CO ₂	0.0191	0.0214
HC total	g HC	0.0565	0.0532
HC urban	g HC	0.027	0.026
NOx total	g NOx	0.100	0.103
NOx urban	g NOx	0.043	0.049
CO total	g CO	0.023	0.075
CO urban	g CO	0.012	0.065
PM10 total	mg PM10	5.42	4.97
PM10 urban	mg PM10	4	3.63
Energy Embodied	MJ LHV	1.18	1.11

 Table 7.9

 Urban and total upstream emissions (per MJ) for diesel and diesohol

		LS Diesel	
Combustion	Units	(Aus)	Diesohol
Greenhouse	kg CO ₂	0.0667	0.059
HC total	g HC	0.0835	0.080
HC urban	g HC	0.0835	0.080
NOx total	g NOx	0.944	0.863
NOx urban	g NOx	0.944	0.863
CO total	g CO	0.230	0.260
CO urban	g CO	0.230	0.260
PM10 total	mg PM10	35.3	26.82
PM10 urban	mg PM10	35.3	26.82
Energy Embodied	MJ LHV	0	0

 Table 7.10

 Urban and total tailpipe emissions (per MJ) from diesel and diesohol

7.5.2 Vehicle emissions - trucks (g/km)

This section gives the calculated values for the emissions from trucks, on a per-kilometre basis.

 Table 7.11

 Urban and total life cycle emissions (per km) for trucks calculated for diesel and diesohol

Full LC	Units	LS Diesel engine	Diesohol engine
Greenhouse	kg CO ₂	0.9250	0.8619
HC total	g HC	1.509	1.430
HC urban	g HC	1.192	1.141
NOx total	g NOx	11.250	10.402
NOx urban	g NOx	10.638	9.823
CO total	g CO	2.723	3.606
CO urban	g CO	2.612	3.501
PM10 total	mg PM10	438.4	342.3
PM10 urban	mg PM10	423.1	327.9
Energy Embodied	MJ LHV	12.7	11.90

Precombustion		LS Diesel (Aus)	Diesohol
Greenhouse	kg CO ₂	0.2060	0.2310
HC total	g HC	0.609	0.573
HC urban	g HC	0.292	0.284
NOx total	g NOx	1.080	1.110
NOx urban	g NOx	0.468	0.531
CO total	g CO	0.243	0.805
CO urban	g CO	0.132	0.700
PM10 total	mg PM10	58.4	53.5
PM10 urban	mg PM10	43.1	39.1
Energy Embodied	MJ LHV	12.7	11.9

 Table 7.12

 Urban and total precombustion emissions (per km) for trucks calculated for diesel and diesohol

Table 7.13
Urban and total tailpipe emissions (per km) for trucks calculated for diesel and diesohol

Combustion		LS Diesel (Aus)	Diesohol
Greenhouse	kg CO ₂	0.719	0.631
HC total	g HC	0.900	0.857
HC urban	g HC	0.900	0.857
NOx total	g NOx	10.17	9.292
NOx urban	g NOx	10.17	9.292
CO total	g CO	2.48	2.801
CO urban	g CO	2.48	2.801
PM10 total	mg PM10	380	288.80
PM10 urban	mg PM10	380	288.80
Energy Embodied	MJ LHV	0	0

7.5.3 Vehicle emissions - buses (g/km)

This section gives the calculated values for the emissions from buses, on a per-kilometre basis. The greenhouse gas emissions and the particulate matter emissions are graphed in Figure 7.3.

Full LC		LS Diesel	Diesohol
Greenhouse	kg CO ₂	1.66	1.55
HC total	g HC	2.71	2.57
HC urban	g HC	2.14	2.05
NOx total	g NOx	20.20	18.68
NOx urban	g NOx	19.10	17.64
CO total	g CO	4.89	6.48
CO urban	g CO	4.69	6.29
PM10 total	mg PM10	787	614.62
PM10 urban	mg PM10	760	588.77
Energy Embodied	MJ LHV	22.8	21.37

 Table 7.14

 Urban and total life-cycle emissions for buses (per km) calculated for diesel and diesohol



Figure 7.4 Exbodied greenhouse gases emissions (kg CO₂-eq) from diesohol production, processing and use in vehicle



Figure 7.5 Exbodied particulate matter (mg - urban) from diesohol production, processing and use in vehicle

Precombustion		LS Diesel (Aus)	Diesohol
Greenhouse	kg CO ₂	0.37	0.41
HC total	g HC	1.09	1.03
HC urban	g HC	0.52	0.51
NOx total	g NOx	1.94	1.99
NOx urban	g NOx	0.84	0.95
CO total	g CO	0.44	1.45
CO urban	g CO	0.24	1.26
PM10 total	mg PM10	104.9	96.06
PM10 urban	mg PM10	77.4	70.21
Energy Embodied	MJ LHV	22.80	21.37

 Table 7.15

 Urban and total precombustion emissions for buses (per km) calculated for diesel and diesohol

Combustion		LS Diesel (Aus)	Diesohol
Greenhouse	kg CO ₂	1.29	1.133
HC total	g HC	1.62	1.538
HC urban	g HC	1.62	1.538
NOx total	g NOx	18.26	16.684
NOx urban	g NOx	18.26	16.684
CO total	g CO	4.45	5.030
CO urban	g CO	4.45	5.030
PM10 total	mg PM10	682.31	518.56
PM10 urban	mg PM10	682.31	518.56
Energy Embodied	MJ LHV	0.00	0

 Table 7.16

 Urban and total tailpipe emissions for buses (per km) calculated for diesel and diesohol

7.5.4 Uncertainties

We use the uncertainty estimates given by Beer et al. (2000) on the basis of the tailpipe emissions to estimate the uncertainties associated with the above results to be as given in Table 7.17.

 Table 7.17

 Estimated one standard deviation uncertainties (in percent) for diesohol emissions

	g/MJ	g/t-km	g/p-km
CO ₂	10	10	10
HC	45	17	73
NOx	17	26	8
CO	51	36	66
PM10	45	45	45

7.6 Viability and Functionality

The flash point of the emulsion becomes that of alcohol when the alcohol content exceeds 5% of the volume. Above a 15% ethanol blend an ignition improver is needed, whereas above 25% ethanol engine modifications are required.

Two problems have been found to date with the use of diesohol according to discussions with Mr Ernie Lom and Dr Russell Reeves of APACE Research Ltd. The first of these is comparable to those with the use of low sulfur diesel, and relate to fuel injection equipment components. The components are: i) some T valves fitted to Bosch type feed pumps swell excessively and result in the valve stem becoming jammed; ii) the drive shaft seal fitted to Nipon Denso rotary pumps can swell and soften resulting in fuel leakage; iii) some filter glues, impregnation resins and epoxy resins (such as in DPA pump and RBA transfer pump blades) are susceptible and need to be identified in service.

The second problem, which has been fixed with the installation of booster pumps, concerns the need to ensure that vapour locks do not occur. Adding ethanol changes the vaporization potential of diesel. Evaporative emissions of VOC from vehicles increase when vapour pressure of the fuel is increased or the ambient temp rises (Carnovale et al., 1991). Diesel fuel has a very low vapour pressure but the addition of alcohol to diesel in diesohol creates a fuel with a similar vapour pressure to ethanol. While modern gasoline vehicles have evaporative

emissions control measures, diesel vehicles do not. Evaporative emissions may be a significant problem from unmodified vehicles using diesohol, but this needs to be tested. To control evaporative emissions from vehicles using alcohol fuels, measures may need to be implemented to control fuel vapour pressure, and control evaporative emissions from diesel fuel vehicles. APACE Research Ltd ensures that there are no vapour locks by installing a booster pump (E. Lom, pers. comm.).

They also point out that diesohol was the only emulsified fuel to pass stability test conducted by Shell. To date diesohol has been a niche fuel and thus the situation with respect to availability and warranty has not been clarified. During testing of buses using diesohol, the fuel was blended by delivering diesel to Manildra, near Nowra, and blending the diesel with ethanol and emulsifier.

7.7 Health and OHS

7.7.1 *Production and transport*

The ethanol used in Australia is manufactured from biomass from the fermentation of sugar derived from grain or sugar crops. Production of these feedstock crops results in a range of particles and air toxic emissions.

Feedstock transport to the ethanol production facility results in a range of particles and air toxic emissions. These will be detailed in subsequent work that deals solely with ethanol. In this review of diesohol, these contributions are noted by the difference in value between the last two columns of Table 7.9. This approach is taken because we were specifically asked to compare each fuel (diesohol in this case) against LSD as the reference fuel. Similarly, emissions of particulate matter and air toxics could be expected from the ethanol production process. The process includes high temperature cooking and fermentation, which emits acetaldehyde.

As the composition of diesohol is 85% diesel the production and transport emissions associated with diesohol production are assumed to be similar to LSD, except for the ethanol and emulsfier component. The emulsifier consists of a styrene-butadiene copolymer dissolved in the diesel fuel that, by steric stabilisation, couples with a polyethyleneoxide-polystyrene (PEOPS) copolymer dissolved in the hydrated alcohol. Manufacture of the emulsifier involves butadiene, which is an air toxic. However, the quantities of emulsifier are small (0.5% v/v) compared to the quantities of diesel and ethanol. Consequently the amount of butadiene is very low.

Particulate matter

The urban precombustion (truck) PM10 estimate for LSD is 43 mg/km compared to 39 mg/km for diesohol (Table 7.12).

Air toxics

The urban precombustion (truck) HC estimate for LSD is 0.292 g/km compared to 0.284 g/km for diesohol (Table 7.12). The public health effects of air toxics will be mainly associated with combustion emissions in the large urban centres.

An accompanying disk to this report from provides details of air toxics emissions from upstream activities.

7.7.2 Use

APACE Research results, as summarised in Table 7.5, indicate that compared to LSD emissions, diesohol (E15) emissions have marginally higher CO emissions, but marginally lower NOx and HC emissions.

Particulate matter

The APACE Research results (Table 7.5) indicate that, compared to LSD emissions, diesohol (E15) emissions have lower PM emissions. The values are 0.04 g/MJ (0.25 g/km) for LSD and 0.03 g/MJ (0.19g/km) for diesohol made from low sulfur diesel. The combustion (truck) PM10 estimate from the LCA for LSD is 380 mg/km compared to 289 mg/km for diesohol.

Air toxics

The APACE Research results (Table 7.6) also indicate that compared to LSD emissions, diesohol (E15) emissions have marginally lower acetaldehyde emissions – 0.038 g/km for LSD compared to 0.033 g/km for LSDiesohol (i.e. diesohol made with low sulfur diesel). However, the Swedish Euro2 bus study found emissions of 0.02 g/km acetaldehyde (Ahlvik & Brandberg, 2000) using low sulfur diesel. This provides a measure of the variability in the data and hence the uncertainty in the results.

There are lower formaldehyde emissions using diesohol. Low sulfur diesel emits from 0.014 g/km, whereas LSDiesohol emits 0.011 g/km. Table 7.7 also indicates that acrolein emissions will be lower with diesohol than with diesel fuels.

Information for diesohol was not available for the other air toxics. However the diesohol HC emissions were marginally lower compared to LSD for the APACE Research results. The combustion (truck) HC (assumed to be equivalent to NMVOC) estimate for LSD is 0.900 g/km compared to 0.857 g/km for diesohol.

7.7.3 Diesohol emissions summary

As the composition of diesohol is 85% diesel the production and transport emissions associated with diesohol production are assumed to be similar to LSD. The LCA indicates that urban precombustion PM10 emissions of diesohol (39 mg/km or 3.63 mg/MJ) are marginally lower than LSD (43 mg/km or 4.0 mg/MJ), though the urban precombustion HC emissions are similar at 0.29 g/km or 0.026 g/MJ.

The LCA indicates that combustion PM emissions from diesohol (289 mg/km or 26.8 mg/MJ) are lower than LSD (380 mg/km or 35.3 mg/MJ).

There is limited information available on air toxic emissions for diesohol. The high proportion of diesel in diesohol suggests that the air toxic emissions are unlikely to be substantially different to LSD. The LCA indicates that HC combustion emissions of diesohol are similar to LSD

7.7.4 OHS Issues

The flash point and flammability characteristics of diesohol are those of alcohol. This requires that diesohol be considered and handled as gasoline (petrol) rather than as diesel fuel, even though the flash point of petrol is considerably lower than that of ethanol (13°C). In practical terms, APACE Research handles the fuel as it would ethanol to ensure safety. Ethanol in solution is hazardous according to Worksafe Australia, with high flammability, moderate toxicity, and a moderate irritant.

Occupation exposure of drivers to diesohol vapours during HDV refuelling was assessed by Workcover in 1992 (NSW Workcover Authority, 1999). Normally refuelling is conducted by

keeping the fuel dispensing nozzle in the automatic mode with only the last 10-12 litres added manually. The drivers are normally only exposed to diesohol vapours during manual refuelling. The results indicate that levels of diesohol vapours are low and do not represent a significant health hazard to drivers.

7.8 Environmental Issues

The present use of ethanol, as in diesohol, is that of a niche fuel. As such, there are no issues related to sustainability. However, if ethanol were to become a dominant fuel then it would have to be based on ligno-cellulose. Foran and Mardon (1999) contains details of ethanol and methanol production technology and supply constraints, and of the environmental consequences of both crop and fuel production processes. They claim that if ligno-cellulosic ethanol production is used then it would be possible to establish biomass plantations over the next 50 years that meet 90% of Australia's oil requirements, and specifically to supply all transportation fuels. To do this using ethanol requires biomass production to cover up to 19 million hectares of Australia's croplands and high rainfall pasture zones. Their modelling approach envisages substantial environmental benefit. In addition to the reduction in greenhouse gas emissions (up to 300 million tonnes by the year 2050), the large-scale planting of tree and shrub crops as ethanol feedstock would help to control dryland salinity and associated problems.

The environmental impact from the production of diesohol are the same as those from the production of the diesohol feedstocks; namely diesel as ethanol, and will be dealt with in the relevant chapters.

In particular, we draw attention to the fact that appropriate disposal of the refinery wasteproducts is crucial to environmental impacts or benefits. Dunder application is often criticised as being the cause of poor waste quality in Queensland, though there is little evidence of this (www.sunfish.org.au/fishkills/fishkills.htm). Conversely, appropriate and careful disposal of dunder means that many farmers in the district near Sarina now use it as a fertiliser and soil condition - even though it was once considered a poison.

We are not aware of any issues related to groundwater contamination.

7.9 **Expected Future Emissions**

Arcoumanis (2000) developed a model that examines a given alternative fuel relative to the reference diesel engine (Euro2) in terms of a specific regulated pollutant. A value of 1 implies identical performance to the low sulfur diesel/Euro2 combination. A value greater than 1 implies inferior performance, whereas a value less than 1 indicates superior performance.

Table 7.18 lists the estimated emissions factors for diesohol. The columns in bold represent the standards relative to the Euro2 standard. The adjacent column gives the expected performance of diesohol. The estimates of Arcoumanis (2000) indicate that diesohol can be expected to meet all future Australian Design Rules for all pollutants except total hydrocarbon which may be slightly above Euro3 and Euro4 standards.

Estimated relative emission factors for diesohol under future technologies. Euro2 diesel (shown in bold) are taken as 1.0										
Technology	со	СО	THC	THC	NOx	NOx	РМ	PM	CO ₂	LCA CO ₂
Euro2	1.0	1.1	1.0	1.1	1.0	0.8	1.0	0.6	1.0	0.9
Euro3	0.53	0.5	0.6	0.7	0.71	0.6	0.67	0.4	1.0	0.9
Euro4	0.38	0.4	0.42	0.5	0.5	0.4	0.2	0.1	1.0	0.9

Table 7.18

APACE Research advises that vapour lock problems had led to higher THC and CO emissions as reflected in Arcoumanis (2000). APACE has indicated that the addition of a booster pump now overcomes vapour lock problems and the resulting THC and CO problems. This means that LSdiesohol should be able to meet future ADRs.

7.10 Summary

7.10.1 Advantages

- As a partly renewable fuel it produces less fossil CO₂ than conventional fuels.
- Particulate emissions are lowered.
- 1,3 butadiene and benzene levels decrease as the ethanol concentration increases.
- Lower sulfur content than conventional diesel.

7.10.2 Disadvantages

• Overseas, the chemical emulsifiers used to blend ethanol and diesel contain harmful chemicals. According to APACE the chemical emulsifier that they use is composed only of hydrocarbons and oxygen and is thus no more harmful than diesel fuel.

7.11 Appendix to Diesohol Fuel Chapter

This appendix (Appendix 6) comprises a separate file of scanned material provided by APACE.

This page left blank intentionally

8. Compressed Natural Gas

8.1 Background

Natural gas (NG) is a mixture of hydrocarbons, mainly methane (CH₄), and is produced either from gas wells or in conjunction with crude oil production. The composition of natural gas used in Melbourne in 1997/98 was 91.6 percent methane, 5.0 percent ethane, 0.4 percent propane, 0.1 percent butane, 0.8 percent nitrogen and oxygen, and 2.1 percent carbon dioxide. Natural gas is consumed in the residential, commercial, industrial, and utility markets.

The interest for natural gas as an alternative fuel stems mainly from its clean burning qualities, its domestic resource base, and its commercial availability to end-users. Because of the gaseous nature of this fuel, it is stored onboard a vehicle in a compressed gaseous state (CNG), though it is also possible to liquefy it and store it in liquid form (LNG).

In Australia, CNG is compressed to around 25 MPa for on-board storage at typically 20 MPa. Refuelling of CNG vehicles is done in the following way. Natural gas is drawn from the distribution network, compressed to 25 MPa and stored in pressure vessels. When a vehicle is being filled and pressure in the storage vessel drops, the compressor draws further gas from the pipeline. The storage vessels are used only to speed up the filling process, not to hold large quantities of compressed gas. In some cases, for example 'slow-fill' refuellers, the pressure vessel stage is bypassed and the compressor compresses gas directly into the cylinder of the vehicle.

8.1.1 Natural gas production

Natural gas consumed in Australia is domestically produced. Gas streams produced from reservoirs contain natural gas, liquids and other materials. Processing is required to separate the gas from petroleum liquids and to remove contaminants. First, the gas is separated from free liquids such as crude oil, hydrocarbon condensate, water, and entrained solids. The separated gas is further processed to meet specified requirements. For example, natural gas for transmission companies must generally meet certain pipeline quality specifications with respect to water content, hydrocarbon dewpoint, heating value, and hydrogen-sulfide content. A dehydration plant controls water content; a gas processing plant removes certain hydrocarbon components to hydrocarbon dewpoint specifications; and a gas sweetening plant removes hydrogen sulfide and other sulfur compounds (if present). As raw natural gas is odourless, a chemical odorant (generally sulfur in the form of a mercaptan) is generally added prior to entering the local distribution system to enable expeditious identification of any gas leaks, although some gas is transmitted without odorant.

8.1.2 Natural gas market

Natural gas is distributed throughout Australia in pipeline systems (Figure 5.1) that extend from the well-head to the end user.



Figure 8.1 Australian gas fields and pipelines

Every mainland State and Territory has access to natural gas through pipelines. The pipeline system consists of long-distance transmission systems, followed by local distribution systems. Some underground storage is also used to help supply seasonal peak needs.

The Australasian Natural Gas Vehicles Council web site in their submission for this study point out that:

Known world reserves of natural gas now constitute over 95% of equivalent oil reserves. In Australia this ratio is more than three times the oil reserve. Proven Australian resources of natural gas currently stand at 109,051 PJ, at existing production levels, this will last 91 years compared to domestic oil reserves which are estimated to last 39 years. Historically, reticulated gas with most of the infrastructure being below ground, has survived most disasters including two world wars. Natural gas is abundant, clean, readily accessible and strategically independent of traditional oil based fuels. The last point is important in that NG is not controlled by a small number of global corporations and cartels and is not reliant on USD exchange rates or world parity pricing.

8.1.3 Fuel characteristics

Natural gas has very different fuel characteristics from the fuels normally used in internal combustion engines. Its density, at 0.70 g/L is lighter than air. Louis (2001) cites a lower heating value of 52.9 MJ/kg.

The energy content (higher heating value) of CNG varies from 38.8 megajoules per cubic metre at atmospheric pressure in New South Wales and South Australia to 38.5 in Victoria, 37.5 in Western Australia and 41.9 in the Northern Territory (National Greenhouse Inventory Committee, 1998)). The average energy content is similar to that of one litre of automotive diesel oil (38.6 megajoules), and about 12 per cent above that of one litre of gasoline (34.2 megajoules) (ABARE, 1991). Pressurised storage of a cubic metre of natural gas as CNG, however, requires a container volume of 4 to 5 litres.

A national fuel standard for CNG is to be developed in 2001-2002 under the *Fuel Quality Standards Act 2000.*

8.1.4 Implications for engine conversions

Because of its characteristics, natural gas can be used in spark ignition engines, but in compression ignition engines a proportion of diesel fuel is usually required to trigger ignition. Alternatively, diesel engines can be converted to spark ignition for natural gas use.

For diesel engines (primarily HDVs in Australia), the conversion to a compression ignition dual (mixed) fuel configuration involves use of a pilot supply of diesel to ignite the natural gas. This requires the addition of a gas fuel system alongside the existing diesel fuel system, together with a mechanism for regulating the proportion of diesel and gas for the engine speed and load conditions. According to the IEA (1993) engine efficiency for this configuration is about the same as that for a diesel engine. BTCE (1994) states that the efficiency of dual (mixed) fuel systems can be equal to or higher than for diesel at high loads, but lower at part loads. For this reason, the overall efficiency in service is lower than for diesel. This chapter deals with single fuel vehicles so that dual fuel vehicles have not been examined. It is to be expected, based on results of LPG dual fuel vehicles, that emissions reductions from dual fuel vehicles will not be as large as those from single fuel vehicles.

Conversion of diesel engines to spark ignition engines running solely on natural gas requires more extensive modification, in that the diesel fuel injectors in the cylinder head will be replaced by spark plugs, and an ignition system added to the engine. A compression ratio lower than that of the diesel is likely to be required. Also, a larger cylinder capacity than that required for a dual (mixed) fuel system may be needed, to provide the same energy content. Though conversions have been the primary source of natural gas engines in Australia to date, increasing availability of OEM engines and vehicles makes conversions less relevant.

8.2 Full Fuel Cycle

Nigge (2000) recently undertook a detailed life cycle assessment of natural gas vehicles in Germany that quantified emissions and health effects.

8.2.1 Tailpipe

The Australasian Natural Gas Vehicle Council (ANGVC) kindly provided emissions data from the latest generation of engines taken from various studies including UK test data on a Scania CNG 113M engine using Mobil CNG (Table 8.1), data from Cummins on their 8.3 litre diesel and C8.3G engine with and without catalyst (Lyford-Pike, 2001) and data from a 9.8 L Transcom

modified Renault 620-45 natural gas engine (AEC Limited), as well as data from South Australian CNG buses (ANGVC, 2001).

	Table 8.1 Scania diesel and CNG test results (g/kWh) in the UK (Andrew, 2001)				
	НС	СО	NOx	PM	CO ₂
Diesel	0.864	1.442	7.014	0.3731	756.3
CNG	0.212	0.018	0.962	0.007	674
LNG	0.18	0.017	1.532	0.013	698

Table 8.1 provides results of tests of the present generation of diesel engines (Scania DSC 11-21) as tested at the Millbrook Proving Ground in January 2001 (Andrew, 2001). The drive cycle was not specified. However, as the European Community requires Euro3 standards for heavy vehicles as from January 2000, we expect that both the engines and the test regime corresponded to Euro3. The specific fuel consumption during the test of the CNG vehicle was 190 g/kWh at 1100 to 1800 rpm. The minimum range of the CNG truck was 560 km. The truck achieved a range in excess of 640 km by increasing the CNG pressure from 20 MPa to 25 MPa

Table 8.2 provides results obtained in December 2000 by a Renault engine tested under the European Transient Cycle (ETC), and by Cummins engines tested in November 2000 under the US EPA 99/00 requirements. These are equivalent to ADR 80 and to Euro3 requirements.

	NMHC	ТНС	CH ₄	NOx + NMHC	со	NOx	РМ	CO ₂
Transcom modified Renault 620-45 with catalyst	0.003	0.531			0.024	2.432		
Cummins (C8.3G) CNG with catalyst (ULEV)	0.28		6.27	2.33	1.04	2.05	0.01	678
Cummins (C8.3G) CNG without catalyst	1.058		6.54	3.63	8.67	2.57	0.034	695
(LEV)								
Cummins Diesel (ISC280) with catalyst					0.67	5.36	0.07	700
Cummins Diesel (ISC280) without catalyst					1.21	5.36	0.12	753

 Table 8.2

 Emissions results (g/kWh) for Renault and Cummins engines

By contrast, Table 8.3 gives the emission results of tests on a MAN NL 202 bus with a D0826 LUH, 6.87 litre, turbocharged, intercooled engine, and with a D2866 DUH, 11.97 litre natural gas engine. These engines are on buses that are actually in service at present. The tests were done using the ECE R-49 cycle. The diesel engines were tested with diesel fuel (2000 ppm), low sulfur diesel (500 ppm) and with Euro3 diesel (300 ppm sulfur).

¹ This value is unduly large. Our subsequent calculations are based on the LSD value for PM in Table 8.3.

	South Australian bus emissions data (g/kWh)					
	нс	СО	NOx	РМ		
Euro 1 Diesel	0.25	0.97	7.8	0.17		
Euro 2 Diesel (LSD)	0.13	0.48	6.66	0.10		
Euro 3 Diesel	0.04	0.65	4.87	0.08		
CNG	0.2	1	1	0.02		

Table 8.3

One problem with certification procedures based on engine dynamometers is that they may report values that substantially differ from those calculated by chassis dynamometers. The NSW EPA (Brown et al. 1999) also tested Scania 11L Turbo Euro2 technology CNG buses for their performance with, and without, a catalyst. The results are reproduced in Table 8.4.

		ТНС	Methane	NMVOC
Without catalyst	Bus #1	2.86	2.64	0.22
Without catalyst	Bus #2	3.37	2.92	0.45
With catalyst	Bus #1	1.88	1.85	0.03
With catalyst	Bus #2	3.02	2.78	0.24

Table 8.4 Methane and non-methanic hydrocarbon emissions (g/kWh) from CNG buses

Another source of representative data is given in Table 8.5, which reproduces the emission factors (based on emissions per MJ of fuel use) for heavy vehicles fuelled by natural gas that are given by the National Greenhouse Gas Inventory Committee (1998). Using these default figures typical methane emission are 2.5 g/km and the N₂O emissions for a natural gas-fuelled urban bus are 0.0247 g/km.

Emission factors (g/MJ) for heavy vehicles fuelled by natural gas					
Gas Emission factor					
CO,	54.4				
$\tilde{CH_4}$	0.101				
N ₂ O	0.001				
ŇŎx	1.2				
СО	0.2				
NMVOC	0.01				

Table 8.5

We note that the estimate of tailpipe emissions of 1344 g CO₂/km for a CNG bus that Beer et al. (2000) obtained corresponds to a fuel efficiency of 24.7 MJ/km. As a typical energy content for natural gas is 39 MJ/m³ the results of Beer et al. (2000) were based on an assumed fuel economy of 1.58 km/m³. According to NSW State Transit (Hardy, pers. comm. 2000) the known fuel

consumption of the CNG buses is 1.6 km/m^3 . The results of Andrew (2001) that were used in this analysis indicate that the present generation of CNG buses are far more fuel efficient, emitting 595 g CO₂/km, which corresponds to a fuel efficiency of 10.9 MJ/km.

8.2.2 Upstream emissions

As CNG is assumed to be produced from high pressure gas supplies in major cities, standard gas production and transmission processes are used for the upstream emissions of Natural Gas. Added to this are compression processes based on either a gas engine driven CNG compressor, or an electrically driven CNG compressor.

Data on natural gas production have been derived from the National Greenhouse Gas Inventory for 1998 (NGGIC, 2000). This data is presented in Table 8.6.

Table 8.6 Energy use data for oil and gas production and refinery processing						
Fuel Energy Use Production 1998 Energy use to energy production r						
		PJ	PJ	GJ/PJ produced		
Oil and gas production and field	Petroleum	0.9	2528.6	0.36		
	Gas	141.1	2528.6	55.80		
Natural gas transmission	Gas	8.6	688.5	12.49		
Gas production and distribution	Gas	2.4	371.5	6.46		

The compression process involves a simple model with natural gas as energy as the main inputs, and CNG as the main output. The energy use is usually quoted in terms of its efficiency compared with the energy value of the gas being compressed. Data on compression are taken from Wang (1999) and are listed in Table 8.7. The emission data for natural gas combustion for compression is taken from standard natural gas combustion data for industrial boilers presented in NGGIC (2000) for greenhouse emissions and in Environment Australia (1999) for air toxics. This data is presented in Table 8.8. The data for electricity combustion are from the same sources for emissions while fuel usage and grid mix are taken from Electricity Supply Association of Australia (2000). Full fuel cycle inputs are presented in Table 8.9 and FFC emissions are presented in Table 8.10 for an average Australian grid mix.

 Table 8.7

 Energy use in natural gas compression for two fuel scenarios

Fuel	Efficiency	Value in MJ	Comment
Energy from Natural Gas	91.70%	4643	90.5 MJ per 1000MJ Gas (51.3MJ/kg) compressed
Australian Electricity	96.60%	1550	30.2 MJ per 1000MJ Gas (51.3MJ/kg) compressed

Emissions	Value	Unit	Source
CO ₂	51.19	g	NGGIC, 1997 Standard data Table 1
methane	10.41	mg	NGGIC, 1997 Standard data Table 1
N ₂ O	0.12	mg	NGGIC, 1997 Standard data Table 1
NOx	220.59	mg	NGGIC, 1997 Standard data Table 1
СО	42.32	mg	NGGIC, 1997 Standard data Table 1
non methane VOC	3.48	mg	NGGIC, 1997 Standard data Table 1
SOx	0.053	mg	(Environment Australia 1999)
particles	3.078	mg	(Environment Australia 1999)
benzene	0.86	μg	(Environment Australia 1999)
formaldehyde	30.38	μg	(Environment Australia 1999)
n-Hexane	734.18	μg	(Environment Australia 1999)
toluene	1.37	μg	(Environment Australia 1999)
PAHs	0.28	μg	(Environment Australia 1999)
As	0.08	μg	(Environment Australia 1999)
Be	0	μg	(Environment Australia 1999)
Cd	0.46	μg	(Environment Australia 1999)
Chromium	0.56	μg	(Environment Australia 1999)
cobalt	0.03	μg	(Environment Australia 1999)
Copper	0.35	μg	(Environment Australia 1999)
Lead	0.2	μg	(Environment Australia 1999)
manganese	0.15	μg	(Environment Australia 1999)
mercury	0.11	μg	(Environment Australia 1999)
Nickel	0.86	μg	(Environment Australia 1999)
Selenium	0.01	μg	(Environment Australia 1999)
Zn	11.65	μg	(Environment Australia 1999)

 Table 8.8

 Air emissions from combustion of 1 MJ of natural gas for process energy

Note: these figures are not for a full fuel cycle – energy input to supply gas for combustion are shown in Table 8.6.

Resources (Inputs from Nature)		
coal 19.5MJ/kg	7.22	g
coal 22.1MJ/kg	28	g
coal 22.6MJ/kg	40	g
crude oil	210	mg
lignite 14.4MJ/kg	4.81	g
lignite 8.2MJ/kg	108	g
natural gas	40.9	mg
pot. energy hydropower	114	kJ

Table 8.9Fuel inputs for 1 MJ of average Australian electricity

Source: Grant, unpublished data from Life Cycle Inventory Databases

8.2.3 Fugitive emissions

Natural gas can contain significant quantities of naturally occurring CO₂, which in the past has often been vented to the atmosphere at the well-head. Le Cornu (1989) pointed to Cooper Basin gas as having up to 35 per cent by weight (12.7 per cent by volume) of naturally occurring CO₂. On a state by state basis, vented CO₂ accounts for between 3 and 15 per cent of full fuel-cycle CO₂ emissions from natural gas combustion (Wilkenfeld, 1991). In some instances CO₂ recovered from natural gas could be compressed and used in enhanced oil recovery.

Fugitive emissions of methane occur at the wellhead (production), processing, transmission and end user distribution. Our analysis indicates that average emissions at production stage in Australia amount to 2.17 kg per tonne of gas, while processing contributes 5.74 kg per tonne of gas.

Australian long distance high pressure (up to 15 MPa) transmission pipelines are relatively modern (the oldest dates back to 1969) and built to high standards. They are well maintained and accidental leaks are a rarity. It is estimated that at transmission stage fugitive emissions are 0.005% of the total network throughput.

Most gas losses from the distribution systems are by way of leakage from the low pressure network (7 kPa). This includes both the reticulation network and appliances operated by end users. Losses from the distribution network are difficult to estimate as they may occur both upstream and downstream from the meters. It is estimated that emissions from the distribution network, called unaccounted gas, i.e. the difference between the gas issued by the utilities and the gas sold to customers may be as high as 7.5% (NGGIC, 1996). We consider this to be an upper bound to likely fugitive emissions.

Emission	Value	Unit	Emission	Value	Unit
acetaldehyde	54.4	μg	Manganese	82.2	pg
antimony	1.69	μg	Methane	332	mg
As	39.4	μg	Methane(sea)	45.7	μg
В	6.47	μg	methyl ethyl ketone	37.5	μg
Ва	115	Ng	Methyl isobutyl ketone	15	μg
Be	2.06	μg	Methyl methacrylate	1.87	μg
benzene	125	μg	Mg	617	μg
benzene sea	69.7	Pg	Mn	47	μg
benzo(a)pyrene	683	Pg	Мо	152	ng
Bi	985	Pg	N ₂ O	2.77	mg
Carbon disulfide	12.2	μg	naphthalene	267	ng
Cd	4.87	μg	n-Hexane	933	ng
Chloroform	5.62	μg	n-hexane (sea)	59	ng
СО	60.4	Mg	Ni	26.3	μg
CO (sea)	5.75	μg	Nickel	1.07	ng
CO_2	253	G	non methane VOC	7.49	mg
cobalt	16.5	Pg	Non methane VOC (sea)	17.5	μg
Copper	255	Pg	NOx	678	mg
Cr (III)	24.4	μg	NOx (sea)	18.4	μg
Cr (VI)	4.49	μg	o-xylene	9.98	pg
Cu	69.2	Ng	o-xylene (sea)	0.182	pg
cumene	506	Ng	РАН	2.01	μg
CxHy sulfur	4.31	Ng	PAH (sea)	24.7	pg
Cyanide	487	μg	Pb	39.7	μg
cyclohexane	17.2	Ng	pentane	3.13	μg
DEHP	6.94	μg	phenol	1.5	μg
Dibutyl phthalate	5.25	μg	PM10	15.4	mg
Dioxin & Furans	165	Pg	PM10 (sea)	83.1	pg
dust	18.5	Mg	Se	73	μg
ethylbenzene	10.2	μg	Selenium	68.6	pg
ethylbenzene (sea)	0.104	Pg	Soot	57.4	μg
F	14.4	Mg	SOx	1.26	g
formaldehyde	32.5	μg	styrene	2.44	μg
formaldehyde (sea)	2.5	Ng	tetrachloroethylene	4.12	μg
H_2S	39.4	Ng	toluene	29.9	μg
HCl	113	Mg	toluene (sea)	122	pg
hexane	6.37	μg	Trichloroethylene	5.81	μg
Hg	4.73	μg	V	902	ng
Lead	316	Pg	xylenes	3.6	μg
Li	65.5	Pg	Zn	123	ng

 Table 8.10

 Air emissions for 1 MJ of average Australian electricity

The values for fugitive emissions used in this study are based on data on fugitive emission from natural gas production and also from the NGGI for 1998. The values are presented in Table 8.11.

		Fuel Quant	ity CO ₂	CH ₄	N ₂ O	NOx	CO	NMVOC
		(PJ)	(Gg)	(Gg)	(Gg)	(Gg)	(Gg)	(Gg)
Oil	Exploration (for both oil and gas)	1257	14.8	0.2				0.1
Gas	Production and processing	1272		1.6				1
	Transmission	689		4.9				0.1
	Distribution	372	10.4	171.7				25.5
Venting and flaring for Oil and	Venting at Gas processing plant	1272	2814	119.6				42.3
Gas Production	Distributed Venting	860	749					
	Flaring	2646	2188	26.6	0.1	1.1	6.6	11.4

 Table 8.11

 Fugitive greenhouse emission data for oil and gas production and refinery processing

Source: Fugitive Emissions from Fuels 1B-2 (sheet 1): Oil and Natural Gas

A process tree for CNG production is shown in Figure 8.2 with the methane emission shown in grams as the lower value in each process box. The largest fugitive emission is in the assumed loss in fuel distribution, which is discussed in more detail below.

Methane emissions from vehicles

Methane, the principal component of natural gas, has a greenhouse radiative forcing (GWP) of 21 over a 100-year period. It is therefore important that tailpipe losses of unburnt fuel and fugitive/evaporative losses are minimised.

As methane is a non-reactive hydrocarbon, tailpipe emissions of methane are not as well controlled by catalytic converters. According to Nylund and Lawson (2000: p.46) the sulfur based odorant used in natural gas at very low concentration levels can have a very detrimental effect on the conversion efficiency of oxidation catalysts, bringing their methane conversion down to 30%. When catalysts are optimised for methane, then conversion efficiencies can be as high as 85-90%.

Methane fugitive losses in distribution

Fugitive losses would have the potential to reduce substantially any advantages that natural gas may have in terms of emissions. Gas supply authorities considered that fugitive losses would be less than 2 per cent, and concentrated entirely on the old town-gas reticulation systems. Refuelling depots or retail gas reticulation systems would be serviced by new medium or high pressure lines, and fugitive losses from this form of distribution might be expected to be very low. BTCE (1994) point out that fugitive losses may be exaggerated through a lack of understanding of the term 'unaccounted for gas,' which is the overall accounting error including metering over a vast distribution network.



Figure 8.2 Methane emission in grams across CNG life cycle per km truck transport

(Kadam, 1999) assumes emissions from gas processing plants are 0.1% while the 1998 NGGI claims total distribution losses for low pressure gas supply are 0.25%. In the final modelling, a figure of 0.1% has been used for fugitive emission of methane from CNG facilities – including all operations from the point of gas supply to the facility, up to, but not including, the combustion of the gas on board the vehicle. A sensitivity analysis showing the effect of different levels of fugitive emissions is presented in Figure 8.3. It shows that up to 1% emission the greenhouse gas emission results are still lower than the baseline diesel fuel, though at 10% the full fuel cycle emission is substantially above the diesel baseline. The exbodied emissions and the baseline are the same at approximately 4% fugitive emissions.



Figure 8.3 Effect of different fugitive emission assumption of full fuel cycle greenhouse emission per km of truck travelled

Two modes of compression were examined: compression using natural gas and compression using electricity.

8.3 Results

8.3.1 Emission per unit energy

Urban and rural life cycle emissions calculated for diesel and CNG							
Full Lifecycle	Units (per MJ)	LS diesel	CNG (Elec.comp)	CNG (NG comp)			
Greenhouse	kg CO ₂	0.0858	0.0665	0.0683			
NMHC total	g HC	0.140	0.027	0.029			
NMHC urban	g HC	0.111	0.003	0.003			
NOx total	g NOx	1.044	0.140	0.152			
NOx urban	g NOx	0.987	0.126	0.137			
CO total	g CO	0.253	0.011	0.014			
CO urban	g CO	0.242	0.005	0.008			
PM10 total	mg PM10	40.7	1.1	1.2			
PM10 urban	mg PM10	39.3	0.9	1.0			
Energy Embodied	MJ LHV	1.18	1.09	1.15			

 Table 8.12

 Urban and rural life cycle emissions calculated for diesel and CNG

Table 8.13
Urban and rural precombustion emissions per MJ for CNG

Precombustion	Units	LS diesel	CNG	CNG
			(Elec.comp)	(NG comp)
Greenhouse	kg CO ₂	0.0191	0.0117	0.0135
NMHC total	g HC	0.0565	0.0248	0.0273
NMHC urban	g HC	0.027	0.001	0.001
NOx total	g NOx	0.100	0.026	0.038
NOx urban	g NOx	0.043	0.013	0.023
CO total	g CO	0.023	0.007	0.011
CO urban	g CO	0.012	0.001	0.004
PM10 total	Mg PM10	5.42	0.439	0.526
PM10 urban	Mg PM10	4	0.257	0.327
Energy Embodied	MJ LHV	1.18	1.09	1.15

Combustion	Units	LS diesel	CNG (Elec.comp)	CNG (NG comp)
Greenhouse	kg CO ₂	0.067	0.054	0.054
NMHC total	g HC	0.084	0.019	0.019
NMHC urban	g HC	0.084	0.019	0.019
NOx total	g Nox	0.944	0.114	0.114
NOx urban	g Nox	0.944	0.114	0.114
CO total	g CO	0.230	0.003	0.003
CO urban	g CO	0.230	0.003	0.003
PM10 total	mg PM10	35.26	0.7	0.7
PM10 urban	mg PM10	35.26	0.7	0.7
Energy Embodied	MJ LHV	0	0	0

 Table 8.14

 Urban and rural combustion emissions per MJ for CNG

 Table 8.15

 Summary of life cycle emissions per MJ from CNG

		LS diesel	CNG (Elec.comp)	CNG (NG comp)
Greenhouse	Precombustion	0.0191	0.0117	0.0135
Greenhouse	Combustion	0.0667	0.0548	0.0548
NMHC total	Precombustion	0.0565	0.0248	0.0273
NMHC total	Combustion	0.0835	0.0019	0.0019
NMHC urban	Precombustion	0.0271	0.0007	0.0010
NMHC urban	Combustion	0.0835	0.0019	0.0019
NOx total	Precombustion	0.1000	0.0262	0.0384
NOx total	Combustion	0.944	0.114	0.114
NOx urban	Precombustion	0.043	0.013	0.023
NOx urban	Combustion	0.944	0.114	0.114
CO total	Precombustion	0.0225	0.0072	0.0108
CO total	Combustion	0.2301	0.0034	0.0034
CO urban	Precombustion	0.0123	0.0014	0.0045
CO urban	Combustion	0.2301	0.0034	0.0034
PM10 total	Precombustion	5.42	0.44	0.53
PM10 total	Combustion	35.26	0.66	0.66
PM10 urban	Precombustion	4.00	0.26	0.33
PM10 urban	Combustion	35.26	0.66	0.66
Energy Embodied	Precombustion	1.18	1.09	1.15

8.3.2 Emissions per unit distance

Urban and rural life cycle emissions per km calculated for diesel, CNG					
Full Lifecycle	Units (per km)	LS diesel	CNG (Elec.comp)	CNG (NG comp)	
Greenhouse	kg CO ₂	0.9250	0.7284	0.7474	
NMHC total	g HC	1.509	0.293	0.320	
NMHC urban	g HC	1.192	0.028	0.032	
NOx total	g NOx	11.250	1.533	1.666	
NOx urban	g NOx	10.638	1.383	1.502	
CO total	g CO	2.723	0.116	0.155	
CO urban	g CO	2.612	0.052	0.086	
PM10 total	mg PM10	438.4	12.0	12.9	
PM10 urban	mg PM10	423.1	10.0	10.7	
Energy Embodied	MJ LHV	12.7	11.90	12.50	

Table 8.16

_

Table 8.17 Urban and rural precombustion emissions per km for diesel and CNG LS diesel CNG CNG Precombustion Units (per km) (Elec.comp) (NG comp) Greenhouse 0.2060 0.1290 0.1480 kg CO₂ NMHC total g HC 0.609 0.272 0.299 NMHC urban g HC 0.292 0.007 0.011 1.080 0.287 0.420 NOx total g NOx NOx urban g NOx 0.468 0.137 0.256 CO total g CO 0.243 0.079 0.118 CO urban g CO 0.132 0.015 0.049 PM10 total 4.81 5.76 mg PM10 58.4PM10 urban mg PM10 43.1 2.81 3.58 Energy Embodied MJ LHV 12.7 12.5 11.9

Combustion	Units	LS diesel	CNG (Elec.comp)	CNG (NG comp)
Greenhouse	kg CO ₂	0.719	0.595	0.595
NMHC total	g HC	0.900	0.212	0.212
NMHC urban	g HC	0.900	0.212	0.212
NOx total	g NOx	10.177	1.246	1.246
NOx urban	g NOx	10.177	1.246	1.246
CO total	g CO	2.480	0.037	0.037
CO urban	g CO	2.480	0.037	0.037
PM10 total	mg PM10	380.00	7.2	7.2
PM10 urban	mg PM10	380.00	7.2	7.2
Energy Embodied	MJ LHV	0	0	0

 Table 8.18

 Urban and rural combustion emissions per km for diesel, CNG

 Table 8.19

 Summary of life cycle emissions per km for diesel, CNG

		LS diesel	CNG (Elec.comp)	CNG (NG comp)
Greenhouse	Precombustion	0.2060	0.1290	0.1480
Greenhouse	Combustion	0.7190	0.5994	0.5994
NMHC total	Precombustion	0.6090	0.2720	0.2990
NMHC total	Combustion	0.9000	0.0212	0.0212
NMHC urban	Precombustion	0.2920	0.0072	0.0108
NMHC urban	Combustion	0.9000	0.0212	0.0212
NOx total	Precombustion	1.0800	0.2870	0.4200
NOx total	Combustion	10.170	1.246	1.246
NOx urban	Precombustion	0.468	0.137	0.256
NOx urban	Combustion	10.170	1.246	1.246
CO total	Precombustion	0.2430	0.0788	0.1180
CO total	Combustion	2.4800	0.0368	0.0368
CO urban	Precombustion	0.1320	0.0154	0.0488
CO urban	Combustion	2.4800	0.0368	0.0368
PM10 total	Precombustion	58.40	4.81	5.76
PM10 total	Combustion	380.00	7.17	7.17
PM10 urban	Precombustion	43.10	2.81	3.58
PM10 urban	Combustion	380.00	7.17	7.17
Energy Embodied	Precombustion	12.70	11.90	12.50



Figure 8.4 Exbodied greenhouse gases from CNG production and use with electrical compression



Figure 8.5 Exbodied particulate matter from CNG production and use with electrical compression



Figure 8.6 Exbodied greenhouse gases from CNG production and use with natural gas compression



Figure 8.7 Exbodied particulate matter from CNG production and use with natural gas compression

8.3.3 Uncertainties

We use the uncertainty estimates given by Beer et al. (2000) on the basis of the tailpipe emissions to estimate the uncertainties associated with the above results to be as given in Table 8.20.

	g/MJ	g/t-km	g/p-km
CO ₂	10	2	12
NMĤC	135	135	135
NOx	50	29	72
CO	15	11	22
PM10	60	17	108

 Table 8.20

 Estimated one standard deviation uncertainties (in percent) for CNG emissions

8.3.4 Discussion

Our results indicate lower greenhouse gas emissions both from tailpipe emissions and from upstream emissions. Earlier studies, such as those reported in the IPCC Second Assessment Report (Watson et al., 1996), the Expert Reference Group (1998) report, or those mentioned at <u>http://www.hsph.harvard.edu/Organizations/hcra/diesel/diesel.pdf</u> obtain different results. There are two reasons for this – changes in vehicle technologies, and the expected fugitive emissions.

Changes in vehicle technologies

The lower vehicle emissions arise from the improved performance of the present series of dedicated CNG engines that are optimised for the use of CNG. Earlier studies were based on a previous generation of CNG engines. This is evident when the history of the Western Australian experience is examined. The Expert Reference Group (1998) report examined issues associated with diesel and natural gas fuels and decided that diesel was the preferred fuel. The ANGVC (2000) responded with a review of the report and discussed what it believed to be the inadequacies of the report.

Following the Western Australian election, the decision to purchase diesel buses was reversed and natural gas buses were ordered. The firm Advanced Engine Components Ltd. was contracted to install its multipoint sequential electronic fuel injection natural gas vehicle system on Daimler-Chrysler M447G engines. The system was tested in June 2001 at the Swiss Federal Laboratories for Materials Testing and Research (EMPA) in Zurich under the official European Transient Cycle. The engine was certified as being compliant with the Euro4 standard. The results of the tests done in June 2001, shown in Table 8.21, demonstrate that the present generation of NGV vehicles perform at Euro4 specifications.

Table 8.21
Emissions (g/kWh) from Daimler-Chrysler M447G engines

Technology	СО	тнс	CH ₄	NMHC	NOx	PM	CO ₂	Specific Fuel Consumption
G20 Fuel Gas ²	0.131	0.167	0.156	0.011	3.09	0.006	626	185-216
G25 Fuel Gas	0.134	0.479	0.459	0.02	2.88	0.007	637	185-216
Euro3 standard	5.45	2.38	1.6	0.78	5.0	0.16		
Euro4 standard	4.0	1.65	1.1	0.55	3.5	0.03		

² EU reference fuel: G20 is 100% methane, G25 is 86% methane.

Fugitive emissions

The reduction in upstream emissions occurs because we assumed for Australia, on the basis of the advice received from stakeholders, that fugitive emissions are 0.1% of supply. This leads to the results, tabulated above, that exbodied emissions of greenhouse gases are less than that of diesel. Earlier studies and overseas studies, based on assumptions of higher fugitive emissions, produce opposite results in relation to greenhouse gases. We undertook a sensitivity study, as depicted in Figure 8.3, that indicates that if fugitive emissions exceed 4 % of supply then exbodied emissions of greenhouse gases exceed those of low sulfur diesel.

8.4 Viability and functionality

8.4.1 Safety

According to the IANGV web site (<u>www.iangv.org/sources/ga.html</u>) natural gas vehicles (NGV) have an excellent safety record (especially when compared to petrol driven vehicles). They cite two fundamental reasons for this: the structural integrity of the NGV fuel system and the physical qualities of natural gas as a fuel.

The fuel storage cylinders used in NGVs are much stronger than petrol tanks. The design of NGV cylinders are subjected to a number of specified "severe abuse" tests, such as heat and pressure extremes, gunfire, collisions and fires.

Though fuel storage cylinders are stronger than petrol tanks, when composite material used to encase the tanks, the materials are fundamentally more susceptible to physical damage than metals under abusive conditions. For this reason, composite materials on NGV cylinders must always be properly handled and protected. Incidents involving natural gas cylinder ruptures revealed that some form of chemical attack or physical damage to the composite overwrap on the cylinder was involved. This has been addressed in new cylinder standards by prescribing a standard acid exposure test.

NGV fuel systems are "sealed", which prevent any spills or evaporative losses. Even if a leak were to occur in an NGV fuel system, the natural gas would dissipate into the atmosphere because it is lighter than air.

Natural gas has a high ignition temperature, about 650°C, compared with about 350°C for gasoline. It also has a narrow range of flammability; that is, in concentrations in air below about 5 percent and above about 15 percent by volume, natural gas will not burn. The high ignition temperature and limited flammability range make accidental ignition or combustion of natural gas unlikely.

8.4.2 Warranty

There are many dedicated natural gas vehicles available. These, are provided with standard manufacturers' warranties. In the case of aftermarket conversions, third party warranties are also available to cover gas related components. As an example, the Cummins warranty for both ISC (Diesel) and C8.3G+ (Natural Gas) engines is identical.

8.4.3 Functionality

The knock resistance of methane is high, which is advantageous for engine performance. The Research Octane Number of methane is about 120, enabling compression ratios of up to 13:1 to be achieved in some OEM engines. Though the maximum efficiency of a spark-ignition gas engine is estimated to be 10-15% lower than the efficiency of a diesel engine (Nyland and

Lawson, 2000), the data that were used in our analysis (based on engine dynamometer information) indicate that CNG is only 1.5% less efficient than low sulfur diesel.

CNG buses appear to display a large discrepancy between their theoretical or engine dynamometer performance, and their on-road performance. According to Bates et al. (2001) in the current French NGV programme, natural gas buses have 28% to 62% worse fuel consumption than diesel buses under real-life driving conditions.

Examination of the literature in relation to the use of CNG as a fuel for bus fleets (Watt, 2000; SRI International, 1996; Los Angeles County Metropolitan Transportation Authority, 1999) reveals that in general, CNG buses require greater maintenance. Stage Coach New Zealand reports that fires have been caused by backfiring problems as a result of faulty maintenance, including a failure to re-install flash arresters. Bell Street Buses in Melbourne report similar problems (Watt, 2000). The ANGVC believes that current generation technology, if properly fitted and maintained, should not give rise to incidents such as these.

The Los Angeles Country Metropolitan Transportation Authority (LACMTA) notes that due to chronic problems with the engine and fuel system components, CNG buses have had a significantly greater defect rate than diesel buses. A fleet of Orion V CNG buses in operation in New York with NYCMTA consistently had twice the road failures as the same model of diesel bus. Engine and fuel system road calls for the CNG buses were also twice as high as the road calls for diesel engines and fuel systems. In Los Angeles, the LACMTA vehicles' engines and fuel system road calls for a fleet of older diesel buses only accounted for 34 percent of total road calls.

Until there is large-scale experience with CNG bus maintenance, reliability problems and the likelihood of faulty maintenance as a result of unfamiliarity with the equipment will be greater with CNG buses than with diesel buses. Fleet operators in Australia often report that a change in maintenance procedures results in improved reliability. Due to the small size and varying ages of CNG fleets in Australia, it is difficult to make an accurate statistical evaluation of vehicle reliability. Adjustments to maintenance procedures and adjustments to driving style may both result in improved reliability.

The performance of CNG engine and fuel system components are expected to improve as the technology matures. The performance of natural gas engine and fuel system components have improved considerably in recent years and are expected to improve further as the technology matures. In the past this has been hampered by low demand for natural gas engines but increasing demand for low emissions engines is likely to accelerate technology improvements and reduce price differentials between natural gas and diesel engines.

8.4.4 *Operating range*

We have noted that a typical range for a CNG truck is 560 km, which can be increased to over 640 km by increasing the number of cylinders on board the vehicle or by increasing the CNG pressure within the tanks at the time of fill. In the case of dual-fuel operations, diesel capacity may also allow for additional range.

CNG buses are heavier than the corresponding diesel vehicle as a result of the weight of the tanks. The Sydney Bus fleet menu on the web provides technical details on the Sydney Bus fleet at http://www.sydneybuses.nsw.gov.au/sb.fleet.html. According to the information provided there, a Scania L113CRB CNG bus has an unladen weight of 11,240 kg and can carry 72 passengers. The equivalent Scania L113CRL diesel bus has an unladen weight of 11,040 kg and can carry 69 passengers. These Scania CNG buses have a range of 250 km. The newer Sydney Bus CNG buses are Mercedes Benz 0405H buses with a range of 400 km. Developments in cylinder technology in recent years have increased the capacity for on board storage. Older model

Scania buses, for example, carried around 530 kg of cylinders (excluding mounting hardware) to deliver a driving range of only 250 km, whereas current Scanias can deliver over 450 kms with only 550 kg of cylinders (including mounting hardware) on board.

8.4.5 Re-fuelling

Sydney Buses describe their refuelling system as follows:

The refuelling station at State Transit Authority's Kingsgrove bus depot has two 500 m³/hr compressors, eight 250 bar storage cylinders and the associated dispensers and reclaiming units. Each depot will have three compressors operating at a rate of 3000 cubic metres per hour at 34 MPa. The storage cascade has a total capacity of 3500 cubic metres. The buses will be able to be filled from empty to 20 MPa in three and a half minutes, with up to 40 buses being filled within two hours. The process is automatic with connection and disconnection of the coupling the only manual requirement.

Currently there are limited public CNG refuelling facilities (total 13) but over 30 public sites are expected to be operational by the end of 2002. In addition the demand for depot-based sites is increasing and it is expect that a similar number of additional depot based stations will be developed over this time. NGVs can also be fuelled from a small dispenser directly connected to a home or business natural gas line. This is commonly known as a Vehicle Refuelling Appliance (VRA). A small electrically driven compressor operates the dispenser.

8.4.6 Availability

Natural gas is abundant in Australia thus, in principle, there are no problems with fuel availability. In practice, natural gas is vulnerable to disruption in the gas supply. This was most evident with the Longford incident in 1998 when gas supplies to Melbourne, and much of the rest of Victoria were halted following the disaster at the Longford plant. New pipelines are under construction to ensure alternate gas supply routes to Sydney and Melbourne.

8.5 CNG conversions

The majority of CNG vehicles in Australia were sourced as new vehicles. However, there has been growing interest in the conversion of conventionally fuelled vehicles to CNG through aftermarket conversions.

The emissions performance of converted Australian CNG vehicles is unclear due to a lack of comprehensive industry-wide data. The only results available were from one system that was used in a small number of vehicles. That system is currently being upgraded and is no longer sold in the previous configuration. Some tailpipe emissions from the previous configuration were much higher than those for OEM vehicles. It is possible that the difference in emission levels between converted vehicles and OEMs may decrease as the heavy-duty vehicles conversion industry becomes more firmly established.

8.6 Health Issues

NGVs have the potential to effect a significant reduction in local air pollutants such as CO, NMHCs, SO_x, particles, smoke and odour. The effects of traces of formaldehyde in NGV exhausts (though less than from alcohol fuels) have yet to be determined.

8.5.1 *Production and transport*

Particulate Matter

The LCA estimate for CNG urban precombustion (truck) PM10 emissions of 3 to 4 mg/km is substantially less than the LSD estimate of 43 mg/km.

Air Toxics

The LCA estimate for CNG urban precombustion (truck) NMHC emissions of 0.007 to 0.011 g/km is substantially less than the LSD estimate of 0.292 g/km.

The public health effects of air toxics will be mainly associated with combustion emissions in large urban centres. An accompanying disk to this report provides details of air toxic emissions from upstream activities.

8.5.2 Use

Anyon (1998) points out that LPG, like CNG, has much lower emissions than diesel, and LPG has low particle levels, which make it an attractive fuel for urban buses and delivery vehicles. However, as diesel particle emissions reduce to Euro4 levels this advantage may be lost.

Exhaust emissions of methane, which is a greenhouse gas, are relatively high.

Particulate Matter

Research consistently shows that CNG (and gaseous fuels in general) with its simple chemistry and very low sulfur content, emit extremely low levels of particles. (Anyon, 1998)

Emissions of particulate matter are almost eliminated with natural gas use as shown in the earlier results tables. The IANGV (1990) noted that the NGV engine's lubricating oil appeared to be the source of remaining particle emissions.

The LCA estimate for CNG combustion (truck) PM10 emissions of 7.2 mg/km is substantially less than the LSD estimate of 380 mg/km.

Air Toxics

CNG produces much lower emissions of the main air toxics such as benzene, 1,3 butadiene, formaldehyde and acetaldehyde, compared with diesel (Anyon, 1998)

CNG contains no benzene, so refuelling and running losses of this toxic would be zero. (USEPA, 1993)

The LCA estimate for CNG combustion (truck) NMHC emissions of 0.212 g/km is less than the LSD estimate of 0.900 g/km.

Summary

CNG upstream emissions of both particles and air toxics are substantially less than LSD. CNG tailpipe emissions of particles are substantially less than LSD. CNG tailpipe emission of benzene, 1,3 butadiene, formaldehyde and acetaldehyde are less than LSD.

No comparative emissions data for CNG and LSD has been identified for:

- polycyclic aromatic hydrocarbons (PAH);
- toluene; and
- xylene

8.7 OHS Issues

Australian long distance high pressure (up to 15 MPa) transmission pipelines are relatively modern (the oldest dates back to 1969) and built to high standards. They are well maintained and accidental leaks are a rarity. Refuelling CNG is considered to be the 'least-safe' moment of its use. CNG is much lighter than air and thus it is safer than spilled diesel. In the case of CNG leak, because of the gaseous nature of the fuel, the gas will issue as a very high velocity jet into surroundings aiding greatly in the rapid dispersion of the fuel.

The OHS issues in the lifecycle of CNG are covered by a range of State and Commonwealth occupational health and safety provisions. While there will be different OHS issues involved in the production process associated with CNG compared with LSD, no OHS issues unique to the production and distribution of CNG have been identified.

8.8 Vapour Pressure Issues

Most gas losses from the distribution systems are by way of leakage from the low pressure network (7 kPa). This includes both the reticulation network and appliances operated by end users. Losses from the distribution network are difficult to estimate as they may occur both upstream and downstream from the meters. It is estimated that emissions from the distribution network, called unaccounted gas, i.e. the difference between the gas issued by the utilities and the gas sold to customers are as high as 7.5% (NGGIC, 1996).

Since the use of CNG as a fuel requires a closed delivery system, evaporative emissions from a dedicated CNG vehicle are assumed to be zero. (USEPA, 1993). Different views are held on evaporative emissions. One is that CNG vehicles do not have any, due to their sealed pressurised fuel system. BTCE (1994), on the other hand, refers to 'frequent leaks' as a technical problem to be solved for NGVs.

8.9 Environmental Impact and Benefit

Noise levels from natural gas buses are less than those of diesel buses. Kadayifci and Bryett (1997) measured a decrease of 2 to 5 dBA during drive-by tests, and 2 to 3 dBA during stationary noise tests. Tests in France on identical diesel and CNG buses found up to 8 dBA reductions in noise outside the bus. Passengers experienced about 4dBA less noise (MVV InnoTec GmbH, 2000).

The operational experience is salutary. Perception problems about poor driveability of CNG buses were put to rest with comparison trials with diesel buses. The conclusion was that lack of noise from the CNG buses gave the drivers the impression of a lack of acceleration (Watt, 2000:p.66)

NGVs have the potential to effect a significant reduction in local air pollutants such as CO, NMHCs, SO_x, particles, smoke and odour. The situation with regard to NO_x is less clear cut, and the effects of traces of formaldehyde in NGV exhausts (though less than from alcohol fuels) have yet to be determined.

The potential for water and soil pollution is effectively eliminated by the use of natural gas.

With respect to sustainability, known world reserves of natural gas now constitute over 95% of equivalent oil reserves. In Australia this ratio is more than three times the oil reserve. Proven Australian resources of natural gas currently stand at 109,051 PJ, at existing production levels, this will last 91 years compared to domestic oil reserves which are estimated to last 39 years. Natural Gas is an indigenous fuel that, if broadly adopted by the transport industry, could result in the order of an additional 100PJ per annum of gas being consumed rather than imported and more expensive crude oil (ANGVC 2001).

CNG can also be a renewable fuel for vehicles because it can be purified from the biogas extracted from waste treatment facilities.

8.10 Expected Future Emissions

Arcoumanis (2000) developed a model that examines a given alternative fuel relative to the reference diesel engine (Euro2) in terms of a specific regulated pollutant. A value of 1 implies identical performance to the low sulfur diesel/Euro2 combination. A value greater than 1 implies inferior performance, whereas a value less than 1 indicates superior performance.

Table 8.22 lists the estimated emissions factors for CNG. The columns in bold represent the standards relative to the Euro2 standard. The adjacent column gives the expected performance of CNG. The estimates of Arcoumanis (2000) indicate that CNG can be expected to meet all future Australian Design Rules for all pollutants.

Estimated emission factors for CNG under future technologies (PM is unregulated)										
Technology	СО	СО	тнс	ТНС	NOx	NOx	PM	РМ	CO ₂	LCA CO ₂
Euro2	1.0	0.3	1.0	0.9	1.0	0.2	1.0	0.1	1.0	0.9
Euro3	0.53	0.2	0.6	0.6	0.71	0.1	0.67	0.1	1.0	0.9
Euro4	0.38	0.1	0.42	0.4	0.5	0.1	0.2	0.05	1.0	0.8

Table 8.22

8.11 Summary

8.11.1 Advantages

- CNG has very low particle emissions because of its low carbon to hydrogen ratio.
- There are negligible evaporative emissions, requiring no relevant control. •
- Due to its low carbon-to-hydrogen ratio, it produces less carbon dioxide per GJ of fuel than diesel.
- It has low cold-start emissions due to its gaseous state. •
- It has extended flammability limits, allowing stable combustion at leaner mixtures. •
- It has a lower adiabatic flame temperature than diesel, leading to lower NOx emissions.
- It has a much higher ignition temperature than diesel, making it more difficult to auto-ignite, • thus safer.
- It contains non-toxic components. •
- It is much lighter than air and thus it is safer than spilled diesel. •
- Methane is not a volatile organic compound (VOC). •
- Engines fuelled with natural gas in heavy-duty vehicles offer more quiet operation than • equivalent diesel engines, making them more attractive for use in urban areas.
- It has nearly zero sulfur levels and, thus, negligible sulfate emissions. •
- Natural gas is distributed via underground pipe networks, removing the need for hazardous • transportation and transfer processes.
- Because of the pipeline delivery, retailers or fleet operators are not required to store large • quantities of fuel, usually prepaid, on site.
- Natural gas use does not give rise to issues with groundwater contamination such as those experienced through diesel/petrol spillage or leakage from underwater storage.

• Natural gas pricing is stable and predictable, removing uncertainty to business caused by fuel price fluctuations.

8.11.2 Disadvantages

- CNG on board a vehicle takes 3 to 4.5 times more volume for storage than diesel, thus storage needs may be reduced.
- It requires dedicated catalysts with high loading of active catalytic components to maximise methane oxidation.
- The composition may vary depending on the CNG source, which affects stoichiometric air/fuel ratios. This has not been a problem in Australia to date.
- It requires special refuelling stations that necessitate new infrastructure.
- The energy required to compress natural gas leads to increased greenhouse gas emissions.
- The extra weight of the fuel tank leads to higher fuel consumption or loss of payload.
- Exhaust emissions of methane, which is a greenhouse gas, are relatively high compared with low sulfur diesel.
- It can give rise to backfire in the inlet manifold if the ignition system is faulty or fails in use.
- Relatively small fugitive emissions of methane can have a significant effect on the exbodied greenhouse gas emissions.